

1 Landscape-level DNA metabarcoding study in the Pannonian forests reveals differential effects of
2 slope aspect on taxonomic and functional groups of fungi

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9

10 **Abstract**

11 In temperate regions, slope aspect is one of the most influential drivers of environmental conditions
12 at landscape level. The effect of aspect on vegetation has been well studied, but virtually nothing is
13 known about how fungal communities are shaped by aspect-driven environmental conditions. I
14 carried out DNA metabarcoding of fungi from soil samples taken in a selected study area of
15 Pannonian forests to compare richness and community composition of taxonomic and functional
16 groups of fungi between slopes of predominantly southerly vs. northerly aspect and to assess the
17 influence of selected environmental variables on fungal community composition. The deep sequence
18 data presented here (i.e. 980 766 quality-filtered sequences) indicate that both niche (environmental
19 filtering) and neutral (stochastic) processes shape fungal community composition at landscape level.
20 Fungal community composition correlated strongly with aspect, with many fungi showing preference
21 for either south-facing or north-facing slopes. Several taxonomic and functional groups showed
22 significant differences in richness between north- and south-facing slopes and strong compositional
23 differences were observed in all functional groups. The effect of aspect on fungal communities likely
24 is mediated through contrasting mesoclimatic conditions, that in turn influence edaphic processes as
25 well as vegetation. Finally, the data presented here provide an unprecedented insight into the
26 diversity and landscape-level community dynamics of fungi in the Pannonian forests.

27

28 **Key words:** beta diversity; fungal community ecology; ITS2 rDNA; topography

29

30 INTRODUCTION

31 In temperate regions, slope aspect is among the most influential factors that drive the
32 physical environment at the landscape level. The aspect effect is driven primarily by the variation in
33 the amount of solar radiation per surface area, which is a function of the angle of incidence of solar
34 radiation. As a result, slopes oriented toward the Equator receive higher intensity and greater
35 duration of solar radiation, which can be 50% higher on south-facing than on north-facing slopes in
36 the northern hemisphere (Geiger 1965; Rosenberg *et al.* 1983; Gilliam *et al.* 2014). These contrasting
37 energy inputs can profoundly alter mesoclimatic conditions, particularly air and upper soil
38 temperature, which, in turn, affect relative humidity, evapotranspiration, soil moisture and edaphic
39 processes (Fekedulegn *et al.* 2003). As a result, slopes with poleward (in this case, northerly) aspect
40 generally are cool and more humid, while slopes with southerly aspect tend to be markedly warmer
41 and drier, particularly in mid- and high latitudes (Holland & Steyn 1975; Méndez-Toribio *et al.*
42 2016).

43 Such aspect-related contrasts in the abiotic environment are intuitively expected to have
44 influence on the composition of biotic communities. Indeed, the effects of aspect on vegetation are
45 well known and clear compositional differences have been found between north- and south-facing
46 slopes in various ecosystems, e.g., in temperate and Mediterranean forests (Whittaker 1956;
47 Sternberg & Shoshany 2001; Gilliam *et al.* 2014), temperate grasslands (Schmidt 2013), boreal
48 forests (Hollingsworth *et al.* 2006), and arctic tundra (Walker *et al.* 1994). On the other hand,
49 virtually nothing is known about the influence of aspect on richness and community composition of
50 fungi at landscape level, except two recent studies focusing on arbuscular mycorrhizal fungal
51 richness in boreal forests and arid steppes in China (Chu *et al.* 2016; Liu *et al.* 2017). Therefore, how
52 slope aspect affects richness and community composition of various taxonomic and functional
53 groups of fungi remains unknown.

54 According to the macroecological study of Tedersoo *et al.* (2014), global-scale fungal
55 diversity and distribution patterns are primarily influenced by climatic factors, mainly mean annual
56 temperature and precipitation, followed by edaphic factors, particularly pH, and spatial patterns due
57 to dispersal limitation. On the other hand, the coupling between vegetation and soil fungal
58 community composition and richness appears to be much weaker, with the exception of
59 ectomycorrhizal (ECM) fungi (Tedersoo *et al.* 2014; Peay *et al.* 2016). Therefore, it is reasonable to
60 hypothesize that the above-mentioned aspect-driven environmental differences are expected to
61 influence the diversity and distribution of fungal communities at landscape level as well.

62 In this study, I tested the effects of aspect on the richness and composition of taxonomic and
63 functional groups of fungi at a selected study site in the Pannonian forests, where regional climatic

64 conditions and geological parent material were uniform in order to minimize confounding factors.
65 The Pannonian biogeographical region represents unique ecosystems in Europe (Sundseth 2009). It
66 covers the inner parts of the Carpathian basin and is characterized by a great wealth of floristic and
67 faunistic elements from different parts of Eurasia. More specifically, in addition to the broadly
68 distributed Eurasian species, there are numerous sub-Mediterranean, continental, Pontic and
69 Balkanian species. The distinctiveness of the Pannonian biogeographic region comes from the
70 combination of these elements, as well as the occurrence of Pannonian endemics (Fekete *et al.* 2016).
71 The geologically diverse mountain range of the Északi-középhegység (North Hungarian Mountains)
72 represents the northern part of the Pannonian biogeographic region. The region is characterized by
73 high habitat diversity, partly due to the complex geology which features a wide variety of calcareous,
74 volcanic and igneous rocks (Pelikán 2010), and partly due to the diverse topography that creates a
75 broad spectrum of mesoclimatic conditions. This diversity of habitats allows for the coexistence of
76 sub-Mediterranean, continental, Atlantic, and Carpathian floristic and faunistic elements often in
77 close proximity, depending on slope, aspect, elevation, and geological parent material (Suba 1983;
78 Vojtkó 2002; Vojtkó *et al.* 2010). With respect to macrofungi, there is a long history of sporocarp-
79 based studies in various areas of the Északi-középhegység (Bohus & Babos 1960; Takács & Siller
80 1980; Rimóczi 1992, 1994; Tóth 1999; Siller *et al.* 2002; Albert & Dima 2005; Egri 2007; Pál-Fám
81 *et al.* 2007; Rudolf *et al.* 2008; Siller 2010; Siller & Dima 2014). In addition, morphological and
82 molecular analyses of roots colonized by ECM genera *Humaria*, *Gevea*, *Tomentella*, and *Tuber* have
83 been carried out in a well-preserved montane beech forest reserve (Kovács & Jakucs 2006; Erős-
84 Honti *et al.* 2008; Jakucs *et al.* 2015). However, the diversity and distribution of fungi, particularly
85 microscopic fungi, in various habitats types are still unexplored and I am not aware of any molecular
86 study assessing the local (alpha) diversity and/or community turnover (beta diversity) of fungal
87 communities in the Pannonian forests.

88 The objectives of this DNA metabarcoding study were 1) to provide the first insights into the
89 kingdom-wide taxonomic and functional diversity of fungi in different Pannonian forest types on a
90 landscape scale; 2) to compare richness and community composition of fungi among sampling sites
91 among predominantly south- and north-facing slopes; and 3) to assess the influence of selected
92 environmental variables on fungal community composition.

93

94 MATERIALS AND METHODS

95 *Study area*

96 Within the broader region of the Északi-középhegység characterized above, the focal area of
97 this study is located in the valley of the Tarna creek situated between the Mátra and the Bükk, the

98 two highest mountains in Hungary. The study area is characterized by rolling hills of low elevations
99 (200-500 m a.s.l.) that are mostly covered with managed secondary forests. The regional climate is
100 subcontinental, with warm summers (mean July temperature ca. 20°C), cold winters (mean January
101 temperature ca. -3°C), and mean annual temperature and precipitation of ca. 9.5°C and ca. 580 mm,
102 respectively (Tóth 1983; Horváth & Gaálová 2007).

103 Pannonian-Balkan turkey oak-sessile oak forests (*Quercetum petraeae-cerris*, Natura 2000
104 code 91M0) form the zonal vegetation type on gentle slopes and in placor settings in the study area
105 as well as in other colline (200-450 m a.s.l.) areas of the Északi-középhegység. They are dominated
106 by oaks, such as *Quercus dalechampii*, *Q. petraea*, and *Q. cerris*, and can attain 15-25 m canopy
107 height. Within the turkey oak-sessile oak zone, Pannonian thermophilous downy oak forests (*Corno-*
108 *Quercetum pubescentis*, 91H0) generally develop on shallow and rocky soil on steep (> 20°), south-
109 facing slopes. These xerothermic forests are also dominated by the same oak species as the zonal
110 forest type, but the canopy remains low or medium tall (8-10 m), with the characteristic appearance
111 of *Q. pubescens*, particularly in the western half of Hungary, where this vegetation type is more
112 widespread. In the Északi-középhegység, thermophilous oak forests tend to occur in isolated patches
113 with particularly warm and dry mesoclimate (Vojtkó 2002; Borhidi 2003; Vojtkó *et al.* 2010; Bölöni
114 *et al.* 2011).

115 North-facing, gentle slopes are mostly covered by Pannonian sessile oak-hornbeam forests
116 (*Carici pilosae-Carpinetum*, 91G0), that are predominantly mesophilous, submontane (400-600 m
117 a.s.l.) forests, characteristic of cool, relatively humid habitats on deep soils, with a canopy height of
118 25-30 m. These mixed forests are dominated by *Carpinus betulus* and *Quercus petraea*, with *Acer*
119 *campestre*, *A. platanoides*, *Cerasus avium*, and *Tilia cordata* occurring sporadically. In the low-
120 elevation landscape of the study area, extrazonal submontane beech forests (*Melittio-Fagetum*) occur
121 very sporadically and are restricted to steep, north-facing slopes with the coldest mesoclimatic
122 conditions. These are tall (25-35 m), mesic forests, where *Fagus sylvatica* generally is
123 monodominant. In zonal settings in the Északi-középhegység, beech forests (9130) are represented
124 by submontane beech forest (*Melittio-Fagetum*) between 400 and 750 m a.s.l. and montane beech
125 forest (*Aconito-Fagetum*) above 750 m a.s.l., the latter featuring many species with Carpathian
126 distributions (Borhidi 2003; Vojtkó *et al.* 2010; Bölöni *et al.* 2011).

127 The fourteen sampling sites were evenly distributed in two hills (Kis-várhegy and Nagy-
128 várhegy) just south of the village of Sirok, Heves county, Hungary (Fig. 1, Table 1). The location
129 was chosen because it features, in an unusually small area (< 2 km²), the above-mentioned major
130 Pannonian forest types of northern Hungary, which are distributed according to aspect-based

131 differences in mesoclimatic conditions in a not spatially clustered manner. Furthermore, potentially
132 confounding factors that often limit the interpretation of ecological studies in natural habitats are
133 minimal: 1) the Kis-várhegy and Nagy-várhegy are uniformly composed of Jurassic calcareous
134 limestone (Pelikán 2010); 2) the regional climate is uniform throughout the small study area; 3) all
135 forest types are represented by old-growth (> 80y) secondary forests; 4) the elevation differences
136 among the sampling sites are minor (< 60 m) and are not related to aspect or to any other
137 environmental conditions; and 5) the spatial proximity and the continuous natural or semi-natural
138 landscape are expected to facilitate unrestricted dispersal of fungal propagules among the sites by
139 wind, water or by wildlife. Therefore, I hypothesized that any observed compositional differences
140 among the samples would be the result of niche-based and stochastic processes, with the former
141 primarily referring to environmental filtering according to differences in abiotic and biotic factors
142 driven by topography: predominantly northerly vs. southerly aspect and to a lesser extent by slope
143 angle.

144

145 *Sampling and molecular work*

146 For this study, fourteen sites were chosen that represented forests of northerly and southerly
147 aspects on both the Kis- and Nagy-várhegy. At each site (ca. 10 × 25 m), 20 samples (5 cm³ each) of
148 top soil from underneath the litter layer were taken at each sampling site in such a way that samples
149 were at least 2 m from each other to minimize the probability of sampling the same genet repeatedly.
150 Soil samples collected at a given site were pooled, resulting in a composite soil sample for each site.
151 Ca. 20 g of each composite sample was kept frozen until DNA extraction, while the rest was used for
152 soil chemical analyses to measure pH (water-based), and total carbon (C) and nitrogen (N) contents
153 following Sparks *et al.* (1996).

154 Genomic DNA was extracted from 0.5 ml of soil from each composite sample using
155 NucleoSpin[®] soil kit (Macherey-Nagel GmbH & Co., Düren, Germany), according to manufacturer's
156 protocol. The ITS2 region (ca. 250 bp) of the nuclear ribosomal DNA repeat was PCR amplified
157 using primers fITS7 (Ihrmark *et al.*, 2012) and ITS4 (White *et al.*, 1990). The ITS4 primer was
158 labelled with sample-specific Multiplex Identification DNA-tags (MIDs). The amplicon library was
159 sequenced at Naturalis Biodiversity Center (Naturalis) using an Ion 318TM Chip and an Ion Torrent
160 Personal Genome Machine (Life Technologies, Guilford, CT, U.S.A.). Detailed protocols of the
161 molecular work are described in Geml *et al.* (2014b).

162

163 *Bioinformatic work*

164 The initial clean-up of the raw data was carried out using Galaxy
165 (<https://main.g2.bx.psu.edu/root>), in which the sequences were sorted according to samples and
166 adapters (identification tags) were removed. The primers were removed and poor-quality ends were
167 trimmed off based on 0.02 error probability limit in Geneious Pro 8 (BioMatters, New Zealand).
168 Subsequently, sequences were filtered using USEARCH v.8.0 (Edgar, 2010) based on the following
169 settings: all sequences were truncated to 200 bp and sequences with expected error > 1 were
170 discarded. For each sample, sequences were collapsed into unique sequence types, while preserving
171 their counts. The quality-filtered sequences from all samples were grouped into operational
172 taxonomic units (OTUs) at 97% sequence similarity and putative chimeric sequences were removed
173 using USEARCH. I assigned sequences to taxonomic groups based on pairwise similarity searches
174 against the curated UNITE+INSD fungal ITS sequence database (version released on October 10,
175 2017), containing identified fungal sequences with assignments to Species Hypothesis (SH) groups
176 delimited based on dynamic sequence similarity thresholds (Kõljalg *et al.* 2013). After excluding
177 OTUs with < 80% similarity or < 150 bp pairwise alignment length to a fungal sequence, 6216
178 fungal OTUs were retained, representing a total of 980 766 quality-filtered sequences, including
179 1930 global singletons. Global singletons are routinely excluded from metabarcoding studies,
180 because the vast majority of them represent erroneous sequences (Lindahl *et al.* 2013; Geml *et al.*
181 2016, 2017). However, this practice likely results in the exclusion of some real high-quality data on
182 locally rare species. Because, to my best knowledge, this is the first metabarcoding study in the
183 Pannonian forests, I intended to keep all high-quality data, including 343 singletons with > 98%
184 sequence similarity to a reference fungal SH, a conservative threshold for conspecificity in most
185 fungal groups (Kõljalg *et al.* 2013), and excluded the remaining 1587 singletons from further
186 analyses. DNA sequences have been deposited in NCBI (accession numbers provided upon
187 manuscript acceptance).

