

# Strategies to Maximize the Wood Production in Amazon Forest

**Aline Canetti**

Independent forest consultant

**Evaldo Muñoz Braz**

Embrapa Florestas

**Patricia Mattos** (✉ [patricia.mattos@embrapa.br](mailto:patricia.mattos@embrapa.br))

Embrapa Florestas <https://orcid.org/0000-0003-4134-8890>

**Renato Olivir Basso**

Elabore Projetos e Consultoria Florestal

**Afonso Figueiredo Filho**

UNICENTRO: Universidade Estadual do Centro-Oeste

---

## Research

**Keywords:** dendrochronology, forest modeling, minimum cutting diameter

**Posted Date:** October 1st, 2020

**DOI:** <https://doi.org/10.21203/rs.3.rs-83620/v1>

**License:**  This work is licensed under a Creative Commons Attribution 4.0 International License.

[Read Full License](#)

---

## TITLE PAGE

### Strategies to maximize the wood production in Amazon forest

Aline Canetti<sup>1</sup>, Evaldo Muñoz Braz<sup>2</sup>, Patrícia Póvoa de Mattos<sup>2\*</sup>, Renato Olivir Basso<sup>3</sup>,  
Afonso Figueiredo Filho<sup>4</sup>

(Canetti, A., Braz, E. M., Mattos, P. P., Basso, R. O., Figueiredo Filho, A.)

\*Corresponding author: [patricia.mattos@embrapa.br](mailto:patricia.mattos@embrapa.br), +55 41 3675-5625 fax: +55 41  
3675-5600

<sup>1</sup>Independent forestry consultant; E-mail: [alinecanetti@gmail.com](mailto:alinecanetti@gmail.com)

<sup>2</sup>Embrapa Florestas. Estrada da Ribeira Km 111, Caixa Postal 319. 83411-000,  
Colombo, PR, Brazil; E-mail: [evaldo.braz@embrapa.br](mailto:evaldo.braz@embrapa.br); [patricia.mattos@embrapa.br](mailto:patricia.mattos@embrapa.br)

<sup>3</sup>Elabore Projetos e Consultoria Florestal. Av. Gov. Júlio Campos, 207 - St. Comercial.  
78550-000, Sinop, Brazil; E-mail: [elabore@terra.com.br](mailto:elabore@terra.com.br)

<sup>4</sup>Universidade Estadual do Centro-Oeste. PR 153, Km 7, s/n - Riozinho, Irati, PR,  
Brazil. 84500-000. E-mail: [afigfilho@gmail.com](mailto:afigfilho@gmail.com)

#### To the Editor-in-Chief:

Dear Sir,

We would like to submit the manuscript “Strategies to maximize the wood production in Amazon forest” to possible publication in the Forest Ecosystems Journal, we strongly believe our theme fits within its scope.

We confirm that the manuscript was not published or it is not under evaluation elsewhere in substantially the same or abbreviated form, either in print or electronically. The innovation of this paper was the use of the analysis of the original structure of the forest combined with growth models to predict the optimal wood production in Amazon forest. It may be used by forest managers and forest law-makers, aiming to maximize sustainable wood production in the Amazon Forest.

Forest Ecology and Management was the most cited journal in our paper and

Sincerely,

Patrícia Póvoa de Mattos

# 1 Strategies to maximize the wood production in Amazon forest

## 3 ABSTRACT

4 **Background:** This study aimed to develop a procedure to determine which logging  
5 diameter would achieve optimal wood production by species, aiming to support  
6 sustainable management of the Amazon forest. Two main methodologies of analysis by  
7 species were combined: probability density function (PDF) and growth modeling. The  
8 growth models were used to derive the volume increment curves at the individual tree  
9 level. To detect the points of maximum annual increment in volume at the population  
10 tree level we used PDF with adjusted growth equations.

11 **Results:** The population maximum annual volumetric increments occurred in smaller  
12 diameters compared to that of the individual-level. When combining shorter cutting  
13 cycles with the population biological rotation point considered as the minimum cutting  
14 diameter (MCD), we observed higher annual increments in volume than that achieved  
15 using the Brazilian law criteria (MCD = 50 cm) or other MCD tested.

16 **Conclusion:** The procedure proposed may be used by forest managers and forest law-  
17 makers, aiming to maximize sustainable wood production in the Amazon forest.

18 **Keywords:** dendrochronology; forest modeling; minimum cutting diameter

## 20 Background

21 Natural forests are steadily accumulating biomass, reflected in the production of carbon  
22 and wood. The forest functions must be maintained after logging, and the remaining  
23 stocks must allow a continuous production of wood for the next interventions, ensuring  
24 forest management sustainability (Ong; Kleine, 1996; Van Gardingen et al., 2006; Braz  
25 et al., 2015; Avila et al., 2017). The management must be carefully planned and be

26 consistent with the initial natural forest structure (e.g., density, diametric distribution,  
27 growth, mortality, and regeneration) of the species of interest (Seydack et al., 1995;  
28 Bick et al., 1998).

29 According to Seydack (2012), a challenge for forestry scientists is to combine modeling  
30 and simulation tools to understand tropical forest dynamics, enabling production  
31 maximization. Minimum cutting diameters (MCD) should be defined for all commercial  
32 species, representing a powerful tool in maximizing increment for tropical forests under  
33 management as in the Amazon region.

34 Knowledge of the natural forest balance allows identification of the ideal moment of  
35 intervention and its production potential (Bick et al., 1998). The forest balance equation  
36 depends on how tree increment occurs over time. Studies of species growth at  
37 individual-level, structure, and reproductive biology are the basis for defining the  
38 management guidelines that guarantee sustainability for wood production in tropical  
39 forests (Vanclay, 1989; Miranda, D. L. C. et al., 2018; Dionisio et al., 2018).

40 Forest growth goes through youth, maturity, senescence, and stagnation phases. Its basic  
41 unit is the tree, which has a distinct sigmoidal growth pattern, tending to a biological  
42 limit (Odum, 1988). Trees that were established decades ago will grow within the  
43 ingress criteria for classes with larger diameters, will stagnate or eventually die. Thus, a  
44 diametric distribution of natural forest is the result of several tree establishments over  
45 different periods.

46 Diametric distribution is a useful factor to describe the forest properties since the  
47 diameter is easily obtained and correlated with other variables such as volume, which  
48 defines the economic value of the forest area (Bailey; Dell, 1973). Diameter is widely  
49 used in the forest sector to assess the effect of environmental and anthropogenic  
50 disturbances (Kohyama, 1986; Coomes et al., 2003; Wright et al., 2003; Bettinger et al.,

51 2009; Hossain et al., 2015), to describe successional patterns (Kohyama, 1986; Wright  
52 et al., 2003; Wang et al., 2009) and for the prediction of the future stock of a stand (De  
53 Liocourt, 1898; Meyer, 1952; Carvalho, 1981; Condit et al. , 1998; Bettinger et al.,  
54 2009; Hossain et al., 2015; Orellana; Figueiredo Filho, 2017). However, future stock  
55 projection is difficult for tropical forest due to its great diversity (Orellana, Figueiredo  
56 Filho, 2017). Moreover, information about individual species distribution pattern is  
57 scarce.

58 Describing forest growth trajectory requires long-term information when permanent  
59 plots are used as data source (Bick et al., 1998). The successive measurements can make  
60 it difficult or even imply a failed simulation, due to the time spent to replicate a  
61 structure that represents the initial diametric classes up to the biological diameter limit  
62 of the species (Brienen; Zuidema, 2006a; Miranda D. L. C. et al., 2018).

63 Most studies of the species forest growth in the Amazon basin were carried out with  
64 permanent plots (Silva, 1989; 1997; Higuchi, 1996; Fortini et al., 2015; Avila et al.,  
65 2017; D'Oliveira et al., 2017) and supported the management guidelines in the Brazilian  
66 legislation (Silva, 1989, 1997; Higuchi, 1996). However, in some species less than one  
67 individual per hectare occurs in Amazonian primary forest structure, especially when  
68 considering trees with commercial dimensions (Miranda, D. L. C. et al., 2018), and  
69 extensive sample areas are therefore required to properly represent a population of  
70 individual species (Groenendijk et al., 2017; Miranda, D. L. C. et al., 2018.).

71 In addition to area size, time scale is also a limiting factor in studying forest dynamics  
72 with permanent plots (Brienen; Zuidema, 2006a; Miranda, D. L. C. et al., 2018). In  
73 Brazil, permanent plots monitored for more than 40 years are rare. These periods are  
74 short compared to the age of commercial trees, which are rarely less than 100 years old  
75 (Brienen; Zuidema, 2006a; Schöngart, 2008; Rosa et al., 2017).

76 Growth series obtained by dendrochronology became more popular as an alternative to  
77 permanent-plot data to supply the demand for information and growth models of species  
78 (Brienen, 2005; Brienen; Zuidema, 2006a; Schöngart, 2011; Mattos et al., 2015; Canetti  
79 et al., 2017; Groenendijk et al., 2017; Miranda, D. L. C. et al., 2018). Growth ring  
80 analysis is a fast and reliable tool for assessing tree age, determining its growth rates  
81 over the life cycle, and identifying growth differences between species (Groenendijk et  
82 al., 2017; Rosa et al., 2017).

83 This study aimed to develop a procedure based on the diametric structure and growth  
84 models to set the minimum cutting diameter (MCD) for optimal wood production. The  
85 procedure was tested on four species in the Brazilian Amazon. The results may be used  
86 as a basis for forest management plans and to aid in the revision and modernization of  
87 forest law in countries with humid tropical forests.

88

## 89 **Methods**

90 This study was conducted in the Sinop micro-region (11°50' S; 54°50' W), Mato Grosso  
91 (MT) state, within the Brazilian Legal Amazon. The region comprises 9 municipalities,  
92 encompassing nearly 40 thousand km<sup>2</sup> (IBGE, 2010). The region's vegetation is  
93 characterized by a transition zone between open tropical rainforest and deciduous  
94 tropical rainforest, and is the largest area composed of this ecotone in the world (IBGE  
95 1992; SIVAM 2002). The micro-region presents flat relief and tropical climate with a  
96 dry season between autumn and winter (Alvares et al., 2013). All compartments from  
97 which data were collected were included in sustainable forest management plans  
98 authorized by the State Environmental Agency (SEMA-MT).

99 The species *Apuleia leiocarpa* (Vogel) J.F. Macbr., *Erisma uncinatum* Warm.,  
100 *Hymenolobium excelsum* Ducke, and *Trattinnickia burserifolia* Mart. were studied.

101 Together, these species represent 25% of the commercial volume logged in the micro-  
102 region of Sinop, the second-largest tropical wood production hub in Brazil (Ribeiro et  
103 al., 2016).

#### 104 **Fitting probability density function of the diametric structure of species**

105 The data were obtained from 100% inventories of six forest compartments carried out in  
106 the municipalities of Santa Carmem, Sinop, and União do Sul, totaling 5,432 ha. Forest  
107 inventories were conducted from 2011 to 2014, by measuring all trees with diameter at  
108 1.30 m above ground level ( $\text{dbh} \geq 40$  cm ( $n = 3,617$  for *A. leiocarpa*, 13,655 for *E.*  
109 *uncinatum*, 1,814 for *H. excelsum*, and 16,162 for *T. burserifolia*).