188 Only OTUs with > 90% similarity to a fungal SH with known ecological function were
189 assigned to one of the following functional groups: animal pathogens, ECM fungi, lichens, litter
190 decomposers, mycoparasites, plant pathogens, root-associated fungi (non-ECM orchid and ericoid
191 mycorrhizal fungi and root endophytes), saprotrophs (generalists), and wood decomposers.
192 Arbuscular mycorrhizal fungi were not included in the analyses, because only one representative
193 OTU was found in the quality-filtered dataset. The initial functional assignments were made by
194 FunGuild (Nguyen *et al.* 2015) and were manually checked afterwards. For genera that are known to
195 comprise species from multiple functional guilds (e.g., *Amanita*, *Entoloma*, *Ramaria*, and many SH
196 groups in the Sebaciniales), I assigned ecological function for each OTU individually, based on
197 available ecological information for the matching SH in the UNITE database.

198

199 *Statistical analyses*

200 I normalized the OTU table for subsequent statistical analyses by rarefying the number of
201 high-quality fungal sequences to the smallest library size (36 946 reads). The resulting matrix
202 contained 4312 fungal OTUs. Linear regression analyses in R (R Development Core Team 2013)
203 were used to examine relationships between aspect and richness of taxonomic and functional groups
204 of fungi as well as between aspect and environmental variables, i.e. edaphic factors (pH, C, N, and
205 C/N) and the relative abundance values of the two dominant tree genera (*Carpinus* and *Quercus*) that
206 were distributed throughout the majority of sites. When aspect is treated as a continuous variable
207 from 0° to 360°, the two extreme values of this interval refer to the same aspect in the landscape
208 (north). Therefore, aspect was expressed as northerly aspect in degrees (south: 0°, north: 180°)
209 following Calef *et al.* (2005), which better reflects the well-known ecological differences between
210 north- and south-facing slopes.

211 I used the *vegan* R package (Oksanen *et al.* 2015) to run non-metric multidimensional scaling
212 (NMDS) on the Hellinger-transformed OTU table and a secondary matrix containing environmental
213 variables, which were standardized using the *scale* function in R. Ordinations were run separately for
214 functional groups as well as for all fungi with the following specifications: distance measure = Bray-
215 Curtis, dimensions = 2, initial configurations = 100, model = global, maximum number of iterations
216 = 200, convergence ratio for stress = 0.999999. I used the *envfit* R function to fit the above-
217 mentioned environmental variables and richness of various taxonomic and functional groups onto the
218 NMDS ordinations and plotted isolines of northerly aspect on the NMDS ordinations using the
219 *ordisurf* function. In addition, I tested whether fungal communities were statistically different among
220 forest types using the multiresponse permutation procedure (MRPP) and determined any preferences
221 of individual OTUs for forest types and for north- or south-facing slopes using indicator species
222 analyses (Duf rene & Legendre 1997) in PC-ORD v. 6.0 (McCune & Grace 2002).

223 A series of mantel tests were carried out in PC-ORD to reveal any spatial autocorrelation in
224 environmental variables as well as fungal community composition among the sampling sites and to
225 measure correlation between fungal community composition and environmental variables.
226 Furthermore, a series of partial mantel tests were applied to differentiate the effects of northerly
227 aspect, tree genera, and edaphic factors on fungal community structure. Finally, I estimated what
228 proportions of the total variation in fungal community composition were explained by topography,
229 tree genera, and edaphic factors using variation partitioning (Borcard *et al.* 1992; Legendre *et al.*
230 2005) with a series of redundancy analyses (RDAs) in Canoco 5 (Microcomputer Power, Ithaca, NY,

231 USA). The expectation in purely neutral (stochastically assembled) communities is no correlation
232 between the environmental distance and the community dissimilarity of two sites (Smith &
233 Lundholm 2010).

234

235 RESULTS

236 *Correlation of elevation with fungal richness and environmental variables*

237 In total, the quality-filtered and rarefied dataset contained 4312 fungal OTUs in the 14 soil
238 samples taken from the representative types of Pannonian forests. Of these, 2196 OTUs had > 90%
239 similarity to a fungal SH with known ecological function were assigned to functional groups.
240 Richness values of most functional groups showed significant ($p < 0.05$) or marginally ($p < 0.1$)
241 significant correlation with northerly aspect. For example, I found positive correlation in root-
242 associated fungi ($r^2 = 0.231$; $p = 0.047$), mycoparasites ($r^2 = 0.148$; $p = 0.096$), and wood
243 decomposers ($r^2 = 0.204$; $p = 0.059$) with northerly aspect, while negative correlation was observed
244 in lichens ($r^2 = 0.154$; $p = 0.091$), plant pathogens ($r^2 = 0.231$; $p = 0.047$), and generalist saprotrophs
245 ($r^2 = 0.615$; $p < 0.001$) (Fig. 2). Richness of several taxonomic groups also correlated with northerly
246 aspect with varying significance, positively in Agaricomycetes ($r^2 = 0.261$; $p = 0.036$),
247 Tremellomycetes ($r^2 = 0.211$; $p = 0.056$), Leotiomycetes ($r^2 = 0.166$; $p = 0.083$), and
248 Mortierellomycota and Mucoromycota that had formerly been classified to Zygomycota ($r^2 = 0.465$;
249 $p = 0.004$), and negatively in Dothideomycetes ($r^2 = 0.605$; $p < 0.001$), Eurotiomycetes ($r^2 = 0.523$; p
250 $= 0.002$), and Sordariomycetes ($r^2 = 0.386$; $p = 0.011$) (Fig. 2). Among the measured edaphic factors,
251 only pH correlated significantly with northerly aspect, showing strong negative correlation ($r^2 =$
252 0.516 ; $p = 0.002$). With respect to vegetation, the relative abundance of *Quercus* correlated
253 negatively ($r^2 = 0.837$; $p < 0.001$), while that of *Carpinus* correlated positively ($r^2 = 0.269$; $p =$
254 0.033) with northerly aspect (Fig. 2).

255

256 *Comparing fungal community composition among the sampling sites*

257 All NMDS analyses resulted in 2-dimensional solutions with the following final stress values
258 for all fungi (0.0693), animal pathogens (0.1396), ECM fungi (0.1046), litter fungi (0.1575),
259 mycoparasites (0.0871), plant pathogens (0.0679), root-associated fungi (0.1342), saprotrophs
260 (0.0947), and wood decomposers (0.1471). In all cases, the NMDS ordinations revealed strong
261 structuring of fungal communities according to aspect. In addition, a weaker, but often clear
262 separation could be observed among the four forest types. Community composition of all fungi in the
263 sampling sites clearly structured according aspect which represented the first axis ($r = 0.999$) (Fig.
264 3). Soil pH ($r = -0.993$) correlated strongly with the first axis in that south-facing slopes had higher

265 pH than north-facing ones. With respect to trees, relative abundance of oaks ($r = -0.977$) correlated
266 negatively, while both hornbeam ($r = 0.587$) and beech ($r = 0.311$) correlated positively with
267 northerly aspect, although these latter two in a weaker manner. The second axis seemed to represent
268 the differences in C, N, and particularly in C/N ratio ($r = 0.998$) among the sites as well as the
269 varying dominance of hornbeam ($r = -0.810$) versus beech ($r = 0.951$) on the north-facing slopes.
270 Richness of only two functional groups showed significant correlation with the NMDS ordinations:
271 generalist saprotrophs correlating negatively with both the first ($r = -0.835$) and second axes ($r = -$
272 0.551), while lichens predominantly correlated with the second axis ($r = 0.934$) and to a lesser extent
273 with the first axis ($r = -0.356$). As for taxonomic groups with significant results, richness values of
274 Dothideomycetes ($r = -0.976$), Eurotiomycetes ($r = -0.848$), and Sordariomycetes ($r = -0.753$) were
275 negatively, while that of Mortierellomycota and Mucoromycota ($r = 0.921$) was positively correlated
276 with northerly aspect (Fig. 3).

277 NMDS plots of the datasets corresponding to the functional groups of fungi showed similar
278 correlations with aspect, forest types, and edaphic variables as detailed above. The NMDS
279 ordinations of four largest functional groups in terms of OTU richness, i.e. ECM fungi, plant
280 pathogens, generalist saprotrophs, and wood decomposers, are shown in Fig. 4, while those of the
281 three less species-rich functional groups, i.e. animal pathogens, litter decomposers, mycoparasites,
282 and root-associated fungi are shown in Fig. S1. Richness values of taxonomic families (or genera,
283 when a family was represented by a single genus) of ECM fungi that correlated with northerly aspect
284 were as follows: only ascomycete genera, i.e. *Elaphomyces*, *Helvella*, and *Tuber* showed negative
285 correlations with northerly aspect, while only basidiomycetes, i.e. *Amanita*, Atheliaceae, *Cortinarius*,
286 *Inocybe*, *Laccaria*, Russulaceae, and *Sebacina* correlated positively with northerly aspect. In
287 addition, *Cenococcum*, *Clavulina*, and Thelephoraceae showed positive correlation with axis 2.
288 Vectors representing the richness of taxonomic groups of plant pathogens seemed to be relatively
289 evenly distributed in the ordination plot. Of these, Bionectriaceae and Mycosphaerellaceae showed
290 preference for hornbeam forests, while *Leptodontitidum*, Taphrinaceae, and Venturiaceae seemed to
291 favor beech forests, both on north-facing slopes. Conversely, richness in Coniochaetaceae,
292 Didymellaceae, Massarinaceae, Phaeosphaeriaceae, *Pyrenochaeta*, and Spizellomycetaceae
293 correlated strongly with southerly aspect. With regard to generalist saprotrophs, the majority of taxa
294 showed clear preference for south-facing slopes, such as *Acremonium*, Aspergillaceae,
295 Chaetomiaceae, Didymosphaeriaceae, Gomphaceae, Helotiaceae, Lasiosphaeriaceae, Microascaceae,
296 Onygenaceae, Sporormiaceae, Stachybotryaceae, *Tetracladium*, Trichocomaceae,
297 Tricholomataceae, Trichomeriaceae, and Trichosporonaceae, while Mortierellaceae, Tremellaceae,
298 Umbellopsidaceae, and to some extent Cunninghamellaceae tended to have higher richness in sites of

299 northerly aspect. In wood decomposers, Corticiaceae, Hyaloscyphaceae, Lophiostomataceae, and
300 *Pluteus* correlated negatively, while Chaetosphaeriaceae, Crepidotaceae, Helotiaceae, *Trechispora*,
301 and Xylariaceae correlated positively with northerly aspect. MRPP confirmed the importance of
302 aspect in shaping fungal community composition (effect size $A = 0.084$, probability $p < 0.001$).
303 There were 137 and 174 significant ($p < 0.05$) indicator fungal OTUs characteristic of north- and
304 south-facing slopes, of which, 46 and 69 were assigned to functional groups, respectively (Table 2).

305

306 *Assessing the effects of environmental factors on fungal community composition*

307 Mantel tests showed that neither spatial proximity nor slope had any significant correlation
308 with fungal community composition or with aspect, relative abundance of tree genera, and edaphic
309 factors. On the other hand, aspect was strongly correlated with the relative abundance of tree genera
310 ($r = 0.526$; $p < 0.001$) and weakly with edaphic factors ($r = 0.151$; $p = 0.046$), while there was
311 moderate correlation between tree genera and edaphic factors ($r = 0.266$; $p < 0.035$) (Table 3).
312 Fungal community composition was strongly correlated with aspect ($r = 0.743$, $p < 0.001$), relative
313 abundance of tree genera ($r = 0.558$, $p < 0.001$), and edaphic factors ($r = 0.535$, $p < 0.001$).
314 Nonetheless, partial mantel tests indicated that aspect in itself had a strongly significant effect on
315 community structure ($r = 0.637$, $p < 0.001$) when tree genera was accounted for (control matrix),
316 while the effect of tree genera was substantially weaker, though still significant ($r = 0.295$, $p =$
317 0.018) when aspect was controlled. Conversely, the correlations of aspect ($r = 0.793$, $p < 0.001$) and
318 edaphic factors ($r = 0.639$, $p < 0.001$) with fungal community composition were both strong when
319 edaphic factors or aspect were controlled for, respectively (Table 3).

320 Variation partitioning analyses indicated that the tested environmental variables explained
321 24.3% of the total variation observed. Topography (aspect and slope) explained 15.1% of the
322 variation, but only 0.8% after accounting for the relative abundance of dominant tree genera and
323 edaphic factors. Tree genera accounted for 14.7% of the variation and 3.1% when the effect of other
324 variables was removed. Measured soil variables explained 17.9% of the variation, of which 5.9%
325 was attributed to their 'pure' effect. The shared variation explained by topography and tree genera
326 and by topography and edaphic factors was 2.5% and 2.9%, respectively, while the shared effects of
327 all three groups of variables accounted for 8.9% of the total variation (Fig. S2).

328

329 4. DISCUSSION

330 *Drivers of landscape-level compositional patterns of fungal communities*

331 The deep sequence data presented here clearly show that fungal community composition at
332 the selected Pannonian forest sites is strongly structured according to slope aspect. Although some

333 fungal species occurred in all samples, the majority of fungal OTUs preferred either south- or north-
334 facing slopes, as suggested by the observed community turnover as well as the numerous indicator
335 species. Similarly, even though total fungal richness was not statistically different between north-
336 and south-facing slopes, many functional and taxonomic groups of fungi were more diverse on
337 slopes of northerly or southerly aspect.