110 Data from sample units of 0.25 ha (10 m x 250 m) were also used to complement the  
111 densities of the smaller diameter trees ( $20 \leq \text{dbh} \leq 40$  cm), not covered in the 100%  
112 inventories. These plots were measured in four compartments in the municipalities of  
113 Santa Carmem and Feliz Natal, MT. All trees with  $\text{dbh} \geq 20$  cm were measured.  
114 Altogether, data from 138 plots were used, adding up to 36 ha inventoried. On average,  
115 there were 35 trees  $\text{ha}^{-1}$  with  $20 \leq \text{dbh} < 40$  cm (5 of *A. leiocarpa*, 11 of *E. uncinatum*, 4  
116 of *H. excelsum*, and 121 of *T. burserifolia*).

117 The dbh data for each species were grouped into diametric classes of 10 cm-range to fit  
118 probability density functions (PDFs). Normal, Beta, Gamma, Log-Normal, Johnson's  
119 SB, and Weibull with three parameters (Scolforo, 1998) PDFs were fitted. The diametric  
120 structure of each species included the largest diameter class in which live individuals  
121 were found in the 100% inventories. To avoid total numbers of individuals less than one  
122 tree per ha per species (as occurred for *A. leiocarpa* and *H. excelsum*), the densities of  
123 individuals were considered on a 10-ha scale (i.e. number of trees  $10 \text{ ha}^{-1}$ ).

124 The goodness-of-fit tests of PDFs by species were performed in two steps, using the  
125 Kolmogorov-Smirnov test ( $\alpha = 0.05$ ). Initially, the goodness-of-fit of the theoretical

126 PDFs to the empirical data was tested in each forest compartment. In the second stage,  
127 the goodness-of-fit of the PDFs fitted to the data of the entire Sinop micro-region (i.e.,  
128 for the total data set) were tested considering sample inventories and 100% inventories.  
129 The best PDF for each species was selected based on the lowest maximum absolute  
130 value between theoretical PDF and observed data ( $D_{\text{calc}}$ . For Kolmogorov-Smirnov test)  
131 and lowest error index (Reynolds et al., 1988).

### 132 **Modeling diametric growth by species**

133 Stem sections (discs) with approximately 5-cm thick were collected from trees from  
134 different forests compartments in the Sinop micro-region. These samples were obtained  
135 from the base of the first commercial log, approximately 1 m above the ground level, or  
136 immediately above the sapopemas. We collected 9 discs samples from *A. leiocarpa*, 10  
137 from *E. uncinatum*, 13 from *H. excelsum*, and 11 from *T. burserifolia*. The samples were  
138 dried and polished and the growth rings were measured with an accuracy of 0.01 mm on  
139 a LINTAB measurement table (Frank Rinn, Heidelberg, Germany), using the TSAP-  
140 Win software (Rinn, 1996). Crossdating was carried out among radii of each tree and  
141 among trees.

142 The mean increment and passage time were analyzed by dbh class. We adjusted the  
143 growth models of Gompertz, Johnson-Schumacher, Lundqvist-Korf, Logistic,  
144 Monomolecular, and Schumacher for each species (Burkhart; Tomé, 2012).

145 The values of the asymptotes ( $\beta_0$ ) were fixed at the upper limit of the largest diametric  
146 class recorded in the 100% inventories used. To increase modeling accuracy, we used  
147 the non-parametric Bootstrap method with 100 interactions (Miller, 2004), previously  
148 used in studies of modeling growth in diameter (Brienen, 2005; Mattos et al., 2015;  
149 Canetti et al., 2017). The models were fitted by non-linear regression using the PROC  
150 NLMIXED procedure in the SAS<sup>®</sup> software. The most suitable growth model by



151 species was selected based on the relative residual standard error ( $S_{yx\%}$ ), corrected  
152 Akaike and Bayesian information criteria.

### 153 **Obtaining growth curves in individual volume for each species**

154 Initially, the commercial height at each diameter was obtained to estimate the  
155 commercial volume at each age and to relate it to the diametric growth curve. The  
156 commercial height/dbh ratio of 30 trees (randomly selected) per species was fitted by  
157 non-linear regression (PROC NLMIXED procedure in the SAS<sup>®</sup> software) according to  
158 equation 1 (Schöngart, 2008). Of the 30 trees, four came from sampling inventories (20  
159  $\leq$  dbh < 40 cm) and 26 from 100% inventories (dbh  $\geq$  40 cm). This selection method  
160 was adopted due to the difference of inclusion levels among inventories and the varying  
161 sizes of the measured areas.

162 To calculate the wood volume, we used commercial height and dbh for each age  
163 (obtained from the diametric growth equation), and taper functions developed for the  
164 municipality of Santa Carmem (MT) to calculate the wood volume. For *E. uncinatum*  
165 and *T. burserifolia*, we used species-level taper functions (equations 2 and 3,  
166 respectively) developed by Lansanova (2012). For *A. leiocarpa* and *H. excelsum* we  
167 used the general taper function for the municipality (equation 4) developed by  
168 Lansanova et al. (2018).

169 The di/dbh ratio of the growth samples was applied to calculate the youngest diameters  
170 at the dbh, from which the volume at each age was calculated. As a result, the volume  
171 growth curve of the individual tree was obtained for each species. From the volume  
172 growth curves, the mean annual volume increment (MAI<sub>v</sub>) and current annual volume  
173 increment (CAI<sub>v</sub>) curves for individual trees were obtained according to equations 5 and  
174 6.

$$H = \frac{dbh\beta_0}{dbh + \beta_1} \quad (1)$$

$$\frac{d_i}{dbh_{E.uncinatum}} = 1.20 + 2.97 \frac{h_i}{H} + 10.95 \left(\frac{h_i}{H}\right)^2 + 21.15 \left(\frac{h_i}{H}\right)^3 + 19.51 \left(\frac{h_i}{H}\right)^4 + 6.80 \left(\frac{h_i}{H}\right)^5 \quad (2)$$

$$\frac{d_i}{dbh_{T.burserifolia}} = 1.13 + 1.89 \frac{h_i}{H} + 7.16 \left(\frac{h_i}{H}\right)^2 + 14.02 \left(\frac{h_i}{H}\right)^3 + 12.56 \left(\frac{h_i}{H}\right)^4 + 4.15 \left(\frac{h_i}{H}\right)^5 \quad (3)$$

$$\frac{d_i}{dbh_{General}} = 1.21 + 2.94 \frac{h_i}{H} + 10.94 \left(\frac{h_i}{H}\right)^2 + 20.46 \left(\frac{h_i}{H}\right)^3 + 17.93 \left(\frac{h_i}{H}\right)^4 + 5.90 \left(\frac{h_i}{H}\right)^5 \quad (4)$$

$$MAI_v = \frac{v_t}{t} \quad (5)$$

$$CAI_v = v_{t+1} - v_t \quad (6)$$

175 where: dbh = diameter at 1.30 m above ground level (cm);  $d_i$  = diameter at the sample collection height  
 176 (disc sample, cm);  $h_i$  = height of sample collection (m);  $H$  = commercial height of the tree (m);  $\beta_0$  and  $\beta_1$   
 177 = model parameters fitted by non-linear regression;  $MAI_v$  = mean annual volume increment ( $m^3$ );  $CAI_v$  =  
 178 current annual volume increment ( $m^3$ );  $v_t$  = accumulated commercial volume ( $m^3$ ), at different ages;  $t$  =  
 179 age (years), according to the diametric growth curve.

180

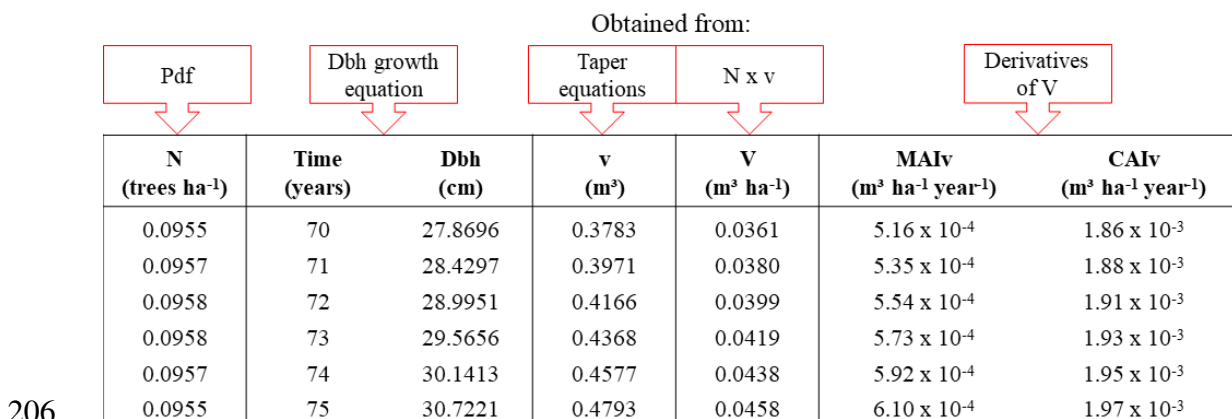
181 The age of biological rotation of individual tree was obtained at the intersection point of  
 182 the  $CAI_v$  and  $MAI_v$  curves. If the age of biological rotation was not reached with the  
 183 ages of the available samples, we extrapolated the growth equation in diameter to  
 184 estimate that age. For the extrapolated diameters, the commercial volume and  $CAI_v$  and  
 185  $MAI_v$  were calculated, following the same procedures.

### 186 **Obtaining growth curves in volume of the population of each species**

187 To obtain growth in volume per unit area for the population of each species over time,  
 188 individual growth curves in volume were associated with the PDFs. It was assumed that  
 189 a species diametric distribution follows the same pattern that occurred in the past  
 190 (Gotelli, 2008; Lundqvist, 2017). From this premise, the maximization of the population  
 191 volume increments was obtained. The population was defined hereby as the number of  
 192 trees from the first to the last diametric class in which there were individuals alive until  
 193 the moment “ $t$ ”.

194 The input of the PDFs is the dbh of the trees and the output is the estimated number of  
 195 trees per unit area. The growth equations developed to generate the diameter values  
 196 were used to estimate the number of trees alive at each time “t”, considering the species’  
 197 growth curve. The procedure illustrated in Fig. 1 allowed us to estimate the population’s  
 198 diametric evolution over time (time x number of trees ha<sup>-1</sup>). As the actual diametric  
 199 distributions were established with data from dbh ≥ 20 cm, the calculations were  
 200 performed from that diameter, according to the following procedure:

- 201 I. The number of trees at each time “t” (identified by the PDF) was multiplied by  
 202 the volume of the individual tree (v) at each time, resulting in the population  
 203 production curve (V) at each time referring to the species growth curve;
- 204 II. From the population volume production curve, MAI<sub>v</sub> and CAI<sub>v</sub> were derived  
 205 (equations 5 and 6).



207 Fig. 1. Scheme showing the methodological process for obtaining the curves of volume increment of the  
 208 population. The boxes with red outline inform the data source of the column to which the arrows are  
 209 directed. PDFs = probability density functions; N = number of trees per hectare; dbh = diameter at 1.30 m  
 210 above ground level; v = volume of individual tree; V = species population volume (m<sup>3</sup> ha<sup>-1</sup>); MAI<sub>v</sub> =  
 211 mean annual volume increment; and CAI<sub>v</sub> = current annual volume increment.