338 Naturally, the effect of slope aspect on fungal community composition and richness is
339 mediated through abiotic and biotic factors that are driven either directly or indirectly by the
340 differences in net solar radiation received in north- and south-facing slopes. For example, aspect
341 strongly influences local air and surface temperature, soil moisture, relative humidity, and soil
342 chemical processes (Gilliam *et al.* 2014). Consequently, the habitats found on north- and south-
343 facing slopes have distinct meso- and microclimatic as well as edaphic conditions. These differences
344 are particularly marked regarding available moisture, soil organic matter, and pH, as had been
345 observed in the study region previously (Dobos 2010), which, in turn, influence the composition of
346 biotic communities and their interactions. Although the relative abundance of the two dominant tree
347 genera (*Carpinus* and *Quercus*) was strongly influenced by aspect, they occurred in most of the sites.
348 Therefore, and because the vast majority of plant-associated fungi are not strictly specific to a tree
349 genus or family, I argue that to a great extent the observed fungal community patterns are caused by
350 a complex array of aspect-driven environmental variables and not by the type of vegetation alone.

351 The availability of soil moisture is one of the most important environmental conditions that
352 determines richness as well as community composition in fungi (Crowther *et al.* 2014; Tedersoo *et al.*
353 2014). Because at landscape scale, as in the study area, annual precipitation is considered uniform
354 across the sampling sites due to their close proximity, topography likely is the most important factor
355 that influences moisture availability through contrasting levels of evapotranspiration between north-
356 and south-facing slopes, as has been observed in several biomes (Fekedulegn *et al.* 2003; Méndez-
357 Toribio *et al.* 2016). Drought-tolerant fungi are generally regarded as generalists, instead of dry
358 specialists, with respect to their ability to grow along a broad moisture gradient (Lennon *et al.* 2012),
359 while drought-intolerant fungi are considered specialists to a narrow range of mesic conditions,
360 where they likely have competitive advantage over the generalists (Crowther *et al.* 2014). The
361 resulting competitive dynamics may force drought-tolerant fungi to the more xeric south-facing
362 slopes, while fungi with a stronger competitive ability under mesic conditions are expected to
363 dominate the mesophilous forests on the north-facing slopes in the study area.

364 Soil pH is also known to play an important role in shaping fungal communities (Porter *et al.*
365 1987; Coughlan *et al.* 2000; Lauber *et al.* 2008; Rousk *et al.* 2010; Geml *et al.* 2014a; Tedersoo *et al.*
366 2014; Glassman *et al.* 2017) and is often influenced by slope aspect (Gilliam *et al.* 2014; Chu *et al.*

367 2016). Because many fungal species have a relatively wide pH optimum (e.g. Wheeler *et al.* 1991;
368 Nevarez *et al.* 2009), it is likely that the observed correlation of pH with community composition is
369 mainly indirect, e.g., via altering nutrient availability and competitive interactions between soil fungi
370 and bacteria (Rousk *et al.* 2008), and other soil biota. In the study sites, there was a strong negative
371 correlation of soil pH with northerly aspect and the interaction of topography (mainly aspect) and
372 edaphic factors (primarily pH) provided a large fraction of the explained variation in fungal
373 community composition. Therefore, it is difficult to disentangle the ‘pure’ effect of pH from that of
374 aspect. Nevertheless, for some fungal groups that are known to be influenced by soil pH, the data
375 presented here are in agreement with previous results. For example, root endophytic fungi had been
376 shown to prefer low soil pH (Postma *et al.* 2006) and the strong preference of non-ECM root-
377 associated fungi for the northerly sites with lower pH in this study confirms the above trend.
378 Similarly, ECM fungi are generally considered acidophilus (Read 1991), and there was weak,
379 although non-significant, positive correlation between ECM fungal richness and northerly aspect (i.e.
380 sites with lower pH). The only other edaphic factors with strong correlation with fungal community
381 structure was C/N ratio, which was not related to aspect. Instead, changes in C/N ratio appeared to be
382 related to different forest types on the north-facing slopes, i.e. oak-hornbeam and submontane beech
383 forests. Because C/N is considered a direct measure of resource quality (Nielsen *et al.* 2010), it is
384 possible that the higher C/N values in the beech forests are driven by differences in litter quality
385 between beech and hornbeam and oak. Measurements from more beech and oak-hornbeam stands are
386 needed to test this hypothesis.

387

388 *The contrasting effects of aspect on functional groups of fungi*

389 The data clearly show a strong emerging pattern driven by slope aspect in all fungal groups.
390 Several functional groups showed strong differences in richness between north- and south-facing
391 slopes. Mycoparasites, non-ECM root-associated fungi, and wood decomposers had higher richness
392 in the sites with predominantly northerly aspect. The cooler microclimate in particular may be more
393 advantageous for root-associated fungi, because this group has been shown to be more diverse at
394 higher elevations in altitudinal gradient studies (Geml *et al.* 2014b). On the other hand, although
395 richness of wood decay fungi and mycoparasites have been shown to decline with decreasing
396 temperature in elevation gradients (Geml *et al.* 2014b), considering the given moisture limitation in
397 this study area, their higher richness in north-facing slopes may be more strongly driven by the
398 greater availability of moisture. Lichens, plant pathogens and generalist saprotrophs were clearly
399 more species rich in the south-facing slopes. With regard to lichens, this may be a consequence of
400 the more open canopy of the thermophilous oak forests that allow more light to penetrate and

401 apparently fosters the growth of the generally shade-intolerant and drought-tolerant lichens. On the
402 other hand, plant pathogens and generalist saprotrophs may benefit more from the higher
403 temperatures characteristic of the south-facing slopes, as richness values of these groups generally
404 correlate positively with temperature in altitudinal as well as in global studies (Geml *et al.* 2014b,
405 Tedersoo *et al.* 2014). Also, many generalist saprotrophs (e.g., in Eurotiomycetes) are known for
406 their drought-tolerance and, based on the above-mentioned theory on competitive dynamics, they
407 may be outcompeted in the mesic sites by mesophilic specialists, resulting in their higher diversity in
408 the warmer and drier south-facing slopes.

409 Even in functional groups with no significant differences in richness, I observed strong
410 compositional differences between slopes of northerly and southerly aspects. For example, in ECM
411 fungi, ascomycete and basidiomycete genera were clearly more diverse in the south- and north-
412 facing slopes, respectively. This is in agreement with other studies showing that ECM communities
413 often are dominated by ascomycetes in arid and semiarid environments, while basidiomycetes tend to
414 dominate more moist habitats (Cavender-Bares *et al.* 2009; Gehring *et al.* 2014). This may be linked
415 to their foraging and fruiting strategies and the related differences in C requirement from the host
416 trees. For example, trees in more xeric conditions are expected to preferentially associate with ECM
417 fungi that have contact and short-distance, as opposed to medium- and long-distance, extrametrical
418 mycelial exploration types and smaller fruiting bodies that are less costly in terms of C requirement.
419

420 *The unexplained component of community turnover*

421 An ever-present feature of fungal community studies is the large amount of unexplained
422 variation in richness and community composition even at small spatial scales (Peay *et al.* 2016). In
423 this study, the most significant compositional differences were observed between slopes of northerly
424 and southerly aspects, but there was also substantial variation among sites within sites on the same
425 slope. The environmental variables measured in this study explained about one quarter of the
426 variation in fungal community composition in all sites, which confirms the above-mentioned
427 substantial unexplained component of community assembly. Because many OTUs were rare, i.e.
428 found only in one or two sites, most such differences may be due to random processes in community
429 assembly as well as due to random sampling, as truly exhaustive soil sampling is practically
430 impossible to achieve in the field. The former is partly explained by the priority effect, i.e. within a
431 given species pool of a particular habitat, stochastic dispersal determines the order in which newly
432 available resources are colonized by different species, which, in turn, drives to a large extent the
433 composition of the community (Peay *et al.* 2016). In addition, other factors not examined here, such
434 as density-dependent processes (e.g., intra- and interspecific competitions and pathogen-host

435 interactions) as well potentially important environmental variables not yet measured, may also
436 contribute to the observed beta diversity. For example, I did not measure soil and air temperature,
437 relative humidity, and soil moisture at the sites, partly because it is already well-known that these
438 factors correlate strongly with aspect on mesoclimate scale, as mentioned above, and partly because
439 obtaining a realistic characterization of these variables would have required measurements taken
440 throughout the growing season at the sampling sites, which was beyond the scope and logistic
441 possibilities of this case study. However, by including site-specific microclimate data as well as a
442 more extensive list of edaphic variables, future studies may obtain more insights into the variation of
443 environmental factors at small spatial scales as well as their influence on fungal community
444 composition and turnover.

445

446 *Contribution to the knowledge on Pannonian forest fungi*

447 This study shows a remarkably high fungal diversity in a small area (< 2 km²) of secondary
448 forests in the eastern edge of the Mátra mountains. After the stringent quality filtering steps, the non-
449 rarefied and rarefied datasets contained 4695 and 4312 fungal OTUs, respectively. Based on the
450 rarefied data, well over 1000 OTUs occurred in any given sample (Fig. 2), each representing an area
451 of ca. 250 m². In total, I detected representatives of 707 fungal genera, of which 467 belonged to
452 Ascomycota, 225 to Basidiomycota, 6 to Chytridiomycota, and 7 represented early-diverging
453 lineages formerly classified in Zygomycota. Nonetheless, the true generic diversity likely is even
454 higher, because the vast majority of fungi in the sampled sites, as well as globally, are microfungi
455 and many of them could only be assigned to families, orders or classes due to the lack of sufficiently
456 identified reference data. Consequently, these represent species with unknown identity, several of
457 which may still be undescribed. With respect to the identified microfungi, the results presented here
458 may be the first data on their diversity and possible habitat preference in the Pannonian
459 biogeographic region, serving as potential reference data for future studies as well. By providing the
460 full list of taxa corresponding to the 2542 unique SHs that matched the OTUs in these samples with
461 high (> 95%) sequence similarity (Table S1), I intend to facilitate other mycological and fungal
462 ecological studies in the region.

463 I was able to assign a relatively high proportion (51%) of fungal OTUs to functional groups,
464 particularly in taxa with macroscopic fruiting bodies for which extensive reference data are available
465 from Europe, e.g., agarics, boletes, coral fungi, polypores, and several true and false truffles. For
466 example, the number of unique matching SHs in some of the most diverse basidiomycete genera
467 were 69 in *Inocybe*, 48 in *Russula*, 18 in *Lepiota*, while somewhat surprisingly *Cortinarius* and
468 *Lactarius* were only represented by 23 and 13 unique SHs, respectively. On the other hand, the data

469 also showed a high diversity ECM basidiomycetes with inconspicuous fruiting bodies, such *Sebacina*
470 (66 SHs) and *Tomentella* (83 SHs) that are notoriously underrepresented in sporocarp surveys
471 (Gardes & Bruns 1996; Kõljalg *et al.* 2000; Geml *et al.* 2012). Furthermore, I detected numerous
472 hypogeous ECM fungi, such as *Elaphomyces muricatus* and *E. papillatus*, *Genea verrucosa*,
473 *Hydnotrya* sp., *Hymenogaster australis*, *H. griseus*, *Hysterangium stoloniferum*, *Gautieria*
474 *graveolens*, *Melanogaster ambiguus*, *M. broomeanus*, and *M. variegatus*, *Wakefieldia macrospora*,
475 as well as several species of the charismatic *Tuber* genus (e.g., *T. aestivum*, *T. borchii*, *T. brumale*, *T.*
476 *maculatum*, *T. puberulum*, *T. rapaeodorum*, *T. rufum* and possibly *T. fulgens*). Most of these are
477 known to occur in Hungary, except for *Gautieria graveolens*, as this genus is only represented by an
478 unidentified species in the list of hypogeous fungi for the Carpathian-Pannonian region compiled by
479 Bratek *et al.* (2013). Representatives of these hypogeous fungi were found in all forest types,
480 indicating that they probably are relatively common in Pannonian forests.

481 In addition, I want to emphasize the suitability of DNA metabarcoding to complement
482 sporocarp-based assessments for biological monitoring and conservation of fungi. Specifically,
483 studies such as this can provide valuable knowledge not only on the total diversity of fungi at any
484 given site, but also on the distribution and habitat preferences of numerous species, including rare
485 and/or protected species. For example, I found at least one of the thirty-five species of fungi
486 protected by law in Hungary (Siller *et al.* 2005, 2006): *Strobilomyces strobilaceus*, which was
487 detected on north-facing slopes of both the Kis- and Nagy-Várhegy, which confirms the habitat
488 preference of this species for mesophilic beech and oak-hornbeam forests in Hungary (Siller *et al.*
489 2005). Furthermore, even though the retained global singletons that were > 98% similar to a
490 reference fungal SH had only minor contribution to richness values and had practically no influence
491 on the community structure patterns, 238 were identified to a genus or species. Of these, the
492 following taxa were only represented by singletons in the dataset: *Aleurodiscus aurantiacus*,
493 *Annulohyphoxylon multiforme*, *Aspergillus cibarius*, *Baeospora myosura*, *Barbatosphaeria dryina*,
494 *Bullera alba*, *Calonectria quinqueseptata*, *Caloplaca obscurella*, *Chaetosphaeria decastyla*,
495 *Chroogomphus mediterraneus*, *Clavariadelphus* sp., *Coprinopsis picacea*, *Coprinus comatus*,
496 *Cortinarius casimiri*, *Cortinarius cotoneus*, *Cortinarius infractus*, *Cortinarius pardinus*, *Crepidotus*
497 *mollis*, *Cryptodiscus pallidus*, *Cylindrocladiella elegans*, *Dactylellina ellipsospora*, *Darksidea*
498 *epsilon*, *Durella connivens*, *Elaphomyces papillatus*, *Elmerina caryae*, *Exidia truncata*, *Filobasidium*
499 *magnum*, *Geastrum fornicatum*, *Geopora* sp., *Gloeocystidiellum kenyense*, *Gromoniopsis idaeicola*,
500 *Haptocillium campanulatum*, *Hyalorbilia erythrostigma*, *Hypogymnia physodes*, *Jattaea tumidula*,
501 *Kretzschmaria deusta*, *Lactarius necator*, *Leiosepium tulasneanum*, *Leptosphaeria rubefaciens*,
502 *Leptospora rubella*, *Leucoagaricus barssii*, *Lophiotrema eburnoides*, *Melanelixia subaurifera*,

503 *Melanospora damnosa*, *Microstroma phylloplanum*, *Mycena renati*, *Mycosphaerella ulmi*,
504 *Neocatelunostroma abietis*, *Neocladophialophora quercina*, *Ossicaulis lignatilis*, *Otidea onotica*,
505 *Paxillus obscurisporus*, *Peltaster fruticola*, *Peniophorella praetermissa*, *Pestalotiposis chinensis*,
506 *Phlebia tremellosa*, *Pholiotina teneroides*, *Phyllozoma linderiae*, *Pilidium concavum*, *Pleurophoma*
507 *ossicola*, *Pluteus velutinus*, *Psilocybe inquilina*, *Resupinatus applicatus*, *Ramphoria pyriformis*,
508 *Rhytisma acerinum*, *Russula albonigra*, *Russula inamoena*, *Russula sororia*, *Sarcosphaera*
509 *coronaria*, *Scheffersomyces stipitis*, *Sistotrema brinkmannii*, *Slopeiomyces cylindrosporus*,
510 *Sphaerollopsis macroconidialis*, *Steccherinum ochraceum*, *Stylonectria norvegica*, *Thielavia*
511 *subthermophila*, *Tomentella badia*, *Trechispora laevis*, *Tremella yokohamensis*, *Tricholoma batschi*,
512 *Urocystis agropyri*, *Ustilago nunavutica*, *Valsivia insitiva*, *Vanrija humicola*, *Verrucaria dolosa*,
513 *Xenasmatella borealis*, *Xylodon rimossimus*, *Xylomelasma* sp., *Zalerion arboricola*, and *Zasmidium*
514 *dalbergiae*. Most of the macrofungi among these, such as *Baeospora myosura*, *Coprinopsis picacea*,
515 *Coprinus comatus*, *Cortinarius infractus*, *Crepidotus mollis*, *Geastrum fornicatum*, *Mycena renati*,
516 *Russula albonigra*, *Russula heterophylla*, *Russula sororia*, and *Tricholoma batschi*, have been found
517 in sporocarp surveys in the Északi-középhegység (Bohus & Babos 1960; Tóth 1999; Siller *et al.*
518 2002; Egri 2007; Rudolf *et al.* 2008; Siller & Dima 2014). With respect to the rest, more sampling is
519 needed to confirm their presence in the region.