## 213 Management simulations

214 The projection method by diametric class (Alder, 1995) was used to compare the  
 215 volumetric increments obtained from the different combinations of cutting cycle and  
 216 minimum cutting diameter (MCD), according to the following management scenarios:

217 I. MCD = 50 cm and cutting cycle = 35 years, according to current legislation  
 218 (Brasil, 2006);

219 II. MCD defined by the biological rotation age of the population (intersection of the  
 220 MAI<sub>v</sub> and CAI<sub>v</sub> curves) and cutting cycle = 35 years;

221 III. MCD defined by the maximum CAI<sub>v</sub> and biological rotation age of the  
 222 population and cutting cycle calculated using equation 7 (Schöngart, 2008).

$$Cutting\ cycle(years) = \frac{Age_{Max.CAIv}}{0.1dbh_{Max.CAIv}} \quad (7)$$

223 where Cutting cycle (years) = time interval between explorations in the forest (Schöngart, 2008);  
 224 dbh<sub>Max.CAIv</sub> = minimum cutting diameter, defined as the diameter at which the species reaches the  
 225 maximum CAI<sub>v</sub>; Age<sub>Max.CAIv</sub> = age at which the species reaches the MCD.

226

227 IV. MCD defined by dbh in which the maximum CAI<sub>v</sub> occurs at the individual tree  
 228 level and cutting cycle calculated according to Schöngart (2008);

229 V. MCD defined by the dbh of biological rotation age of the population in volume  
 230 and increasing cutting cycles (10 to 70 years).

231

232 The mathematical procedure for simulations in the projection method by diametric class  
 233 is presented in equation 8 (Alder, 1995).

$$N_{k,t+1} = N_{k,t} + I_k - O_k - M_k - H_k \quad (8)$$

234 where N<sub>k,t+1</sub> = number of trees in dbh class *k* in period *t+1*; N<sub>k,t</sub> = number of trees in class *k* in period *t*; I<sub>k</sub>  
 235 = ingrowth in class *k* during the period; O<sub>k</sub> = number of trees passing from class *k* to subsequent classes,  
 236 M<sub>k</sub> = mortality in class *k*; H<sub>k</sub> = trees logged during the period.

237

238 The matrices were built from the 30-cm dbh class, for every 5 years (t), until reaching  
 239 the desired cutting cycle. We considered the wood volume of trees with a dbh higher  
 240 than the MCD the volume to be harvested ( $H_k$ ). The following data were used:

- 241 I. Initial diametric structure ( $N_{k,t}$ ): obtained from PDFs fitted by species,  
 242 considering  $\text{dbh} \geq 30$  cm;
- 243 II. Mean increment by diameter class (obtained from the growth equations). The  
 244 number of trees passing to the subsequent classes ( $O_k$ ) was calculated from the  
 245 passage time between diametric classes (equation 9), as described by Alder  
 246 (1995).

247

$$O_k = \frac{t \cdot i}{w} \quad (9)$$

248 where:  $O_k$  = number of trees migrating from class  $k$  to subsequent classes during the period considered;  $i$   
 249 = mean increment ( $\text{cm year}^{-1}$ ) of the diametric class  $k$  (obtained from the growth series);  $w$  = interval  
 250 between diametric classes (10 cm in the present study);  $t$  = period considered.

251

- 252 III. Ingrowth ( $I_k$ ): the number of trees in the 20-cm diameter class (smallest dbh  
 253 class obtained) was considered as ingrowth.
- 254 IV. Mortality ( $M_i$ ): the mortality was estimated using equation 10 adapted from  
 255 Lundqvist (2017).

$$M_i = 100 \left( 1 - \frac{N_{i+1} i_{i+1} \frac{2}{t_i + t_{i+1}}}{N_i i_i} \right) \quad (10)$$

256 where:  $M_i$  = percentage of annual mortality in diametric class  $i$ ;  $N_i$  = number of trees in class  $i$ ;  $N_{i+1}$  =  
 257 number of trees in the class subsequent to  $i$ ;  $i_i$  and  $i_{i+1}$  = mean annual increment in diameter (cm) of  
 258 classes  $i$  and  $i+1$  (obtained from the growth equation);  $t_i$  and  $t_{i+1}$  = passage time (years) from the diametric  
 259 classes  $i$  and  $i+1$  to the subsequent classes (obtained from the growth equation).

260

261 After obtaining the final diametric structure of the projection matrix by diametric class,  
 262 the number of trees was converted into volume (applying the taper functions previously  
 263 described) to obtain the total wood production in the considered period. This production  
 264 was divided by the cutting cycles, in order to proportionally compare the production of  
 265 the tested MCD and cutting cycles.

266

## 267 **Results**

### 268 **Diametric structure**

269 When testing the adherence of general probability density functions (PDFs) adjusted to  
 270 the measured density distributions for each species and for each forest compartment, the  
 271 majority of tested PDF fitted (Table 1). The only exceptions found were the Normal  
 272 function (all areas) and the Beta function (areas 3, 4 and 6) for *T. burserifolia*.

273

274 Table 1. Goodness-of-fit test of the probability density function (PDF) fitted considering the total data set  
 275 to the observed diametric distributions in each study area by the Kolmogorov-Smirnov test.

Species	D	PDF	Area 1	Area 2	Area 3	Area 4	Area 5	Area 6
<i>Apuleia leiocarpa</i>	D <sub>calc.</sub>	Normal	0.048	0.100	0.078	0.139	0.089	0.063
		Beta	0.047	0.094	0.081	0.141	0.092	0.062
		Gamma	0.054	0.100	0.070	0.154	0.077	0.054
		Log Normal	0.075	0.109	0.080	0.167	0.060	0.048
		Johnson's SB	0.045	0.095	0.075	0.147	0.085	0.060
		Weibull 3p	0.044	0.095	0.076	0.145	0.087	0.061
	D <sub>tab.</sub>	0.410	0.410	0.521	0.565	0.432	0.375	
<i>Erisma uncinatum</i>	D <sub>calc.</sub>	Normal	0.049	0.256	0.084	0.055	0.082	0.044
		Beta	0.072	0.190	0.151	0.107	0.026	0.050
		Gamma	0.069	0.261	0.081	0.045	0.099	0.060
		Log Normal	0.095	0.278	0.063	0.068	0.122	0.085

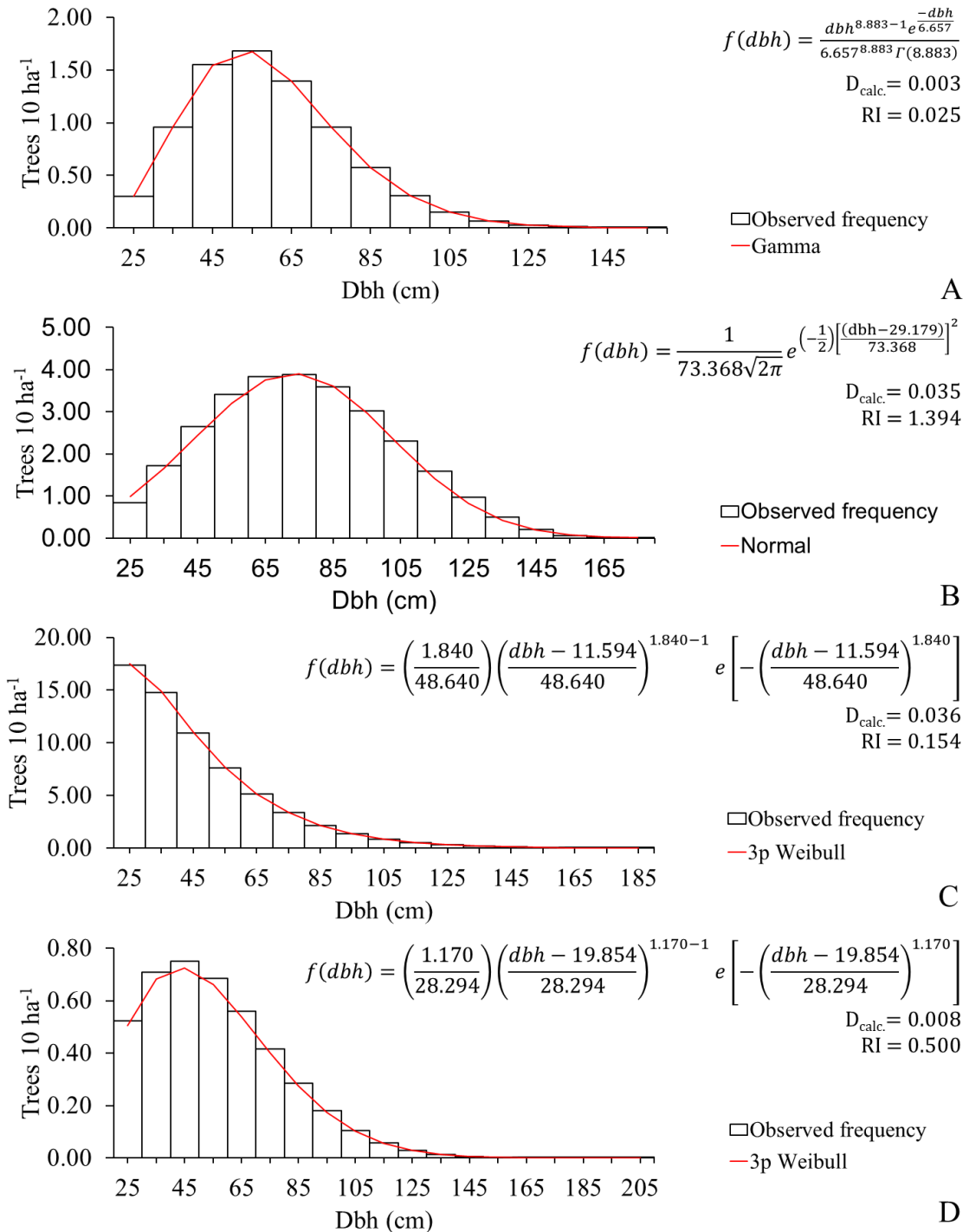
		Johnson's SB	0.047	0.257	0.083	0.056	0.082	0.043
		Weibull 3p	0.049	0.253	0.088	0.061	0.078	0.038
		D <sub>tab.</sub>	0.189	0.294	0.240	0.240	0.270	0.270
<i>Hymenolobium excelsum</i>	D <sub>calc.</sub>	Normal	0.124	0.124	0.124	0.124	0.151	0.124
		Beta	0.061	0.060	0.060	0.060	0.131	0.092
		Gamma	0.049	0.061	0.062	0.056	0.132	0.093
		Log Normal	0.051	0.068	0.065	0.068	0.132	0.093
		Johnson's SB	0.049	0.058	0.063	0.050	0.129	0.090
		Weibull 3p	0.038	0.051	0.067	0.054	0.125	0.086
		D <sub>tab.</sub>	0.565	0.624	0.521	0.521	0.708	0.708
<i>Tratinnickia burserifolia</i>	D <sub>calc.</sub>	Normal	0.238*	0.238*	0.238*	0.238*	0.238*	0.238*
		Beta	0.187	0.170	0.170*	0.170*	0.170	0.170*
		Gamma	0.180	0.149	0.149	0.149	0.149	0.149
		Log Normal	0.167	0.119	0.106	0.106	0.106	0.106
		Johnson's SB	0.127	0.071	0.071	0.071	0.071	0.092
		Weibull 3p	0.149	0.080	0.046	0.042	0.054	0.077
		D <sub>tab.</sub>	0.193	0.183	0.166	0.170	0.178	0.169

276 Area = sampled forest compartment; Dcalc. = maximum absolute value between fitted PDF and observed  
277 values in each compartment; Dtab. = critical value of Kolmogorov-Smirnov test ( $\alpha = 0,05$ ); \* = the PDF  
278 did not fit to the observed values ( $D_{tab.} \leq D_{calc.}$ ).

279

280 When considering the adherence of all PDFs tested to the total set of observed data by  
281 species all functions fitted. The Gamma function fitted better to the observed data for *A.*  
282 *leiocarpa*, the Normal function for *E. uncinatum* and *H. excelsum*, and the Weibull with  
283 three parameters for *T. burserifolia* (Fig. 2). A unimodal pattern was observed for the  
284 diametric distributions of *A. leiocarpa*, *E. uncinatum*, and *H. excelsum*. On the other  
285 hand, *T. burserifolia* decreased in a regular number of trees, similar to the negative

286 exponential distribution. The decreased tendency starts in *H. excelsum* at 45-cm, in *A.*  
 287 *leiocarpa* at 55-cm, and in *E. uncinatum* at 75-cm diameter.  
 288



289  
 290 Fig. 2. Probability density functions with better fitting for *Apuleia leiocarpa* (A), *Erisma uncinatum* (B),  
 291 *Hymenolobium excelsum* (C) and *Trattinnickia burserifolia* (D). Dbh = diameter at 1.30 m above ground



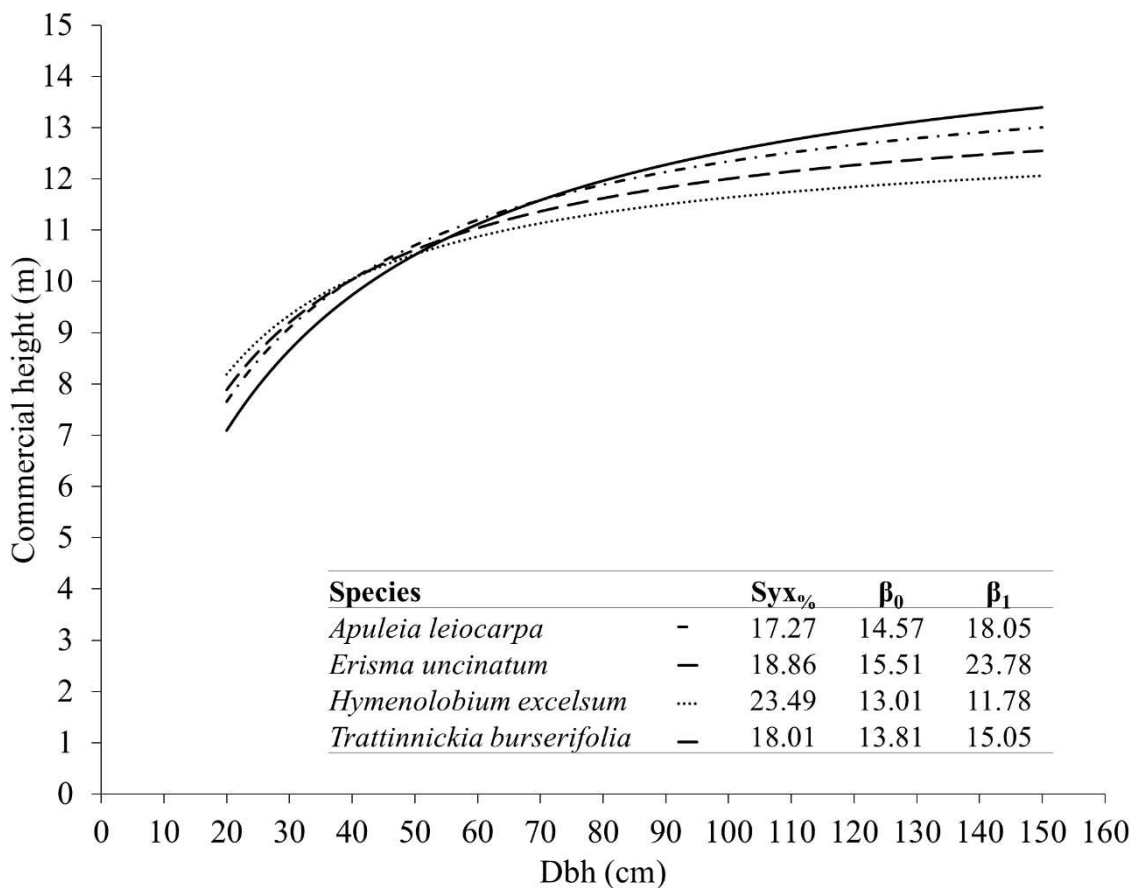
292 level (cm);  $\Gamma$  = gamma function; Dcalc. = maximum absolute value between fitted pdf and observed  
 293 values in each compartment; IR = error index (Reynolds et al., 1988).

294

295 **Height/dbh ratio**

296 The fitted height/dbh model showed satisfactory statistics and residual distribution for  
 297 all species (Fig. 3). The species that reached the highest commercial heights were *E.*  
 298 *uncinatum* and *H. excelsum*. The largest commercial height range occurred before  
 299 reaching 60-cm dbh, ranging from 6.5 to 10.9 m. After 60-cm dbh, the commercial  
 300 heights tended to stabilize, ranging from 10.9 to 13.4 m.

301



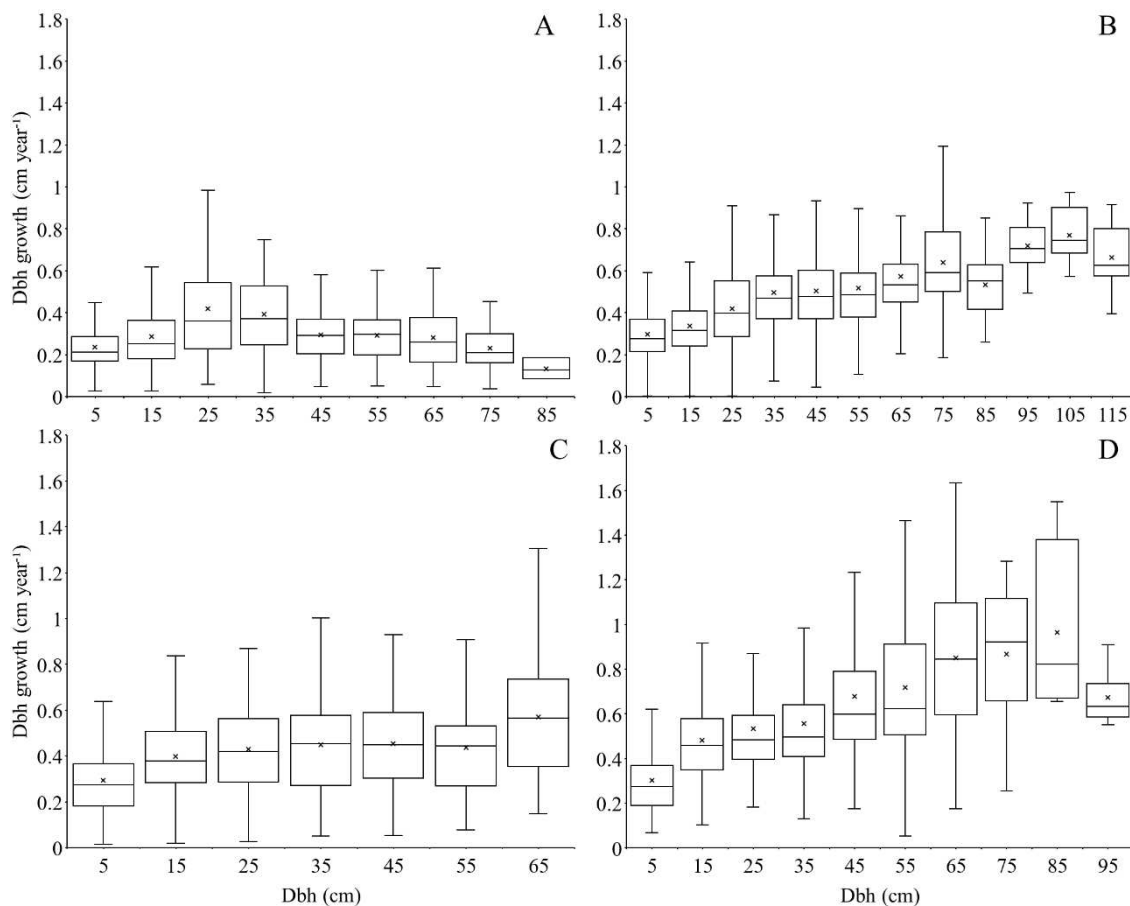
302

303 Fig. 3. Heigh/dbh models of *Apuleia leiocarpa*, *Erisma uncinatum*, *Hymenolobium excelsum*, and  
 304 *Trattinnickia burserifolia*. dbh = diameter at 1.30 m above ground level (cm);  $\beta_0$  and  $\beta_1$  = equation  
 305 parameters fitted by non-linear regression;  $S_{yx\%}$  = relative residual standard error (%)

306

### 307 **Diametric increment in dbh classes**

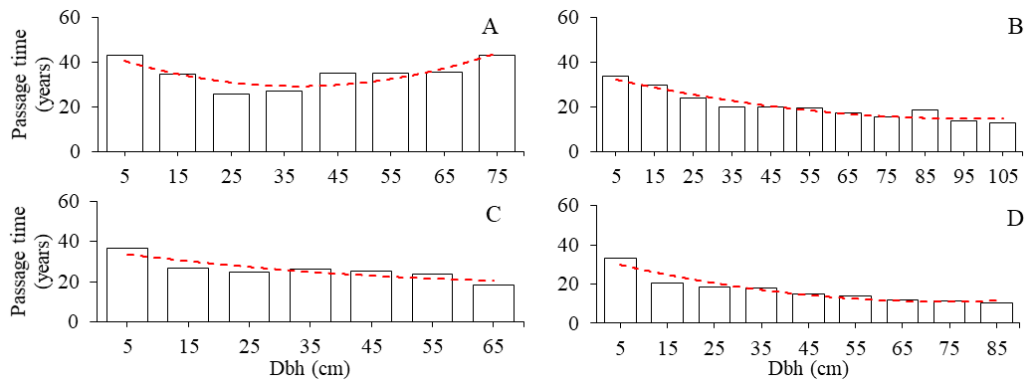
308 The mean increment by dbh class and passage time are shown in Fig. 4 and 5. *Apuleia*  
309 *leiocarpa* presented an initial increase in dbh increment, and a decrease for the upper  
310 diameter classes (Fig. 4A), characterizing the U shape for the passage time (Fig. 5A).  
311 *Erisma uncinatum*, *H. excelsum*, and *T. burserifolia* showed a higher increase in the  
312 upper classes (Fig. 4 B, C and D), assuming a decreasing passage time (Fig. 5 B, C and  
313 D).