520 Finally, soil communities are extremely diverse and there is increasing evidence pointing to
521 soil biodiversity as having key roles in determining the structure and ecological responses of
522 terrestrial ecosystems (Bardgett & van der Putten 2014). Soil fungi in particular are known to drive
523 plant diversity and productivity and are crucial for ecosystem functioning and resilience towards
524 disturbance (van der Heijden *et al.* 2008). Because most fungi have high habitat specificity and tend
525 to respond quickly to changes in environmental conditions (Nielsen *et al.* 2010; Geml *et al.* 2015,
526 2016; Morgado *et al.* 2015, 2016; Mundra *et al.* 2016; Semenova *et al.* 2015, 2016), fungi have a
527 promising potential as indicators of habitat quality in biological monitoring programs. More
528 specifically, assessments of richness and community composition of fungal communities in a variety
529 of habitats can inform decision-makers with respect to land use strategies that foster the sustainable
530 preservation of diverse and resilient ecosystems with a wide range of ecosystem functions.

531

532 *Conclusions*

533 A prominent finding of this study is that fungal diversity and community structure are
534 strongly influenced by slope aspect, despite the short distance separating north- and south-facing
535 slopes. Even though this finding may appear trivial due to the well-known effects of aspect on
536 vegetation, fungal communities are surprisingly little studied in this regard and the data presented

537 here offer unprecedented insights into the landscape-level distribution of various taxonomic and
538 functional groups with respect to topography. Furthermore, while aspect-driven differences in
539 vegetation tend to be related to relative importance as opposed the presence/absence of plant species
540 on slopes with different aspect (Gilliam et al. 2014), many fungal species were detected exclusively
541 on either north- or south-facing slopes. This reflects the often-observed high habitat specificity
542 exhibited by many fungi, which offers possibilities for biological monitoring and habitat
543 characterization and I strongly advocate for incorporating fungi in biodiversity assessments and
544 conservation efforts.

545

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556

557 REFERENCES

- 558 Albert L, Dima B. 2005 Ritka nagygombafajok (Basidiomycetes) előfordulása Magyarországon I.
559 *Mikológiai Közlemények Clusiana* **44**: 3–22.
- 560
- 561 Bardgett RC, van der Putten WH. 2014. Belowground biodiversity and ecosystem functioning.
562 *Nature* **515**: 505–511.
- 563
- 564 Bohus G, Babos M. 1960. Notes on the occurrence in Hungary of *Russula* species, with regard to
565 their range in Europe. *Annales Historico-Naturales Musei Nationalis Hungarici* **52**: 123–146.
- 566
- 567 Borcard D, Legendre P, Drapeau P. 1992. Partialling out the spatial component of ecological
568 variation. *Ecology* **73**: 1045–1055.
- 569
- 570 Borhidi A. 2003. *Magyarország növénytársulásai*. Akadémiai Kiadó, Budapest.

571

572 Bölöni J, Botta-Dukát Z, Illyés E, Molnár Zs. 2011. Hungarian landscape types: classification of
573 landscapes based on the relative cover of (semi-)natural habitats. *Applied Vegetation Science* **14**:
574 537–546.

575

576 Bratek Z, Merényi Zs, Varga T. 2013. Changes of hypogeous funga in the Carpathian-Pannonian
577 region in the past centuries. *Acta Mycologica* **48**: 33–39.

578

579 Calef MP, McGuire D, Epstein H, Rupp ST, Shugart HH. 2005. Analysis of vegetation distribution
580 in Interior Alaska and sensitivity to climate change using a logistic regression approach. *Journal of*
581 *Biogeography* **32**: 863–878.

582

583 Cavender-Bares J, Izzo A, Robinson R, Lovelock CE. 2009. Changes in ectomycorrhizal community
584 structure on two containerized oak hosts across an experimental hydrologic gradient. *Mycorrhiza* **19**:
585 133–142.

586

587 Chu H, Xiang X, Yang J, Adams JM, Zhang K, Li Y, Shi Y. 2016. Effects of slope aspects on soil
588 bacterial and arbuscular fungal communities in a boreal forest in China. *Pedosphere* **26**: 226–234.

589

590 Coughlan AP, Dalpé Y, Lapoint L, Piché Y. 2000. Soil pH-induced changes in root colonization,
591 diversity, and reproduction of symbiotic arbuscular mycorrhizal fungi from healthy and declining
592 maple forests. *Canadian Journal of Forest Research* **30**: 1543–1554.

593

594 Crowther TW, Maynard DS, Crowther TR, Peccia J, Smith JR, Bradford MA. 2014. Untangling the
595 fungal niche: a trait-based approach. *Frontiers in Microbiology* doi:10.3389/fmicb.2014.00579

596

597 Dobos E. 2010. A Mátravidék talajai. In: Baráz Cs. (Ed.) *A Mátrai Tájvédelmi Körzet – Heves és*
598 *Nógrád határán*. Bükk Nemzeti Park Igazgatóság, pp. 141–147.

599

600 Dufrière M, Legendre P. 1997. Species assemblages and indicator species: the need for a flexible
601 asymmetrical approach. *Ecological Monographs* **67**: 345–366.

602

603 Edgar, R.C. 2010 Search and clustering orders of magnitude faster than BLAST. *Bioinformatics* **26**:
604 2460–2461

605

606 Egri K. 2007. Újabb adatok a Zempléni-hegység és a Bodrogköz veszélyeztetett nagygombáiról.
607 *Mikológiai Közlemények Clusiana* **46**: 149–164.

608

609 Erős-Honti Zs, Kovács GM, Szedlay Gy, Jakucs E. 2008. Morphological and molecular
610 characterization of *Humaria* and *Genea* ectomycorrhizae from Hungarian deciduous forests.
611 *Mycorrhiza* **18**: 133–143.

612

613 Fekedulegn D, Hicks RR, Colbert JJ. 2003. Influence of topographic aspect, precipitation and
614 drought on radial growth of four major tree species in an Appalachian watershed. *Forest Ecology*
615 *and Management* **177**: 409–425.

616

617 Fekete G, Király G, Molnár Z. 2016. Delineation of the Pannonian vegetation region. *Community*
618 *Ecology* **17**: 114–124.

619

620 Gardes M, Bruns T. 1996. Community structure of ectomycorrhizal fungi in a *Pinus muricata* forest:
621 above- and below-ground views. *Canadian Journal of Botany* **74**: 1572–1583.

622

623 Gehring CA, Mueller RC, Haskins KE, Rubow TK, Whitham TG. 2014. Convergence in
624 mycorrhizal fungal communities due to drought, plant competition, parasitism, and susceptibility to
625 herbivory: consequences for fungi and host plants. *Frontiers in Microbiology*
626 doi:10.3389/fmicb.2014.00306

627

628 Geiger R. 1965. *The climate near the ground*. Harvard University Press, Cambridge, MA.

629

630 Geml J, Gravendeel B, Neilen M, Lammers Y, Raes N, Semenova TA, Noordeloos ME. 2014a.
631 DNA metabarcoding reveals high fungal diversity and pH-correlated habitat partitioning in protected
632 coastal *Salix repens* communities in the Netherlands. *PLOS ONE* 9(6): e99852.

633

634 Geml J, Morgado LN, Semenova-Nelsen TA, Schilthuizen M. 2017. Changes in richness and
635 community composition of ectomycorrhizal fungi among altitudinal vegetation types on Mount
636 Kinabalu in Borneo. *New Phytologist* **215**: 454–468.

637

- 638 Geml J, Morgado LN, Semenova TA, Welker JM, Walker MD, Smets E. 2015. Long-term warming
639 alters richness and composition of taxonomic and functional groups of arctic fungi. *FEMS*
640 *Microbiology Ecology* doi: 10.1093/femsec/fiv095
641
- 642 Geml J, Pastor N, Fernandez L, Pacheco S, Semenova TA, Becerra AG Wicaksono CY, Nouhra ER.
643 2014b. Large-scale fungal diversity assessment in the Andean Yungas forests reveals strong
644 community turnover among forest types along an altitudinal gradient. *Molecular Ecology* **23**: 2452–
645 2472
646
- 647 Geml J, Semenova TA, Morgado LN, Welker JM. 2016. Changes in composition and abundance of
648 functional groups of arctic fungi in response to long-term summer warming. *Biology Letters*
649 **12**:20160503
650
- 651 Geml J, Timling I, Robinson CH, Lennon N, Nusbaum HC, Brochmann C, Brochmann C,
652 Noordeloos ME, Taylor DL. 2012. An arctic community of symbiotic fungi assembled by long-
653 distance dispersers: phylogenetic diversity of ectomycorrhizal basidiomycetes in Svalbard based on
654 soil and sporocarp DNA. *Journal of Biogeography* **39**: 74–88.
655
- 656 Gilliam FS, Hédli R, Chudomelová M, McCulley RL. 2014. Variation in vegetation and microbial
657 linkages with slope aspect in a montane temperate hardwood forest. *Ecosphere* **5**: 1–17.
658
- 659 Glassman SI, Wang IJ, Bruns TD. 2017. Environmental filtering by pH and soil nutrients drives
660 community assembly in fungi at fine spatial scales. *Molecular Ecology* **26**: 6960–6973.
661
- 662 Holland PG, Steyn DG. 1975. Vegetational responses to latitudinal variations in slope angle and
663 aspect. *Journal of Biogeography* **2**: 179–183.
664
- 665 Hollingsworth TN, Walker MD, Chapin III FS, Parsons AL. 2006. Scale-dependent environmental
666 controls over species composition in Alaskan black spruce communities. *Canadian Journal of Forest*
667 *Research* **36**: 1781–1796.
668
- 669 Horváth G, Gaálová K. 2007. Éghajlati viszonyok. In: Kiss G, Baráz Cs, Gaálová K, Judik B. (Eds.)
670 *A Karancs-Medves és Cseres-hegység Tájvédelmi Körzet – Nógrád és Gömör határán*. Bükki
671 Nemzeti Park Igazgatóság, pp. 91–92.

672

673

674 Ihrmark K, Bödeker ITM, Cruz-Martinez K, Friberg H, Kubartova A, Schenk J, Strid Y, Stenlid J,
675 Brandstöm-Durling M, Clemmensen KE, *et al.* 2012. New primers to amplify the fungal ITS2
676 region -- evaluation by 454-sequencing of artificial and natural communities. *FEMS Microbiology*
677 *Ecology* **82**: 666–677.

678

679 Jakucs E, Erős-Honti Zs, Seress D, Kovács GM 2015. Enhancing our understanding of anatomical
680 diversity in *Tomentella* ectomycorrhizas characterization of six new morphotypes. *Mycorrhiza* **25**:
681 419–429.

682

683 Kovács GM, Jakucs E. 2006. Morphological and molecular comparison of white truffle
684 ectomycorrhizae. *Mycorrhiza* **16**: 567–574.

685

686 Kõljalg U, Dahlberg A, Taylor AFS, Larsson E, Hallenberg N, Stenlid J, Larsson KH, Fransson PM,
687 Kårén O, Johnsson L. 2000. Diversity and abundance of resupinate theleporoid fungi as
688 ectomycorrhizal symbionts in Swedish boreal forests. *Molecular Ecology* **9**: 1985–1996.

689

690 Kõljalg U, Nilsson RH, Abarenkov K, Tedersoo L, Taylor AFS, Bahram M, Bates ST, Bruns TD,
691 Bengtsson-Palme J, Callaghan TM, *et al.* 2013. Towards a unified paradigm for sequence-based
692 identification of Fungi. *Molecular Ecology* **22**: 5271–5277.

693

694 Lauber CL, Strickland MS, Bradford MA, Fierer N. 2008. The influence of soil properties on the
695 structure of bacterial and fungal communities across land-use types. *Soil Biology and Biochemistry*
696 **40**: 2407–2415.

697

698 Legendre P, Borcard D, Peres-Neto PR. 2005 Analyzing beta diversity: partitioning the spatial
699 variation of community composition data. *Ecological Monographs* **75**: 435–450.

700

701 Lennon JT, Aanderud ZT, Lehmkuhl BK, Schoolmaster DR. 2012. Mapping the niche space of soil
702 microorganisms using taxonomy and traits. *Ecology* **93**: 1867–1879.