314

315 Fig. 4. Boxplot of the mean annual periodic increment by diameter class of *Apuleia leiocarpa* (A), *Erisma*  
316 *uncinatum* (B), *Hymenolobium excelsum* (C), and *Trattinnickia burserifolia* (D). Markers (x) represent the  
317 mean increment per diameter class, which were calculated only for the diameter classes with three or  
318 more trees. Dbh = diameter at 1.30 m above ground level.

319



320

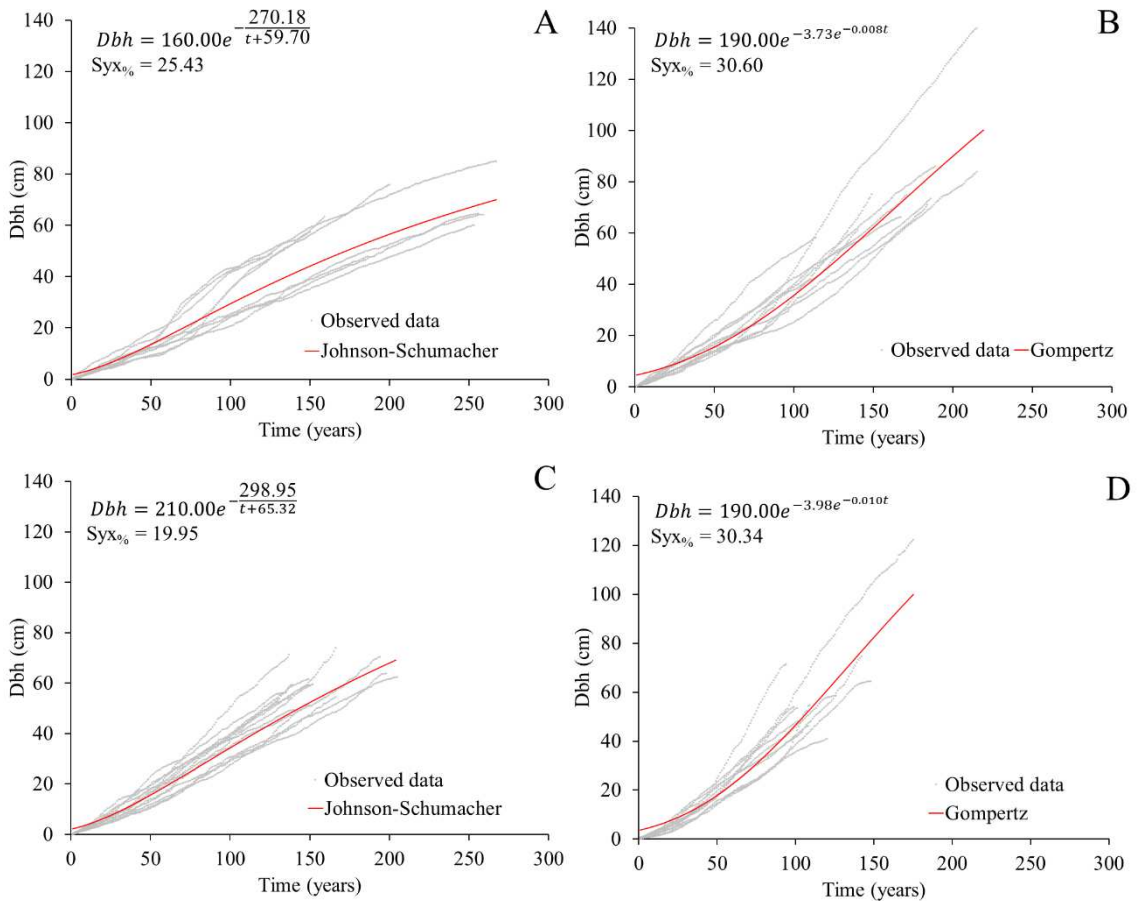
321 Fig. 5. Mean passage time by diameter class for *Apuleia leiocarpa* (A), *Erismia uncinatum* (B),  
 322 *Hymenolobium excelsum* (C), and *Trattinnickia burserifolia* (D) and their respective tendency lines. Only  
 323 passage times for diameter classes with three or more trees were considered. Dbh = diameter at 1.30 m  
 324 above ground level.

325

### 326 **Modeling accumulated diameter growth and generating increment curves**

327 The Johnson-Schumacher model better fitted the time series of *A. leiocarpa* and *H.*  
 328 *excelsum*, while the Gompertz was better for *E. uncinatum* and *T. burserifolia*. For  
 329 diameter growth, *A. leiocarpa* and *H. excelsum* reached the point of maximum tangency  
 330 close to 35 cm and 50 cm, respectively, registering lower increments after these  
 331 diameter classes (Fig. 6 A and C). However, *E. uncinatum* and *T. burserifolia* did not  
 332 reach this point within the measured diameters (Fig. 6 B and D), showing upward  
 333 growth throughout the entire time series.

334



335

336 Fig. 6. Accumulated diametric growth equations fitted for *Apuleia leiocarpa* (A), *Erisma uncinatum* (B),  
 337 *Hymenolobium excelsum* (C), and *Trattinnickia burserifolia* (D). The growth equations were fitted for the  
 338 ages represented by three or more samples. Dbh = diameter at 1.30 m above ground level; t = time  
 339 (years);  $S_{yx\%}$  = Relative Residual Standard Error (%).

340

341 *Apuleia leiocarpa*, *E. uncinatum*, and *T. burserifolia* reached the maximum current  
 342 annual increment in volume (CAI<sub>v</sub>) when considering the available data (Fig. 7 A, B  
 343 and D). The maximum CAI<sub>v</sub> for *H. excelsum* was estimated at approximately 245 years  
 344 (Fig. 7C).

345 For the four species, the growth equation was used to estimate the biological rotation  
 346 age, which for individual trees occurred at the 95-cm dbh for *A. leiocarpa*, 121-cm for  
 347 *H. excelsum*, and 165-cm for *E. uncinatum* and *T. burserifolia*. The maximum CAI<sub>v</sub>

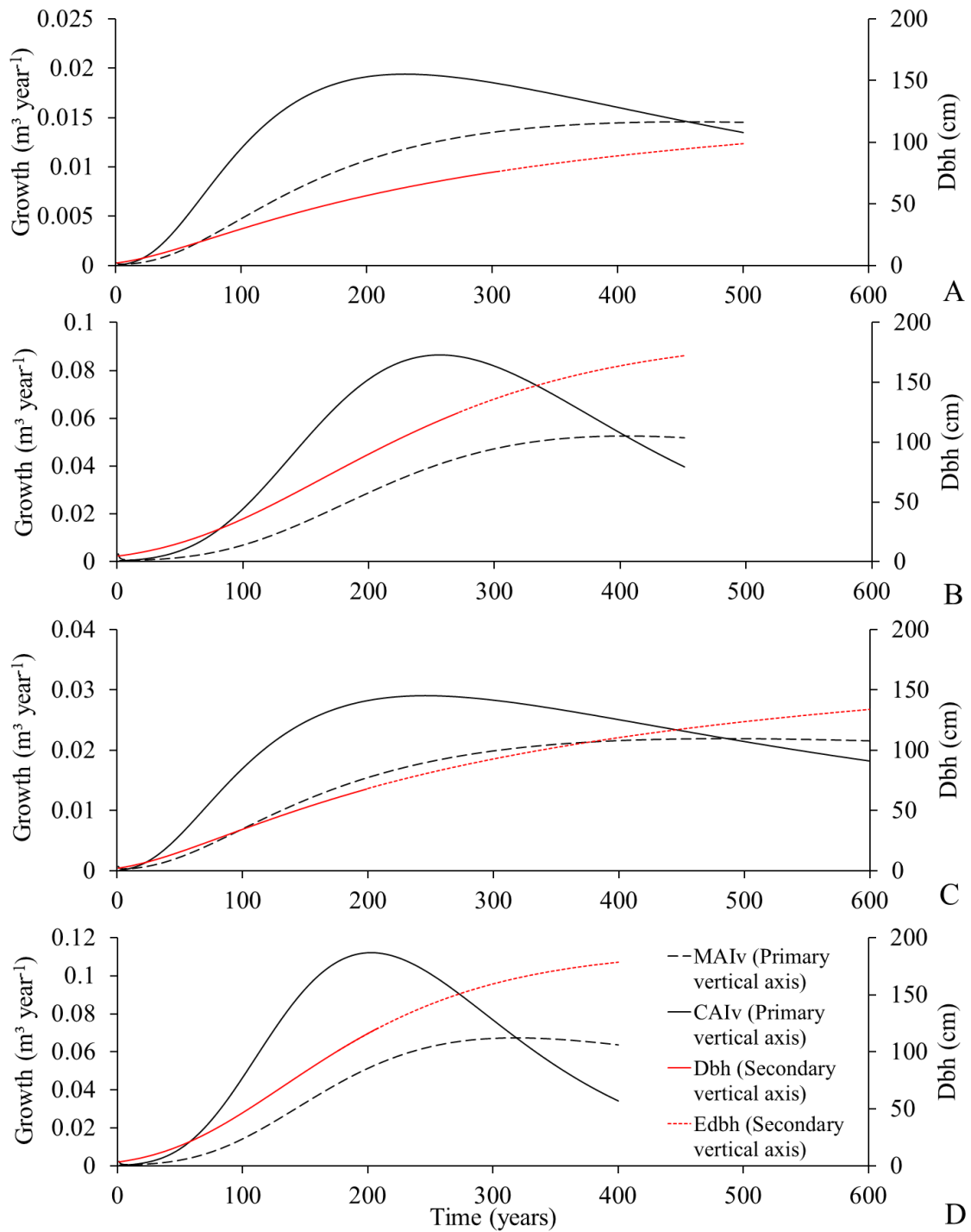
348 occurred close to the 60-cm dbh for *A. leiocarpa*, 80-cm for *H. excelsum*, and 120-cm  
 349 for *E. uncinatum* and *T. burserifolia* (Table 2).

350

351 Table 2. Estimated ages and diameters on the biological rotation and the maximum current annual  
 352 increment of individual tree volume for the species.

Species	Biological rotation		Maximum CAI <sub>v</sub>	
	Dbh (cm)	Age (years)	Dbh (cm)	Age (years)
<i>Apuleia leiocarpa</i>	95	455	63	231
<i>Erismia uncinatum</i>	165	409	119	257
<i>Hymenolobium excelsum</i>	121	480	80	245
<i>Trattinnickia burserifolia</i>	165	319	118	203

353 Dbh = diameter at 1.30 m above ground level; maximum CAI<sub>v</sub> = maximum current annual increment for  
 354 individual tree volume; Biological rotation = intersection of the CAI<sub>v</sub> and MAI<sub>v</sub> (current and mean annual  
 355 increment for individual tree volume) curves.