703

- 704 Lindahl BD, Nilsson RH, Tedersoo L, Abarenkov K, Carlsen T, Kjøller, Kõljalg U, Pennanen T,
705 Rosendahl S, Stenlid J *et al.* 2013. Fungal community analysis by high-throughput sequencing of
706 amplified markers – a user's guide. *New Phytologist* **199**: 288–299.
707
- 708 Liu M, Zheng R, Bai S, Bai Y, Wang J. 2017. Slope aspect influences arbuscular mycorrhizal fungus
709 communities in arid ecosystems of the Daqingshan Mountains, Inner Mongolia, North China.
710 *Mycorrhiza* **27**: 189–200.
711
- 712 McCune B, Grace JB. 2002. *Analysis of ecological communities*. Glenden Beach, OR, USA: MjM
713 Software.
714
- 715 Méndez-Toribio M, Meave JA, Zemeño-Hernández I, Ibarra-Manríquez G. 2016. Effects of slope
716 aspect and topographic position on environmental variables, disturbance regime and tree community
717 attributes in a seasonal tropical dry forest. *Journal of Vegetation Science* **27**: 1094–1103.
718
- 719 Morgado LN, Semanova TA, Welker JM, Walker MD, Smets E, Geml J. 2015. Summer temperature
720 increase has distinct effects on the ectomycorrhizal fungal communities of moist tussock and dry
721 tundra in Arctic Alaska. *Global Change Biology* **21**: 959–972.
722
- 723 Morgado LN, Semanova TA, Welker JM, Walker MD, Smets E, Geml J. 2016. Long-term increase
724 in snow depth leads to compositional changes in arctic ectomycorrhizal fungal communities. *Global*
725 *Change Biology* **22**: 3080–3096
726
- 727 Mundra S, Halvorsen R, Kauserud H, Bahram M, Tederoo L, Elberling B, Cooper EJ, Eidesen PB.
728 2016. Ectomycorrhizal and saprotrophic fungi respond differently to long-term experimentally
729 increased snow depth in the High Arctic. *Microbiology Open* **5**: 856–869.
730
- 731 Nevarez L, Vasseur V, Le Madec L, Le Bras MA, Coroller L, Leguérinel I, Barbier G. 2009.
732 Physiological traits of *Penicillium glabrum* strain LCP 08.5568, a filamentous fungus isolated from
733 bottled aromatised mineral water. *International Journal of Food Microbiology* **130**: 166–171.
734
- 735 Nguyen NH, Song Z, Bates ST, Branco S, Tedersoo L, Menke J, Schilling JS, Kennedy PG. 2015.
736 FUNGuild: an open annotation tool for parsing fungal community datasets by ecological guild.
737 *Fungal Ecology* **20**: 241–248.

738

739 Nielsen UN, Osler GHR, Campbell CD, Burslem DFRP, van der Wal R. 2010. The influence of
740 vegetation type, soil properties and precipitation on the composition of soil mite and microbial
741 communities at the landscape scale. *Journal of Biogeography* **37**: 1317–1328.

742

743 Oksanen J, Blanchet FGK, Roeland Legendre P, Minchin PR, O'Hara RB, Simpson GL, Solymos P,
744 Stevens MHH, Wagner H. 2012. *Vegan: Community Ecology Package*. Version 2.0-5. [http://cran.r-](http://cran.r-project.org)
745 [project.org](http://vegan.rforge.r-project.org/). <http://vegan.rforge.r-project.org/>.

746

747 Pál-Fám F, Siller I, Fodor L. 2007. Mycological monitoring in the Hungarian Biodiversity
748 Monitoring System. *Acta Mycologica* **42**: 35–58.

749

750 Peay KG, Kennedy PG, Talbot JM. 2016. Dimensions of biodiversity in the Earth mycobiome.
751 *Nature Reviews Microbiology* **14**: 434–447.

752

753 Pelikán P. 2010. A Mátra és közvetlen környezetének földtana. In: Baráz Cs. (Ed.) *A Mátrai*
754 *Tájvédelmi Körzet – Heves és Nógrád határán*. Bükk Nemzeti Park Igazgatóság, pp. 17–26.

755

756 Porter WM, Robson AD, Abbott LK. 1987. Field survey of the distribution of Vesicular-arbuscular
757 mycorrhizal fungi in relation to soil pH. *Journal of Applied Ecology* **24**: 659–662.

758

759 Postma JWM, Olsson PA, Falkengren-Grerup U. 2006. Root colonization by arbuscular mycorrhizal,
760 fine endophytic and dark septate fungi across a pH gradient in acid beech forests. *Soil Biology and*
761 *Biochemistry* **39**: 400–408.

762

763 R Development Core Team, 2015. *R: a Language and Environment for Statistical Computing*. R
764 Foundation for Statistical Computing. Vienna, Austria. [http:// www.R-project.org](http://www.R-project.org)

765

766 Read DJ. 1991. Mycorrhizas in ecosystems. *Experientia* **47**: 376–391.

767

768 Rimóczi I. 1992. A Tarnavölgyi erdők nagygombái. *Folia Historico Naturalis Musei Matraensis* **17**:
769 131–138.

770

- 771 Rimóczi I. 1994. Nagygombáink cönológiai és ökológiai jellemzése. *Mikológiai Közlemények*
772 *Clusiana* **33**: 3–180.
- 773
- 774 Rosenberg NJ, Blad SB, Verma SB. 1983. *Microclimate: the biological environment*. Wiley, New
775 York.
- 776
- 777 Rousk J, Bååth E, Brookes PC, Lauber CL, Lozupone C, Caporaso JG, Knight R, Fierer N. 2010.
778 Soil bacterial and fungal communities across a pH gradient in an arable soil. *The ISME Journal* **4**:
779 1340–1351.
- 780
- 781 Rudolf K, Pál-Fám F, Morschhauser T. 2008. A Cserehát nagygombái. *Mikológiai Közlemények*
782 *Clusiana* **47**: 45–74.
- 783
- 784 Semenova TA, Morgado LN, Welker JM, Walker MD, Smets E, Geml J. 2015. Long-term
785 experimental warming alters community composition of ascomycetes in Alaskan moist and dry arctic
786 tundra. *Molecular Ecology* **24**: 424–437.
- 787
- 788 Semenova TA, Morgado LN, Welker JM, Walker MD, Smets E, Geml J. 2016. Compositional and
789 functional shifts in arctic fungal communities in response to experimentally increased snow depth.
790 *Soil Biology and Biochemistry* **100**: 201–209.
- 791
- 792 Siller I. 2010. A gombák világa. In: Baráz Cs. (Ed.) *A Mátrai Tájvédelmi Körzet – Heves és Nógrád*
793 *határán*. Bükki Nemzeti Park Igazgatóság, pp. 175–180.
- 794
- 795 Siller I, Dima B. 2014. Adatok a Heves-Borsodi-dombság és az Upponyi-hegység nagygombáihoz.
796 In: Diczházi I., Schmotzer A. (Eds.) *Apoka – A Heves-Borsodi-dombság és az Upponyi-hegység*
797 *élővilága*. Bükki Nemzeti Park Igazgatóság, pp. 35–54.
- 798
- 799 Siller I, Dima B, Albert L, Vasas G, Fodor L, Pál-Fám F, Bratek Z, Zagyva I. 2006. Védett
800 nagygombafajok Magyarországon. *Mikológiai Közlemények Clusiana* **45**: 3–158.
- 801
- 802 Siller I, Turcsányi G, Maglóczky Zs, Czájlik P. 2002. Lignicolous macrofungi of the Kékes North
803 forest reserve in the Mátra Mountains, Hungary. *Acta Microbiologica et Immunologica Hungarica*
804 **49**: 193–205.

805

806 Siller I, Vasas G, Pál-Fám F, Bratek Z, Zagyva I, Fodor L. 2002. Hungarian distribution of the
807 legally protected macrofungi species. *Studia Botanica Hungarica* **36**: 131–163.

808

809 Smith TW, Lundholm JT. 2010. Variation partitioning as a tool to distinguish niche and neutral
810 processes. *Ecography* **33**: 648–655.

811

812 Sparks DL, Page AL, Helmke PA, Loeppert RH. 1996. *Methods of Soil Analysis Part 3 – Chemical*
813 *Methods*. Soil Society of America Book Series, American Society of Agronomy, Inc, Madison,
814 Wisconsin.

815

816 Sternberg M, Shoshany M. 2001. Influence of slope aspect on Mediterranean woody formations:
817 comparison of a semiarid and an arid site in Isreal. *Ecological Research* **16**: 335–345.

818

819 Suba J. 1983. A Bükk növényei. In: Sándor A. (ed.) *Bükk Nemzeti Park – Kilátás kövekről*.
820 Mezőgazdasági Kiadó, pp. 189–235.

821

822 Sundseth K. 2009. *Natura 2000 in the Pannonian region*. European Commission, Environment
823 Directorate General, Brussels.

824 (<http://ec.europa.eu/environment/nature/info/pubs/docs/biogeos/pannonian.pdf>)

825

826 Takács B, Siller I. 1980. A Bükk hegységi Ősbükkös gombái. *Mikológiai Közlemények, Clusiana* **3**:
827 121–132.

828

829 Tedersoo L, Bahram M, Pölme S, Kõljalg U, Yorou NS, Wijensundera R, Ruiz LV, Vasco-Palacios
830 AM, Thu QP, Suija A, *et al.* 2014. Global diversity and geography of soil fungi. *Science* **346**:

831 1256688

832

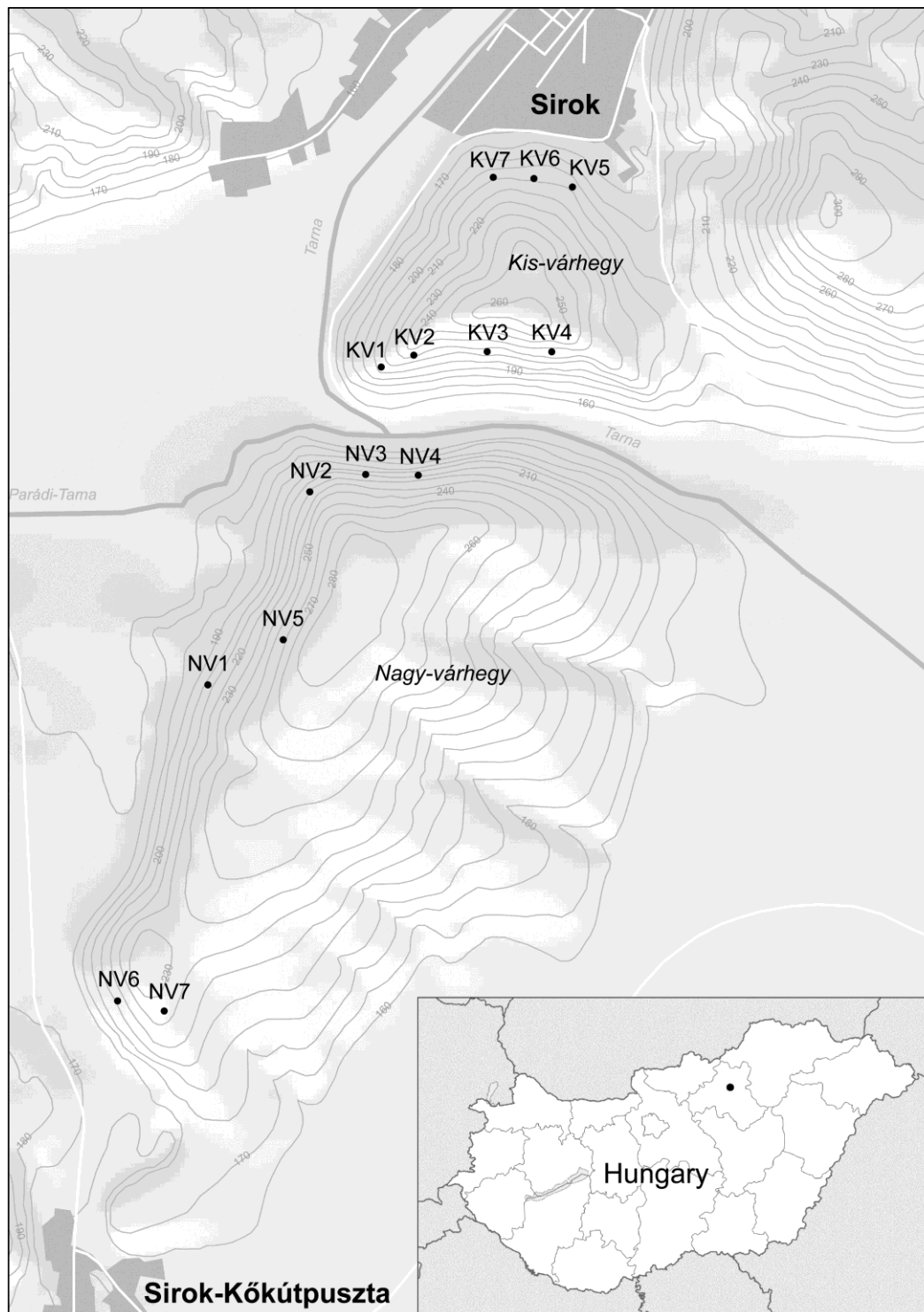
833 Tóth B. 1999. Adatok a Gyepes-völgy (Heves-Borsodi dombág) nagygombáiról. *Kitaibelia* **4**: 261–
834 270.

835

836 Tóth G. 1983. A bükk karszt vízrendszere. In: Sándor A. (ed.) *Bükk Nemzeti Park – Kilátás*
837 *kövekről*. Mezőgazdasági Kiadó, pp. 107–134.

838

- 839 van der Heijden MGA, Bardgett RD, van Straalen NM. 2008. The unseen majority: soil microbes as
840 drivers of plant diversity and productivity in terrestrial ecosystems. *Ecology Letters* **11**:296–310.
841
- 842 Vojtkó A. 2002. A hegység növénytakarója. In: Baráz Cs. (Ed.) *A Bükk Nemzeti Park – Hegyek,*
843 *erdők, emberek.* Bükk Nemzeti Park Igazgatóság, pp. 237–261.
844
- 845 Vojtkó A, Sramkó G, Magos G, Harnos K. 2010. Növényvilág. In: Baráz Cs. (Ed.) *A Mátrai*
846 *Tájvédelmi Körzet – Heves és Nógrád határán.* Bükk Nemzeti Park Igazgatóság, pp. 149–174.
847
- 848 Walker MD, Walker DA, Auerbach NA. 1994. Plant communities of a tussock tundra landscape in
849 the Brooks Range Foothills, Alaska. *Journal of Vegetation Science* **5**: 843–866.
850
- 851 Wheeler KA, Hurdman BF, Pitt JI. 1991. Influence of pH on the growth of some toxigenic species of
852 *Aspergillus*, *Penicillium* and *Fusarium*. *International Journal of Food Microbiology*, **12**: 141–150.
853
- 854 White TM, Bruns, T, Lee S, Taylor J. 1990. Amplification and direct sequencing of fungal ribosomal
855 RNA for phylogenetics. In: Innis MA, Gelfand DH, Sninsky JJ, White TJ (Eds.) *PCR protocols: a*
856 *guide to methods and applications.* Academic Press, San Diego, CA, pp. 315–321.
857
- 858 Whittaker RH. 1956. Vegetation of the Great Smoky Mountains. *Ecological Monographs* **26**: 1–80.
859
860
861

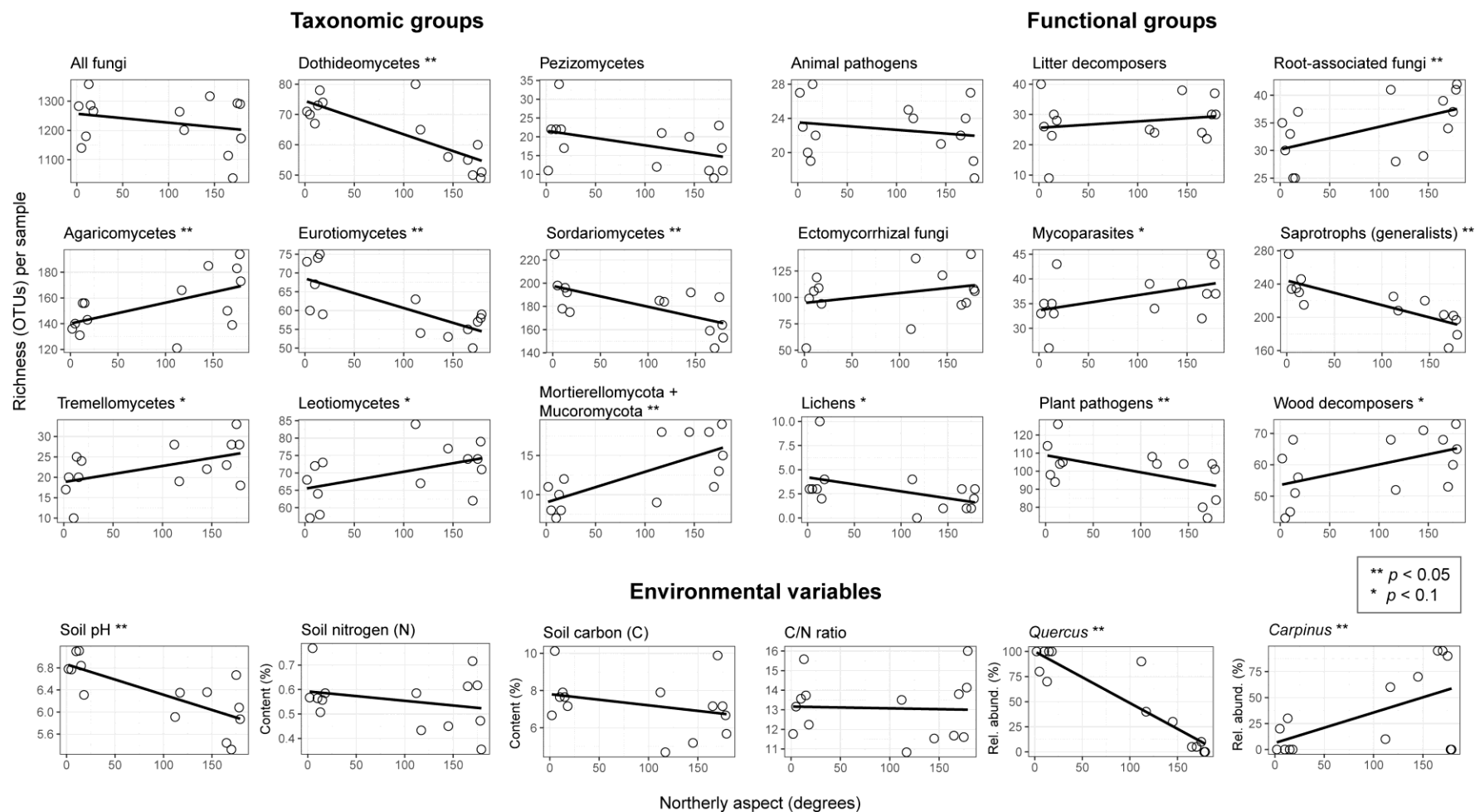


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863 Fig. 1. A map of the sampling localities, with the location of the region of study in Hungary (inset).

864 Full names, vegetation types, topographic variables, and geographic coordinates corresponding to the

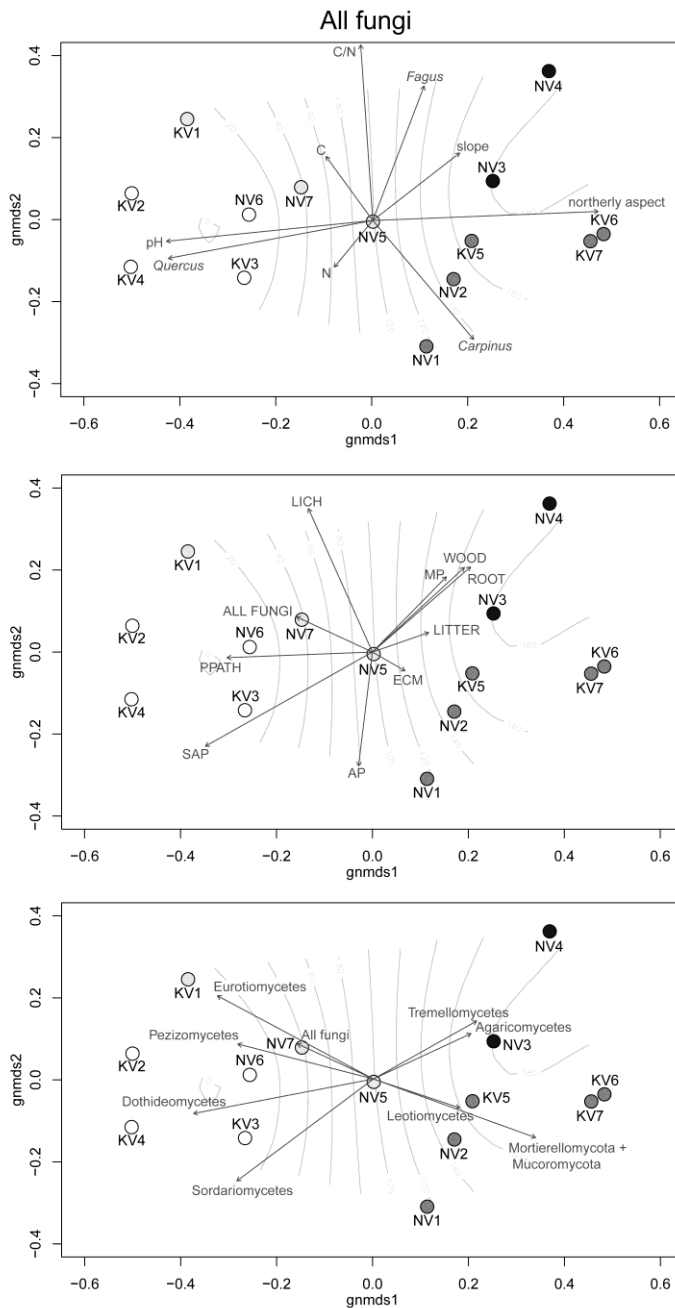
865 sampling localities are listed in Table 1.



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868 Fig. 2. Correlations between northerly aspect and operational taxonomic unit (OTU) richness of functional and taxonomic groups as well as
 869 environmental variables explored using linear regression. Significant ($p < 0.05$) and marginally significant ($p < 0.1$) correlations are marked by
 870 ** and *, respectively.

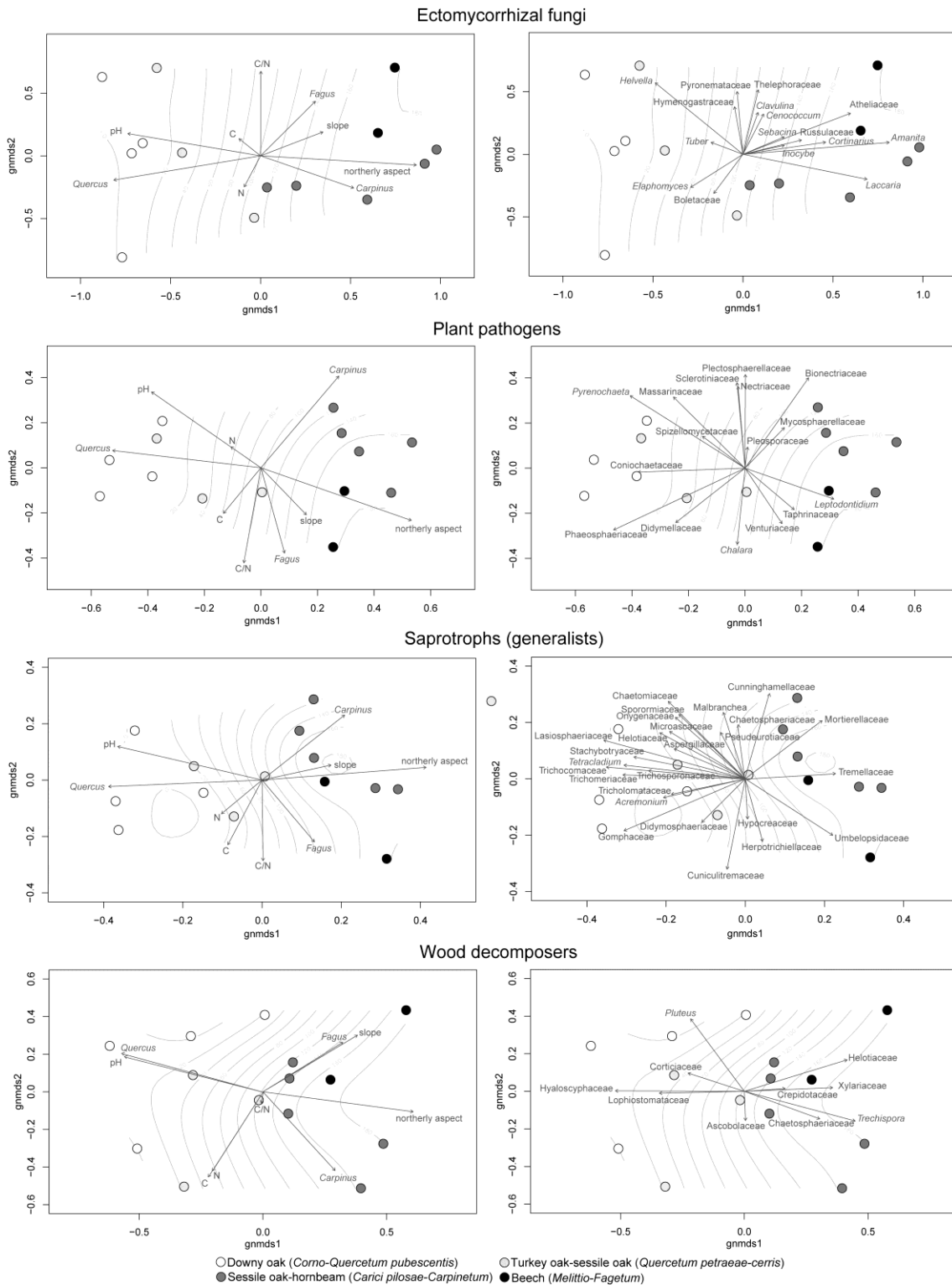


○Downy oak (*Corno-Quercetum pubescentis*) ○Turkey oak-sessile oak (*Quercetum petraeae-cerris*)
 ●Sessile oak-hornbeam (*Carici pilosae-Carpinetum*) ●Beech (*Melittio-Fagetum*)

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872

873 Fig. 3. Non-metric multidimensional scaling (NMDS) ordination plots of the fungal communities in
 874 the sampled forest types based on Hellinger-transformed data, with northerly aspect displayed as
 875 isolines. Labels, localities and descriptions of the sampling sites are given in Table 1. Vectors of
 876 environmental variables and richness of functional and taxonomic groups of fungi correlated with
 877 ordination axes are displayed in three identical ordination plots. Abbreviations for functional guilds:
 878 AP = animal pathogen, ECM = ectomycorrhizal fungus, MP = mycoparasite, PPATH = plant
 879 pathogen, ROOT = root-associated (non-ECM) fungus, SAP = generalist saprotroph, WOOD = wood
 880 decomposer.



881

882

883 Fig. 4. Non-metric multidimensional scaling (NMDS) ordination plots of the four more diverse
 884 functional groups of fungi in the sampled forest types based on Hellinger-transformed data, with
 885 northerly aspect displayed as isolines. Vectors of environmental variables and OTU richness of
 886 fungal families correlated with ordination axes are displayed. Where a certain family was represented
 887 by only one genus, the genus name is displayed.

888 Table 1. Sampling sites included in this study with code, locality, forest type, dominant tree species, slope aspect, northerly aspect expressed in
 889 degrees (south=0, north=180), slope angle, and geographic coordinates. Abbreviations for tree genera are: *C.*: *Carpinus* (Betulaceae), *F.*: *Fagus*
 890 (*Fagaceae*), and *Q.*: *Quercus* (*Fagaceae*). Locations are displayed in a map in Figure 1.