356

357 Fig. 7. Volumetric increment curves (black) and dbh growth equation (red) for *Apuleia leiocarpa* (A),  
 358 *Erismia uncinatum* (B), *Hymenolobium excelsum* (C), and *Trattinnickia burserifolia* (D). In the primary y-  
 359 axis:  $MAI_v$  = mean annual volumetric increment;  $CAI_v$  = current annual volumetric increment. On the  
 360 secondary y-axis: dbh = diameter at 1.30 m above ground level (cm) (growth equation fitted within the  
 361 measured data range); Edbh = accumulated diameter at 1.30 m above ground level (growth equation  
 362 outside the measured data range).

363 **Growth by species population**

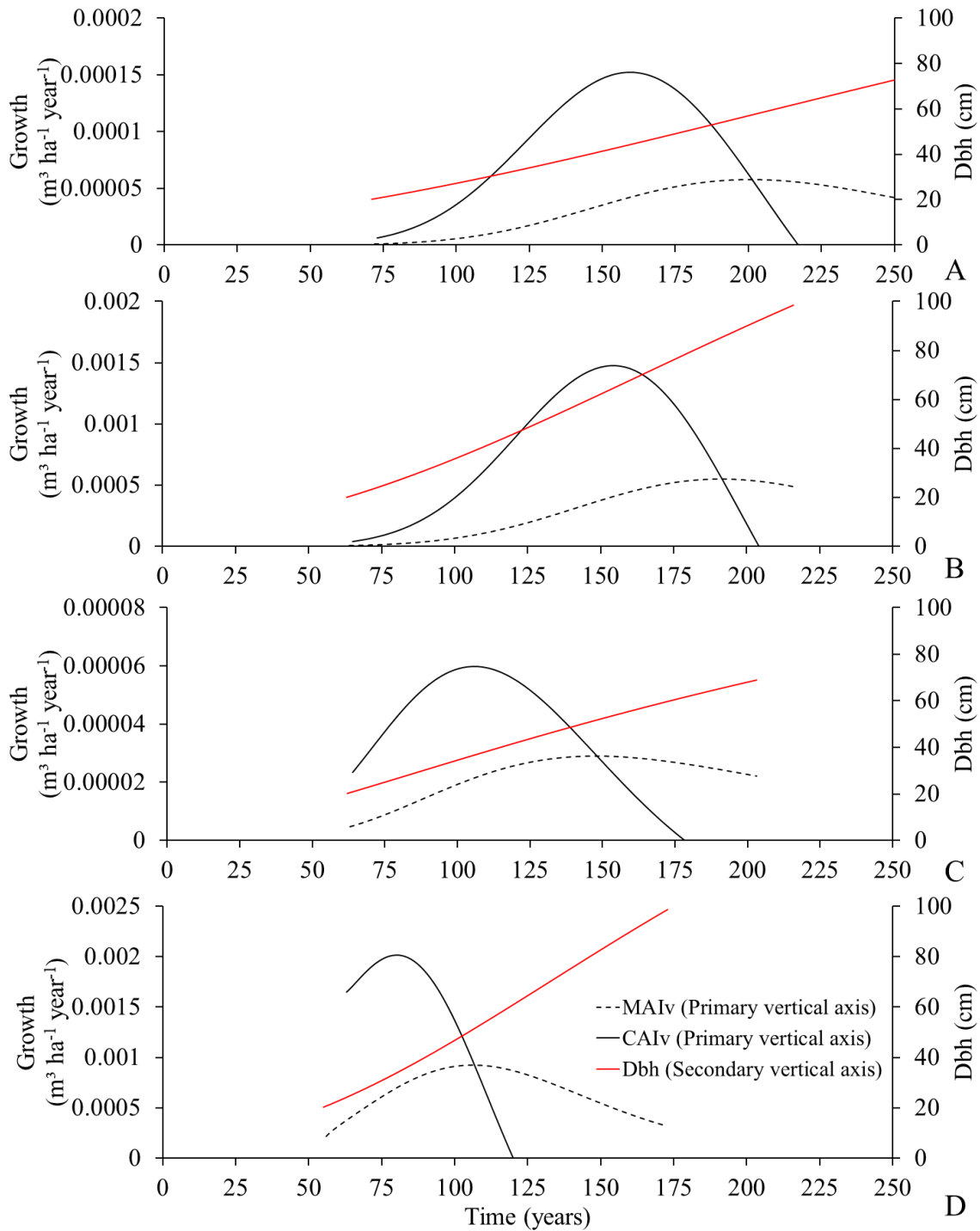
364 The maximum  $CAI_v$  and biological rotation ages of the population volume (Table 3;  
 365 Fig. 8) occurred in lower diameters than those recorded for individual trees (Fig. 7;  
 366 Table 2). Those ages were reached for diameters within the stem samples. There was  
 367 therefore no need for the extrapolation of the volume growth equations that was  
 368 required for the individual diameter growth (Fig. 7).

369

370 Table 3. Estimated ages and diameters on the biological rotation and the maximum current annual  
 371 increment by species population.

Species	Biological rotation		Max. $CAI_v$	
	Dbh (cm)	Age (years)	Dbh (cm)	Age (years)
<i>Apuleia leiocarpa</i>	58	203	44	160
<i>Erismia uncinatum</i>	86	192	64	154
<i>Hymenolobium excelsum</i>	52	150	37	106
<i>Trattinnickia burserifolia</i>	58	116	34	80

372 Dbh = diameter at 1.30 m above ground level; Max.  $CAI_v$  = maximum current annual increment for  
 373 population volume; biological rotation = intersection of the curves of  $CAI_v$  and mean annual increment in  
 374 volume for individual tree.



375

376 Fig. 8. Volumetric and diametric increment curves for the population of *Apuleia leiocarpa* (A), *Erisma*  
 377 *uncinatum* (B), *Hymenolobium excelsum* (C) and *Trattinnickia burserifolia* (D) (dbh  $\geq$  20 cm). In the  
 378 primary y-axis: MAI<sub>v</sub> = mean annual increment in volume; CAI<sub>v</sub> = current annual increment in volume.  
 379 In the secondary y-axis: dbh = diameter at 1.30 m above ground level (cm), obtained from the growth  
 380 equation.

381



382 **Minimum cutting diameter (MCD) and cutting cycle simulations**

383 The MCD criteria determined by the biological rotation age of the population and with  
 384 short cutting cycles (10 years) (Table 4) resulted in a higher annual volume yield for the  
 385 four species. For *A. leiocarpa*, *H. excelsum* and *T. burserifolia*, the maximum CAI<sub>v</sub> age  
 386 of the population and the cutting cycle, as suggested by Schöngart (2008), resulted in  
 387 the least increment (m<sup>3</sup> ha<sup>-1</sup> year<sup>-1</sup>) of all criteria. For *E. uncinatum* the management  
 388 criteria defined in the legislation resulted in the least increment.

389

390 Table 4. Simulation results using different criteria for defining the minimum cutting diameter and cutting  
 391 cycle.

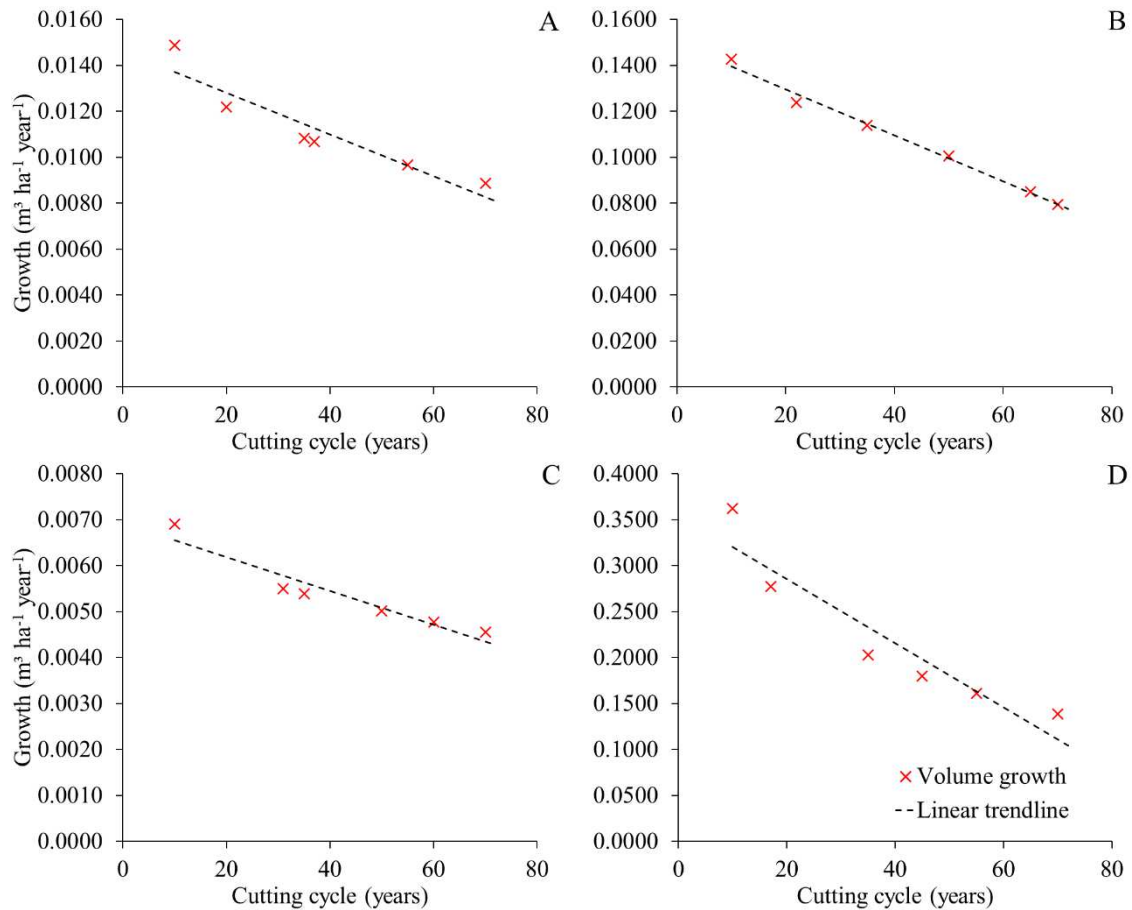
Species	MCD	Law	Rot. pop.	Rot. pop.	CAI <sub>v</sub> pop.	CAI <sub>v</sub> ind.	Rot. pop.	Rot. pop.
	Cutting cycle	Law	Law	SM	SM	SM	Short	Long
<i>Apuleia leiocarpa</i>	MCD (cm)	50	57	57	44	63	57	57
	Cycle (years)	35	35	37	37	37	10	70
	Inc. (m <sup>3</sup> ha <sup>-1</sup> year <sup>-1</sup> )	0.0078	0.0107	0.0106	0.0055	0.0113	<b>0.0159</b>	0.0086
	Vol. (m <sup>3</sup> 100 ha <sup>-1</sup> )	27.46	37.52	39.09	20.49	41.82	15.86	60.53
<i>Erisma uncinatum</i>	MCD (cm)	50	86	86	64	119	86	86
	Cycle (years)	35	35	22	22	22	10	70
	Inc. (m <sup>3</sup> ha <sup>-1</sup> year <sup>-1</sup> )	0.0238	0.1140	0.1238	0.0643	0.0800	<b>0.1427</b>	0.0795
	Vol. (m <sup>3</sup> 100 ha <sup>-1</sup> )	83.27	398.90	272.28	141.51	176.01	142.74	556.66
<i>Hymenolobium excelsum</i>	MCD (cm)	50	52	52	37	80	52	52
	Cycle (years)	35	35	31	31	31	10	70
	Inc. (m <sup>3</sup> ha <sup>-1</sup> year <sup>-1</sup> )	0.0051	0.0054	0.0055	0.0029	0.0049	<b>0.0069</b>	0.0046

	Vol. (m <sup>3</sup> 100 ha <sup>-1</sup> )	17.69	18.84	17.03	8.86	15.31	6.90	31.86
<i>Trattinnickia burserifolia</i>	MCD (cm)	50	58	58	34	118	58	58
	Cycle (years)	35	35	17	17	17	10	70
	Inc. (m <sup>3</sup> ha <sup>-1</sup> year <sup>-1</sup> )	0.1511	0.2027	0.2776	0.1216	0.1966	<b>0.3622</b>	0.1384
	Vol. (m <sup>3</sup> 100 ha <sup>-1</sup> )	528.95	709.37	471.85	206.64	334.20	362.25	968.89

392 MCD = minimum cutting diameter; Cycle = cutting cycle; SM = Schöngart (2008) method; Rot.<sub>pop.</sub> =  
393 biological rotation age of the population; CAI<sub>v pop.</sub> = maximum current annual increment of the  
394 population; Legisl. = Normative Instruction parameters (Brasil, 2006); Inc. = simulated mean annual  
395 volumetric increment; Vol. (m<sup>3</sup> 100 ha<sup>-1</sup>) = total volume produced in a 100-ha area, during the cutting  
396 cycle length. The highest simulated annual increments in volume are presented in bold.