891

Site code	Locality	Forest type	Dominant tree species	Aspect	Northerly aspect (degrees)	Slope angle (degrees)	Latitude (degrees)	Longitude (degrees)
KV1	Kis-várhegy	<i>Quercetum petraeae-cerris</i>	<i>Q. dalechampii, Q. petraea, Q. cerris</i>	SW	13	22.85	47.918885	20.187345
KV2	Kis-várhegy	<i>Corno-Quercetum pubescentis</i>	<i>Q. dalechampii, Q. petraea, Q. cerris</i>	S	5	5.71	47.918840	20.188193
KV3	Kis-várhegy	<i>Corno-Quercetum pubescentis</i>	<i>Q. dalechampii, Q. petraea, Q. cerris</i>	S	2	29.71	47.919033	20.190292
KV4	Kis-várhegy	<i>Corno-Quercetum pubescentis</i>	<i>Q. dalechampii, Q. petraea, Q. cerris</i>	S	10	24	47.918927	20.192524
KV5	Kis-várhegy	<i>Carici pilosae-Carpinetum</i>	<i>C. betulus, Q. petraea</i>	N	175	19.43	47.921958	20.193339
KV6	Kis-várhegy	<i>Carici pilosae-Carpinetum</i>	<i>C. betulus, Q. petraea</i>	N	170	22.86	47.921977	20.192281
KV7	Kis-várhegy	<i>Carici pilosae-Carpinetum</i>	<i>C. betulus, Q. petraea</i>	N	165	25.14	47.921792	20.191034
NV1	Nagy-várhegy	<i>Carici pilosae-Carpinetum</i>	<i>C. betulus, Q. petraea</i>	WNW	115	17.14	47.912658	20.181935
NV2	Nagy-várhegy	<i>Carici pilosae-Carpinetum</i>	<i>C. betulus, Q. petraea</i>	NW	145	33.14	47.916270	20.184321
NV3	Nagy-várhegy	<i>Melittio-Fagetum</i>	<i>F. sylvatica</i>	N	178	37.71	47.917315	20.186580
NV4	Nagy-várhegy	<i>Melittio-Fagetum</i>	<i>F. sylvatica</i>	N	179	38.85	47.917191	20.188443
NV5	Nagy-várhegy	<i>Quercetum petraeae-cerris</i>	<i>Q. dalechampii, Q. petraea, Q. cerris</i>	WNW	115	22.85	47.913477	20.183917
NV6	Nagy-várhegy	<i>Corno-Quercetum pubescentis</i>	<i>Q. dalechampii, Q. petraea, Q. cerris</i>	SW	15	24.28	47.906740	20.180339
NV7	Nagy-várhegy	<i>Quercetum petraeae-cerris</i>	<i>Q. dalechampii, Q. petraea, Q. cerris</i>	SW	18	8	47.906650	20.181740

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893

894 Table 2. Fungal OTUs considered as significant indicators of northerly (N) or southerly (S) slope aspects with corresponding p -values, assigned
 895 functional guild, matching Species Hypothesis, ITS2 rDNA sequence similarity (%) and taxonomic classification of the most similar matching
 896 sequence in the UNITE+INSD dynamic Species Hypotheses database (version released on October 10, 2017). Only indicators with a known
 897 function are shown, displayed in the order of aspect, functional guild, and decreasing significance of indicator value. Abbreviations for
 898 functional guilds: AP = animal pathogen, ECM = ectomycorrhizal fungus, MP = mycoparasite, PPATH = plant pathogen, ROOT = root-
 899 associated (non-ECM) fungus, SAP = generalist saprotroph, WOOD = wood decomposer.

900

OTU	Aspect	p	Function	Species Hypothesis	%	Matching taxon	Phylum	Class
OTU_2513	S	0.0116	AP	SH007609.07FU	95.5	<i>Arthroderma insingulare</i>	Ascomycota	Eurotiomycetes
OTU_892	S	0.0186	AP	SH012485.07FU	99.4	<i>Beauveria</i> sp.	Ascomycota	Sordariomycetes
OTU_4940	S	0.0192	AP	SH193551.07FU	95.6	<i>Metacordyceps chlamyosporia</i>	Ascomycota	Sordariomycetes
OTU_423	S	0.0262	AP	SH196109.07FU	97.5	<i>Simplicillium minatense</i>	Ascomycota	Sordariomycetes
OTU_1990	S	0.0488	AP	SH214395.07FU	100	<i>Metarhizium flavoviride</i>	Ascomycota	Sordariomycetes
OTU_178	S	0.0036	ECM	SH182661.07FU	98.8	<i>Melanogaster ambiguus</i>	Basidiomycota	Agaricomycetes
OTU_462	S	0.0036	ECM	SH182459.07FU	99.4	<i>Scleroderma areolatum</i>	Basidiomycota	Agaricomycetes
OTU_3	S	0.0046	ECM	SH220826.07FU	100	<i>Russula insignis</i>	Basidiomycota	Agaricomycetes
OTU_674	S	0.0068	ECM	SH629220.07FU	95.5	<i>Helvella</i> sp.	Ascomycota	Pezizomycetes
OTU_20	S	0.0144	ECM	SH214605.07FU	96.9	<i>Sebacina incrustans</i>	Basidiomycota	Agaricomycetes
OTU_722	S	0.02	ECM	SH497502.07FU	98.2	<i>Helvella latispora</i>	Ascomycota	Pezizomycetes
OTU_219	S	0.0208	ECM	SH222268.07FU	99.4	<i>Membranomyces spurius</i>	Basidiomycota	Agaricomycetes
OTU_138	S	0.0208	ECM	SH006539.07FU	100	<i>Helvella</i> sp.	Ascomycota	Pezizomycetes
OTU_112	S	0.0414	ECM	SH190455.07FU	98.3	<i>Astraeus telleriae</i>	Basidiomycota	Agaricomycetes
OTU_2018	S	0.0228	MP	SH495246.07FU	92.5	<i>Cosmospora</i> sp.	Ascomycota	Sordariomycetes
OTU_501	S	0.0012	PPATH	SH443332.07FU	96.3	<i>Fusarium concentricum</i>	Ascomycota	Sordariomycetes
OTU_541	S	0.0032	PPATH	SH198276.07FU	99.4	Nectriaceae sp.	Ascomycota	Sordariomycetes
OTU_435	S	0.0056	PPATH	SH175284.07FU	99.4	<i>Myrothecium</i> sp.	Ascomycota	Sordariomycetes
OTU_434	S	0.0062	PPATH	SH217882.07FU	99.4	<i>Pyrenochaeta</i> sp.	Ascomycota	Dothideomycetes
OTU_279	S	0.0068	PPATH	SH202579.07FU	96.9	<i>Pyrenochaeta</i> sp.	Ascomycota	Dothideomycetes
OTU_1610	S	0.0096	PPATH	SH177600.07FU	100	<i>Microbotryum anomalum</i>	Basidiomycota	Microbotryomycetes

OTU_1629	S	0.0108	PPATH	SH213880.07FU	99.4	<i>Gibberella tricineta</i>	Ascomycota	Sordariomycetes
OTU_4998	S	0.0112	PPATH	SH345969.07FU	97.5	<i>Stagonosporopsis dorenboschii</i>	Ascomycota	Dothideomycetes
OTU_1425	S	0.0154	PPATH	SH196352.07FU	100	<i>Spizellomyces pseudodichotomus</i>	Chytridiomycota	Spizellomycetes
OTU_2945	S	0.0192	PPATH	SH175285.07FU	95.7	<i>Myrothecium inundatum</i>	Ascomycota	Sordariomycetes
OTU_5594	S	0.0194	PPATH	SH191389.07FU	97.5	<i>Coniochaeta ligniaria</i>	Ascomycota	Sordariomycetes
OTU_525	S	0.0224	PPATH	SH174362.07FU	100	<i>Lectera longa</i>	Ascomycota	Sordariomycetes
OTU_2810	S	0.0246	PPATH	SH180770.07FU	100	<i>Helminthosporium solani</i>	Ascomycota	Dothideomycetes
OTU_1113	S	0.0032	ROOT	SH020834.07FU	96.3	<i>Phialocephala compacta</i>	Ascomycota	Leotiomycetes
OTU_1268	S	0.0238	ROOT	SH204998.07FU	94.3	<i>Acephala</i> sp.	Ascomycota	Leotiomycetes
OTU_890	S	0.0246	ROOT	SH008258.07FU	96.6	<i>Cadophora</i> sp.	Ascomycota	Leotiomycetes
OTU_748	S	0.0276	ROOT	SH216991.07FU	95.7	<i>Oidiodendron pilicola</i>	Ascomycota	Leotiomycetes
OTU_185	S	0.0012	SAP	SH180481.07FU	98.8	<i>Knufia tsunedae</i>	Ascomycota	Eurotiomycetes
OTU_1707	S	0.0012	SAP	SH176700.07FU	100	Hypocreales sp.	Ascomycota	Sordariomycetes
OTU_935	S	0.0012	SAP	SH180772.07FU	98.1	<i>Corynespora</i> sp.	Ascomycota	Dothideomycetes
OTU_3865	S	0.003	SAP	SH629325.07FU	94.9	<i>Chaetomium homopilatum</i>	Ascomycota	Sordariomycetes
OTU_368	S	0.003	SAP	SH216342.07FU	98.1	<i>Kavinia alboviridis</i>	Basidiomycota	Agaricomycetes
OTU_350	S	0.0068	SAP	SH213544.07FU	98.8	Hypocreales sp.	Ascomycota	Sordariomycetes
OTU_5475	S	0.0076	SAP	SH527119.07FU	98.1	<i>Cladophialophora</i> sp.	Ascomycota	Eurotiomycetes
OTU_367	S	0.0114	SAP	SH217071.07FU	100	<i>Podospora</i> sp.	Ascomycota	Sordariomycetes
OTU_388	S	0.0118	SAP	SH008915.07FU	98.2	Hypocreales sp.	Ascomycota	Sordariomycetes
OTU_1557	S	0.013	SAP	SH184177.07FU	98.3	<i>Preussia persica</i>	Ascomycota	Dothideomycetes
OTU_1071	S	0.015	SAP	SH640120.07FU	91.3	<i>Durella macrospora</i>	Ascomycota	Leotiomycetes
OTU_682	S	0.0154	SAP	SH021128.07FU	99.4	Hypocreales sp.	Ascomycota	Sordariomycetes
OTU_479	S	0.0188	SAP	SH205999.07FU	100	<i>Acrostalagmus luteoalbus</i>	Ascomycota	Sordariomycetes
OTU_5347	S	0.0192	SAP	SH527119.07FU	95.7	<i>Cladophialophora</i> sp.	Ascomycota	Eurotiomycetes
OTU_3468	S	0.0196	SAP	SH014661.07FU	96.9	Hypocreales sp.	Ascomycota	Sordariomycetes
OTU_96	S	0.0206	SAP	SH204873.07FU	99.4	<i>Chaetomium nigricolor</i>	Ascomycota	Sordariomycetes
OTU_1105	S	0.0208	SAP	SH019302.07FU	95	Hypocreales sp.	Ascomycota	Sordariomycetes
OTU_409	S	0.0216	SAP	SH201192.07FU	100	<i>Schizothecium carpinicola</i>	Ascomycota	Sordariomycetes
OTU_2009	S	0.022	SAP	SH003731.07FU	98.8	<i>Articulospora</i> sp.	Ascomycota	Leotiomycetes
OTU_3406	S	0.022	SAP	SH026623.07FU	100	<i>Trichoderma delicatulum</i>	Ascomycota	Sordariomycetes
OTU_1576	S	0.0224	SAP	SH007110.07FU	100	Hypocreales sp.	Ascomycota	Sordariomycetes

OTU_3912	S	0.0228	SAP	SH194607.07FU	97.2	<i>Penicillium christenseniae</i>	Ascomycota	Eurotiomycetes
OTU_1178	S	0.0228	SAP	SH183440.07FU	99.4	<i>Plectania melastoma</i>	Ascomycota	Pezizomycetes
OTU_1694	S	0.023	SAP	SH128133.07FU	95.7	<i>Microcera larvarum</i>	Ascomycota	Sordariomycetes
OTU_1502	S	0.0246	SAP	SH207979.07FU	96.4	<i>Acremonium persicinum</i>	Ascomycota	Sordariomycetes
OTU_1025	S	0.0332	SAP	SH197542.07FU	92.5	<i>Didymosphaeria</i> sp.	Ascomycota	Dothideomycetes
OTU_1444	S	0.0342	SAP	SH204465.07FU	94.4	Hypocreales sp.	Ascomycota	Sordariomycetes
OTU_680	S	0.0344	SAP	SH175276.07FU	95.2	Stachybotryaceae sp.	Ascomycota	Sordariomycetes
OTU_878	S	0.035	SAP	SH029684.07FU	90.4	<i>Arachnomycetes gracilis</i>	Ascomycota	Eurotiomycetes
OTU_1550	S	0.0366	SAP	SH030697.07FU	94.4	<i>Myrmecridium phragmitis</i>	Ascomycota	Sordariomycetes
OTU_425	S	0.0404	SAP	SH213541.07FU	97.2	Hypocreales sp.	Ascomycota	Sordariomycetes
OTU_165	S	0.0428	SAP	SH014661.07FU	98.8	Hypocreales sp.	Ascomycota	Sordariomycetes
OTU_1223	S	0.0442	SAP	SH194321.07FU	99.4	<i>Dictyochaeta simplex</i>	Ascomycota	Sordariomycetes
OTU_1337	S	0.0478	SAP	SH184969.07FU	90.1	Hypocreales sp.	Ascomycota	Sordariomycetes
OTU_1810	S	0.0012	WOOD	SH492473.07FU	96.3	<i>Monodictys capensis</i>	Ascomycota	Dothideomycetes
OTU_1642	S	0.0178	WOOD	SH216764.07FU	100	<i>Lophiostoma cynaroidis</i>	Ascomycota	Dothideomycetes
OTU_1194	S	0.039	WOOD	SH216909.07FU	95.7	<i>Arachnopeziza aurata</i>	Ascomycota	Leotiomycetes
OTU_108	N	0.0012	AP	SH184964.07FU	99.4	<i>Tolypocladium cylindrosporium</i>	Ascomycota	Sordariomycetes
OTU_895	N	0.0118	AP	SH192574.07FU	100	<i>Pochonia bulbillosa</i>	Ascomycota	Sordariomycetes
OTU_344	N	0.0034	ECM	SH642176.07FU	94.6	<i>Laccaria</i> sp.	Basidiomycota	Agaricomycetes
OTU_6	N	0.0052	ECM	SH214650.07FU	98.2	<i>Sebacina</i> sp.	Basidiomycota	Agaricomycetes
OTU_57	N	0.0056	ECM	SH186707.07FU	100	<i>Russula cyanoxantha</i>	Basidiomycota	Agaricomycetes
OTU_37	N	0.0056	ECM	SH204757.07FU	97.5	<i>Tarzetta</i> sp.	Ascomycota	Pezizomycetes
OTU_1741	N	0.0076	ECM	SH016378.07FU	100	<i>Piloderma</i> sp.	Basidiomycota	Agaricomycetes
OTU_60	N	0.0154	ECM	SH214626.07FU	99.4	<i>Sebacina</i> sp.	Basidiomycota	Agaricomycetes
OTU_67	N	0.0202	ECM	SH210482.07FU	95.4	<i>Paxillus cuprinus</i>	Basidiomycota	Agaricomycetes
OTU_182	N	0.0212	ECM	SH195878.07FU	100	<i>Inocybe maculata</i>	Basidiomycota	Agaricomycetes
OTU_450	N	0.0228	ECM	SH188589.07FU	100	<i>Cortinarius decipiens</i>	Basidiomycota	Agaricomycetes
OTU_15	N	0.0242	ECM	SH214721.07FU	96.8	<i>Sebacina</i> sp.	Basidiomycota	Agaricomycetes
OTU_312	N	0.0398	ECM	SH220557.07FU	99.4	<i>Russula chloroides</i>	Basidiomycota	Agaricomycetes
OTU_111	N	0.041	ECM	SH185240.07FU	100	<i>Tomentella</i> sp.	Basidiomycota	Agaricomycetes
OTU_2977	N	0.021	MP	SH190871.07FU	96.6	<i>Trichoderma harzianum</i>	Ascomycota	Sordariomycetes
OTU_1367	N	0.0212	MP	SH211287.07FU	98.8	Cephalothecaceae sp.	Ascomycota	Sordariomycetes