397

398 Considering the MCD defined by the population's biological rotation age, the annual  
399 volumetric production decreased for the four species as the cutting cycle increased (Fig.  
400 9).



401

402 Fig. 9. Annual volume increment by the MCD defined by the population's biological rotation age and  
 403 different cutting cycles for *Apuleia leiocarpa* (A), *Erismia uncinatum* (B), *Hymenolobium excelsum* (C),  
 404 and *Trattinnickia burserifolia* (D).

405

## 406 Discussion

### 407 Diametric distribution pattern of the species

408 Most of the probability density functions (PDFs) (Table 1) fitted to data in each forest  
 409 compartment. Thus, a single PDF per species was fitted at the Sinop microregion (Fig.  
 410 2). Fitting only one PDF per species across the microregion suggests the existence of  
 411 similarities in the species diametric structure in the typology, contrasting to the claim  
 412 that ecosystems are chaotic and disordered (O'Hara, 2014). According to this same  
 413 author, ecosystems are not deterministic or ordered, due to the occurrence of repeated  
 414 disturbances on small scales, which prevent the forest from being homogeneous over

415 time. However, for Larson (1992), the occurrence of environmental or anthropic  
416 disturbances combined with the genetic variation of mixed forest does not imply that  
417 forest development patterns cannot be discerned.

418 Also, the micro-region's pattern agrees with the demographic equilibrium theory, which  
419 states that the diameter distribution balance may be scalar (Muller-Landau et al., 2006).

420 The forest structure is regulated by general principles of growth, mortality (Wang et al.,  
421 2009) and ecological succession (Oliver, 1992).

422 Even in distant or geographically isolated regions, species with resembling  
423 physiological and morphological characteristics can have a similar population structure  
424 (Oliver, 1992). This occurs because the structures repeat over time in the forest (Gotelli,  
425 2008). Diametric structure models allow us to infer the diameter in which tree number  
426 reduces as a result of stagnant growth and increased mortality (Durrieu de Madron;  
427 Forni, 1997; Braz et al., 2014).

428 Ecologists use the diametric structure to indicate the health of the forest. A high number  
429 of smaller trees, as recorded for *T. burserifolia* (Fig. 2D), would represent forest  
430 sustainability, since it would restore the upper classes (Kohyama, 1986; Muller-Landau  
431 et al., 2006; Wang et al., 2009). These considerations originated from De Liocourt's  
432 (1898) and Meyer's (1952) theories that state the concept of a balanced forest following  
433 a negative exponential pattern.

434 However, several studies have found that some forest types and many individual species  
435 may not follow such a negative exponential model (Condit et al., 1998; Dawkins;  
436 Philip, 1998; Nyland, 2002; Pascal, 2003; Braz, 2010; Braz et al., 2014; Hossain et al.,  
437 2015), as occurred for three of the species in this study (Fig. 2 A, B and C).

438 A minor number of smaller trees, as recorded for *A. leiocarpa*, *E. uncinatum* and *H.*  
439 *excelsum*, could indicate that the populations are declining (Condit, 1998; Hossain et al.,

440 2015). However, the forest diametric structure is not the only indicator defining  
441 population maintenance. Schaaf et al. (2006) pointed out that the highest density does  
442 not guarantee the species maintenance in the community, but their ability to compete  
443 within the ecological niche does. The authors stated that if the species has few trees with  
444 smaller diameters, but if the individuals tolerate competition for light in the lower  
445 canopy, they can remain in the forest. Therefore, it is necessary to consider other  
446 factors, particularly the species growth in the diameter classes (Condit et al., 1998).

447 Braz et al. (2014) identified over 60% of the basal area in the upper stratum in the Sinop  
448 micro-region, a situation described by Pascal (2003). This author stated that the deficit  
449 of trees in some dbh classes or the accumulation in larger dbh classes can result from  
450 different causes. For example, a massive fruiting and germination due to favorable  
451 climatic conditions can engender an overstock of trees in a particular dbh class. These  
452 individuals can outgrow to larger dbh classes after a certain time, resulting in an  
453 imbalanced diametric structure curve. These same individuals may be stagnated in a  
454 certain dbh class when under the closed canopy of the forest.

455 An over-stocked upper stratum may have its regeneration constrained due to inadequate  
456 access to light by trees in the lower canopy (Yegang; Jinxuan, 1988; Lamprecht, 1990;  
457 Felfili, 1997; Nyland, 2002; Bettinger et al. 2009). In such cases, some species may  
458 need a longer time scale to regenerate (Felfili, 1997). Over time, this factor de-  
459 characterizes both the negative exponential structure (Braz, 2010; Bettinger et al., 2009)  
460 and the whole forest increment (Dawkins; Philip, 1998).

#### 461 **Production modeling at individual tree level**

462 Commercial height is an essential dendrometric variable for modeling production  
463 forests since it is used for volume prediction (Lappi, 1997). Fitting parameters for the  
464 commercial height/dbh ratio (Fig. 3) allows us to calculate commercial volumes of

465 individual trees and stands more accurately where there is no field measuring. Other  
466 authors used the commercial height/dbh ratio to model the wood volume in the Amazon  
467 Forest (Brienen; Zuidema, 2006a; Fortini; Zarin, 2011; Miranda, Z. P. et al., 2018).  
468 According to Alder (1995), in natural forests the initial growth of trees is slow, followed  
469 by a phase of higher increments until it reaches its maximum. After that maximum  
470 point, growth reduces until stagnation. *Apuleia leiocarpa*, *E. uncinatum* and *T.*  
471 *burserifolia* followed this pattern (Fig. 4 A, B and D and 5 A, B and D), while *H.*  
472 *excelsum* maintained constant growth during five diameter classes, not exceeding 0.5  
473 cm year<sup>-1</sup>, except in the 65-cm class (Fig. 4C). The constant growth of *H. excelsum* until  
474 larger sizes is a feature of late and emerging secondary species (Silva et al., 1985),  
475 whose crowns extend above the canopy average level. These functional groups have  
476 several micro-habitats during their ontogenesis (Clark; Clark, 1992). They are  
477 characterized as partial sciophytes, shade-tolerant in the early development stages.  
478 However, they require a high degree of illumination to overcome the intermediate stages  
479 to maturity, increasing its growth during canopy opening (Maciel et al., 2017).  
480 The decreasing tendency of passage time (Fig. 5 B, C, and D) for the late secondary  
481 species (*E. uncinatum*, *H. excelsum*, and *T. burserifolia*) is also a feature of partial  
482 sciophytes, that are shade tolerant, but do not demand shade for growth (Maciel et al.,  
483 2017).  
484 The U-shaped passage time of the initial secondary species (*A. leiocarpa*) has already  
485 been pointed out by Brienen and Zuidema (2006b) for *Bertholletia excelsa* Bonpl., an  
486 initial secondary species (Gouveia et al., 2011). The same tendency was also registered  
487 for *Goupia glabra* Aubl. in the micro-region of Sinop-MT (Oliveira et al., 2015), which  
488 shows maximum growth in the 25-cm and 35-cm diameter classes, despite being a late  
489 secondary species (Araújo et al., 2009), as occurred for *A. leiocarpa* (Fig. 5A).

490 The biological rotation and maximum  $CAI_v$  ages (Fig. 7; Table 2) were consistent with  
491 the increment by diameter class (Fig. 4). *Apuleia leiocarpa*, an initial secondary species,  
492 reached biological rotation and maximum  $CAI_v$  ages in smaller diameters, while *H.*  
493 *excelsum* reached it in intermediate sizes. The other two species achieved biological  
494 rotation for an estimated dbh higher than 160 cm. The maximum  $CAI_v$  of an individual  
495 tree were used in scientific studies to define minimum cutting diameters and cutting  
496 cycles for species in Amazonas State (Schöngart, 2008; Rosa et al., 2017; Miranda D. L.  
497 C., 2018), and has already been included in State forest law (Amazonas, 2010). If  
498 logging happens before the maximum  $CAI_v$  or after the maximum  $MAI_v$ , the volumetric  
499 production will be inefficient. In this case, the species would not have reached or would  
500 have passed their optimal growth (Schöngart, 2008; Braz; Mattos, 2015). Therefore,  
501 knowing the maximum  $CAI_v$  and  $MAI_v$  per species is essential for sustainable forest  
502 management in the Amazon forest.

503 The curves of individual wood volume increment indicate the ideal parameters for  
504 sustainable management of each species. However, other variables representing the  
505 species dynamics must be considered for decision-making. For example, it is essential  
506 to define available stock of trees by diameter class and survival rates (Sebben et al.,  
507 2008; Free et al., 2014; Groenendijk et al., 2017; Free, 2017).

### 508 **Volumetric population growth**

509 The maximum  $CAI_v$  estimated from the population increment curves occurred for  
510 diameters lower than 50 cm for *A. leiocarpa*, *H. excelsum*, and *T. burserifolia*, and at  
511 64-cm diameter for *E. uncinatum* (Table 3; Fig. 8). The dbh when biological rotation  
512 age occurs, at which  $CAI_v$  and  $MAI_v$  curves intersected, was around 50 cm for *A.*  
513 *leiocarpa*, *H. excelsum*, and *T. burserifolia*, while for *E. uncinatum* this occurred at 86  
514 cm. *Erisma uncinatum* has increased growth (Fig. 4) and a high concentration of trees in

515 the upper diameter classes, which explains its yield culmination in larger dbh sizes. This  
516 feature suggests the maintenance of larger dbh trees in the forest. In polycyclic  
517 management systems, the trees available for logging must reach at least the diameters  
518 with the highest growth potential in volume (Miranda, Z. P. et al., 2018) represented by  
519 the MCD. For this reason, it is essential to determine the diameter that provides the  
520 highest wood production for each species, based on the individual growth and the  
521 population demographic characteristics.