OTU_570	N	0.0368	MP	SH628832.07FU	100	<i>Tremella</i> sp.	Basidiomycota	Tremellomycetes
OTU_731	N	0.0404	MP	SH195954.07FU	100	<i>Hypomyces perniciosus</i>	Ascomycota	Sordariomycetes
OTU_43	N	0.0032	PPATH	SH211202.07FU	100	Bionectriaceae sp.	Ascomycota	Sordariomycetes
OTU_841	N	0.0052	PPATH	SH211202.07FU	96.3	Bionectriaceae sp.	Ascomycota	Sordariomycetes
OTU_699	N	0.0056	PPATH	SH218629.07FU	98.8	<i>Chalara</i> sp.	Ascomycota	Leotiomycetes
OTU_1254	N	0.0142	PPATH	SH194809.07FU	100	<i>Gnomonia virginiana</i>	Ascomycota	Sordariomycetes
OTU_1572	N	0.016	PPATH	SH213283.07FU	96.9	<i>Cladosporium</i> sp.	Ascomycota	Dothideomycetes
OTU_3715	N	0.018	PPATH	SH217194.07FU	98.3	Nectriaceae sp.	Ascomycota	Sordariomycetes
OTU_2484	N	0.0214	PPATH	SH187445.07FU	100	<i>Erysiphe arcuata</i>	Ascomycota	Leotiomycetes
OTU_2587	N	0.0242	PPATH	SH211202.07FU	90.9	Bionectriaceae sp.	Ascomycota	Sordariomycetes
OTU_340	N	0.0262	PPATH	SH209885.07FU	94.4	Nectriaceae sp.	Ascomycota	Sordariomycetes
OTU_870	N	0.0354	PPATH	SH116565.07FU	97.8	<i>Passalora californica</i>	Ascomycota	Dothideomycetes
OTU_5159	N	0.0382	PPATH	SH202969.07FU	96.9	<i>Nectria ramulariae</i>	Ascomycota	Sordariomycetes
OTU_387	N	0.0056	ROOT	SH207784.07FU	99.4	<i>Meliniomyces</i> sp.	Ascomycota	Leotiomycetes
OTU_5186	N	0.0056	ROOT	SH216990.07FU	93.8	<i>Oidiiodendron chlamydosporicum</i>	Ascomycota	Leotiomycetes
OTU_1785	N	0.0096	ROOT	SH217001.07FU	98.2	<i>Oidiiodendron</i> sp.	Ascomycota	Leotiomycetes
OTU_937	N	0.0134	ROOT	SH217004.07FU	99.4	<i>Oidiiodendron echinulatum</i>	Ascomycota	Leotiomycetes
OTU_970	N	0.0336	ROOT	SH181589.07FU	93.8	<i>Chloridium</i> sp.	Ascomycota	Sordariomycetes
OTU_489	N	0.0104	SAP	SH193250.07FU	98.2	<i>Cladophialophora</i> sp.	Ascomycota	Eurotiomycetes
OTU_1409	N	0.0118	SAP	SH013614.07FU	100	<i>Wardomyces humicola</i>	Ascomycota	Sordariomycetes
OTU_780	N	0.0128	SAP	SH180117.07FU	99.4	<i>Mortierella</i> sp.	Mortierellomycota	Mortierellomycetes
OTU_1191	N	0.0146	SAP	SH019878.07FU	90.1	<i>Curreya</i> sp.	Ascomycota	Dothideomycetes
OTU_2865	N	0.0192	SAP	SH199695.07FU	99.4	<i>Absidia cylindrospora</i>	Mucoromycota	Mucoromycetes
OTU_1876	N	0.0236	SAP	SH213264.07FU	97.5	<i>Cladophialophora chaetospora</i>	Ascomycota	Eurotiomycetes
OTU_332	N	0.024	SAP	SH196089.07FU	100	<i>Umbelopsis dimorpha</i>	Mucoromycota	Umbelopsidomycetes
OTU_655	N	0.0242	SAP	SH200349.07FU	90.8	<i>Malbranchea dendritica</i>	Ascomycota	Eurotiomycetes
OTU_1528	N	0.036	SAP	SH028610.07FU	98.8	<i>Leuconeurospora</i> sp.	Ascomycota	Leotiomycetes
OTU_471	N	0.048	SAP	SH216098.07FU	99.4	<i>Umbelopsis changbaiensis</i>	Mucoromycota	Umbelopsidomycetes
OTU_859	N	0.0042	WOOD	SH205445.07FU	100	<i>Nectriopsis rexiana</i>	Ascomycota	Sordariomycetes
OTU_1116	N	0.0052	WOOD	SH198389.07FU	97	<i>Hymenoscyphus fructigenus</i>	Ascomycota	Leotiomycetes

903 Table 3. Correlation of environmental and spatial variables with each other and with fungal community composition of the sampling sites.
 904 Correlations among standardized environmental variables and with Hellinger-transformed fungal community matrix were tested individually
 905 using Mantel tests showing correlation coefficients (r) and significance values (p). Non-significant results (n.s.) are not shown. In addition, to
 906 disentangle the ‘pure’ effects of environmental variables on fungal community composition, partial Mantel tests were used in all combinations of
 907 tested and controlled environmental variables.
 908

Correlation among environmental variables				
	aspect	slope	tree genera	edaphic factors
aspect	-			
slope	n.s.	-		
tree genera	$r=0.5256; p<0.0001$	n.s.	-	
edaphic factors	$r=0.1508; p=0.0461$	n.s.	$r=0.2661; p=0.0354$	-
spatial	n.s.	n.s.	n.s.	n.s.

Correlation with fungal community composition						
Tested variables	Mantel test	Control variables for partial mantel test				
		aspect	slope	tree genera	edaphic factors	spatial
aspect	$r=0.7429; p<0.0001$	-	$r=0.7395; p<0.0001$	$r=0.6368; p<0.0001$	$r=0.7928; p<0.0001$	$r=0.7449; p<0.0001$
slope	n.s.	n.s.	-	n.s.	n.s.	n.s.
tree genera	$r=0.5584; p<0.0001$	$r=0.2949; p=0.0178$	$r=0.5486; p=0.0002$	-	$r=0.5108; p=0.0002$	$r=0.5579; p=0.0001$
edaphic factors	$r=0.5347; p<0.0001$	$r=0.6388; p<0.0001$	$r=0.5247; p=0.0002$	$r=0.4828; p=0.0006$	-	$r=0.5397; p<0.0001$
spatial	n.s.	n.s.	n.s.	n.s.	n.s.	-

909

910 **Supporting Information:**

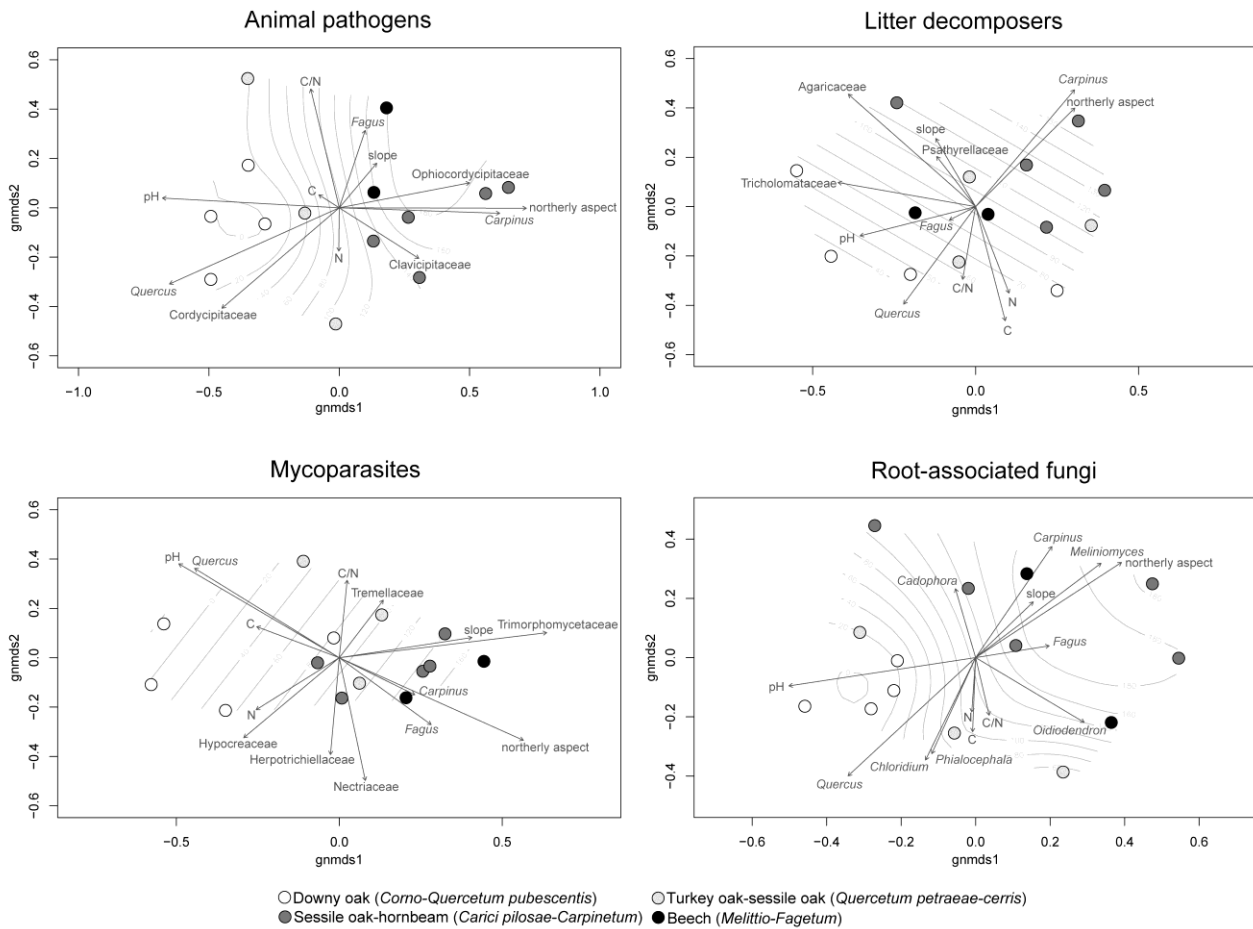


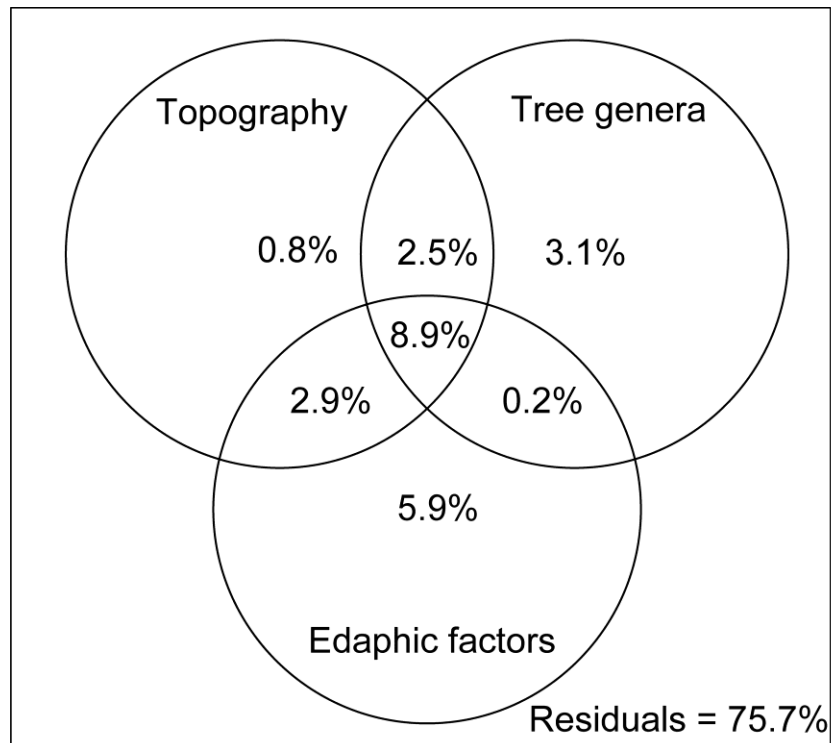
Fig. S1. Non-metric multidimensional scaling (NMDS) ordination plots of animal pathogenic, litter decomposer, mycoparasitic, and root-associated fungi in the sampled forest types based on Hellinger-transformed data. Vectors of environmental variables and OTU richness of fungal families correlated with ordination axes are displayed. Where a certain family was represented by only one genus, the genus name is shown.

911

912

913 Fig. S1. Non-metric multidimensional scaling (NMDS) ordination plots of animal pathogenic, litter
 914 decomposer, mycoparasitic, and root-associated fungi in the sampled forest types based on
 915 Hellinger-transformed data. Vectors of environmental variables and OTU richness of fungal families
 916 correlated with ordination axes are displayed. Where a certain family was represented by only one
 917 genus, the genus name is shown.

918



919

920

921 Fig. S2. The contribution of environmental variables to explaining the variation in fungal community
922 composition among the sampling sites as estimated by variation partitioning analyses. Tested
923 environmental variables included topography (aspect, slope), relative abundance of tree genera
924 (*Carpinus*, *Fagus*, *Quercus*), and edaphic factors (soil pH, C, N, and C/N)

925

926

927 Table S1. The full list of taxa corresponding to the 2542 unique SHs that matched the OTUs in the
928 collected samples with high (> 95%) sequence similarity, with % ITS2 rDNA sequence similarity to
929 the most similar OTU, Species Hypothesis code, and taxonomic classification. Taxa are ordered
930 according to phylum, class, order, family, genus, and the identified taxon name (provided upon
931 manuscript acceptance).