522 According to Miranda, Z. P. et al. (2018), the species management in the mixed forest  
523 should be carried out within the diameters of the higher  $CAI_v$  and the biological rotation  
524 of the population to maximize the volume. However, we observed a large range between  
525 the maximum  $CAI_v$  and the biological rotation ages, of at least 36 years and 14-cm dbh  
526 (Table 3; Fig. 8). It remains unclear which MCD should be applied the highest wood  
527 yield for the next cutting cycle.

528 The  $CAI_v$  curves and biological rotation ages for all species population culminated in  
529 lower diameters and ages compared to the increment curves at the individual level (Fig.  
530 7; Table 2). The inclusion of the number of trees per diameter class anticipates the  
531 culmination since it indirectly represents the population mortality and survival rates  
532 (Assmann, 1970; Rubin et al., 2006). As trees increase in size, the number of trees  
533 decreases due to mortality, causing a deceleration in the gross wood production, even  
534 though the remaining trees become larger (Seydack, 2000; Lundqvist, 2017). Therefore,  
535 even if individual trees are growing, the total production per unit area will decline much  
536 earlier (Assmann, 1970).

537 According to Seydack (2000) the population demographic factors of commercial species  
538 are essential to determine the optimal parameters for forest management. In the natural  
539 forest structure, the available stock of trees by diameter class is a crucial factor for yield



540 projection in terms of the number of trees, basal area and volume (Ong; Kleine, 1996),  
541 since they indicate survival by diameter class. The recovery time of the wood volume in  
542 a post-intervention cycle is related to the species-specific growth, and also to the  
543 number of trees in the smaller diameter classes to the intervention diameter (Brienen;  
544 Zuidema, 2006a). Therefore, methods based on the inflection point age at the population  
545 level are recognized for being more accurate in predicting maximum volume yield in a  
546 natural forest than methods that consider only the individual tree growth.

#### 547 **Volumetric production using different management criteria**

548 The increments ( $\text{m}^3 \text{ha}^{-1} \text{year}^{-1}$ ) of each species were, in general, compatible with other  
549 studies carried out in the Amazon, using permanent plots (Reis et al., 2010; Braz et al.,  
550 2018). Reis et al. (2010) obtained similar increments for *E. uncinatum* (Table 4),  
551 considering the cutting cycle and MCD of the Brazilian forest law (Brasil, 2006).

552 The cutting cycles and alternative MCD from those defined in Brazilian forest law  
553 produced higher volumetric increment for the 4 studied species, similar to that obtained  
554 by Groenendijk et al. (2017) in tropical rainforest in Cameroon. *Erismia uncinatum*  
555 produced approximately seven times higher wood volume ( $\text{ha}^{-1} \text{year}^{-1}$ ) applying short  
556 cutting cycle (10 years) and MCD 70% larger (86 cm) than the Brazilian mandatory  
557 MCD (50 cm). *Erismia uncinatum* presented the highest yield gain with the alternative  
558 MCD and cutting cycles, due to the higher number of trees and growth in dbh classes  
559 above 50 cm (Fig. 2 and 7). All results, especially those of *E. uncinatum*, showed the  
560 potential of maximizing the forest management yield considering population  
561 characteristics at the species level. The ideal MCD can increase volumetric increments,  
562 even without silvicultural treatments, and so should be considered to increase  
563 management productivity (Avila et al., 2017).

564 The MCD defined by the dbh of the biological rotation of the population produced  
565 higher annual increments for the four species, especially for shorter cutting cycles  
566 (Table 4). It is important to consider the optimal population increment rate instead of  
567 average growth data. This was previously considered by Ong and Kleine (1996) and  
568 Bick et al. (1998) when studying natural forests in Malaysia. These authors obtained the  
569 population growth from zero using permanent plots and simulations of forest production  
570 and found the maximum sustainable yield population, defining optimum logging rates  
571 and cutting cycles. Glauner et al. (2003) pointed out that the underlying principle of  
572 management is to improve the forest condition, converting its stock to an optimum level  
573 of increment of the commercial species.

574 In this study the dbh of the maximum  $CAI_v$  (lowest among those tested as MCD)  
575 produced the lowest volume increments considering the tested cycles (Table 4).  
576 According to Bick et al. (1998), the harvested volume of two cycles must oscillate in the  
577 population growth curve between a lower and an upper limit of the maximum yield  
578 volume (i.e., the age of maximum  $CAI_v$ ). These authors recommended an MCD above  
579 the dbh of maximum  $CAI_v$ . Our results are consistent with these recommendations,  
580 since the yield was higher when the logging was carried out at an age after the  
581 maximum population  $CAI_v$ , such as the biological rotation age.

582 When considering only the individual tree growth curve to define the logging  
583 parameters we observed that the MCD was always higher, although the volumetric  
584 increments were lower, except for *A. leiocarpa* (Table 4). Contrary to the statement of  
585 Sebbenn et al. (2008) and Lacerda et al. (2013), the excessive MCD increase does not  
586 always imply higher forest productivity for all species. As mentioned, after reaching  
587 their productivity peak the trees grow at moderate rates and the mortality is high. Slow

588 growth trees should be logged before dying, through selective logging of mature trees of  
589 declining vigor (Seydack, 2012; Lundqvist, 2017).

590 The cutting cycles must be long enough so that remaining trees reach the MCD  
591 (Seydack et al., 1995). However, it cannot be too long, to ensure the trees' potential  
592 growth and forest economic sustainability (Mattioli et al., 2015). The simulations  
593 revealed that long cutting cycles resulted in smaller volumetric increments for all  
594 species (Fig. 9), reaching a loss higher than 60% compared to the 10-year cutting cycle,  
595 considering the same MCD (defined by population biological rotation). Therefore,  
596 increasing the cutting cycle indefinitely (Sebbenn et al., 2008) represents a technical  
597 error in a productive forest (Braz et al., 2015). Moreover, increasing the cut cycle  
598 requires long-term investment (Glauner et al., 2003), which is difficult in unstable  
599 political-economic environments (Groenendijk et al., 2017). On the other hand, short  
600 cutting cycles require logging planning, following the reduced-impact logging  
601 standards.

602 Suitable cutting cycles depend on economic considerations (Bick et al., 1998) besides  
603 the MCD and the forest remaining structure. Economic analysis should support  
604 selecting the best cutting cycle for each specific case, considering the optimal  
605 parameters and exploitation feasibility for each species. Cost-benefit analyses can help  
606 assess forest management's economic and social aspects (Rosa et al., 2017). Moreover,  
607 higher MCD requires studies on tree health and risks of rot in older and bigger trees.

608

## 609 **Conclusions**

610 The diametric distribution of each species is similar throughout the transition zone  
611 between open tropical rainforest and deciduous tropical rainforest in Brazilian Amazon.  
612 Modifications of the standard diametric structure by species can be combined with

613 growth models for inferring tropical forest dynamics, as done in this study. The  
614 diametric distribution should support wood production planning and be the basis for  
615 monitoring forest logging.

616 The optimum population minimum cutting diameter (MCD) of a species can  
617 significantly increase the volumetric increment and the total wood volume, maximizing  
618 the forest production under management. After determining the MCD, the best cutting  
619 cycle must be defined according to economic analysis, which will specify the best  
620 intervention time for the productive forest.

621 The methodology for species-level MCDs proved to be a useful tool and superior to the  
622 fixed parameters defined in the Brazilian law or the methodologies that consider only  
623 the individual tree growth. The procedure provides higher reliability for the maximized  
624 wood production. The ideal MCDs based on dendrochronology substantially reduces  
625 costs and time spent collecting data for forest modeling.

626

## 627 **Availability of data and materials**

628 The data that support the findings of this study are available from Embrapa Florestas  
629 and Elabore Projetos e Consultoria Florestal but restrictions apply to the availability of  
630 these data, which were used under license for the current study, and so are not publicly  
631 available. Data are however available from the authors upon reasonable request and  
632 with permission of Embrapa Florestas and Elabore Projetos e Consultoria Florestal.

633

## 634 **Abbreviations**

635 **CAIv:** current annual volume increment

636 **DBH:** diameter at 1.30 m above ground level

637 **MAIv:** mean annual volume increment

638 **MCD:** minimum cutting diameter

639 **PDF:** probability density function

640

## 641 **Acknowledgements**

642 We are grateful to Elabore Projetos e Consultoria Florestal for providing inventory data  
643 and discs samples and Simon Drew for the language editing.

644

## 645 **Funding**

646 This work was supported by Cipem (Center of Wood Producing and Exporting  
647 Industries of Mato Grosso State), CAPES (Coordination for the Improvement of Higher  
648 Education Personnel) and Embrapa Florestas (Brazilian Agricultural Research  
649 Corporation).

650

## 651 **Authors' contributions**

652 AC participated of elaborating the paper conceptualization and methodology, made the  
653 formal analysis and wrote the original draft of the paper.

654 EMB participated of elaborating the paper conceptualization and methodology,  
655 collected field data and supervised the paper writing and design.

656 PPM participated of elaborating the paper conceptualization and methodology and  
657 supervised the paper writing and design.

658 ROB participated of elaborating the paper conceptualization and collected field data.

659 AFF participated of elaborating the paper conceptualization and methodology and  
660 supervised the paper writing and design.

661

## 662 **Ethics approval and consent to participate**

663 Not applicable

664

665 **Consent for publication**

666 Not applicable

667

668 **Competing interests**

669 The authors declare that they have no competing interests.

670

671 **References**

672 Alder D (1995) Growth Modelling for Mixed Tropical Forest, 30th edn. ODA Forestry

673 Research programme project R4676, Oxford

674 Alvares CA, Stape JL, Sentelhas PC, et al (2013) Köppen's climate classification map

675 for Brazil. Meteorol Zeitschrift 22:711–728. <https://doi.org/10.1127/0941->

676 2948/2013/0507

677 Amazonas (2010) Instrução Normativa no 9, de 12 de novembro de 2010. Dispõe sobre

678 manejo florestal sustentável em áreas de várzea no Estado do Amazonas, e dá

679 outras providências. Diário Oficial do Estado do Amazonas, Brazil

680 Araujo R de A, Costa RB da, Felfili JM, et al (2009) Florística e estrutura de fragmento

681 florestal em área de transição na Amazônia Matogrossense no município de Sinop.

682 Acta Amaz 39:865–877. <https://doi.org/10.1590/S0044-59672009000400015>

683 Assmann E (1970) The principles of forest yield study. Pergamon Press Ltd.

684 Avila AL, Schwartz G, Ruschel AR, et al (2017) Recruitment, growth and recovery of  
685 commercial tree species over 30 years following logging and thinning in a tropical  
686 rain forest. For Ecol Manage 385:225–235.  
687 <https://doi.org/10.1016/j.foreco.2016.11.039>

688 Bailey RL, Dell R (1973) Quantifying Diameter Distributions with the Weibull  
689 Function. For Sci 19:97–104

690 Bettinger P, Boston K, Siry J, Grebner D (2009) Forest Management and Planning.  
691 Academic Press, New York

692 Bick U, Droste H-J, Glauner R, Heuveldop JO (1998) Assessment and measurement of  
693 forestry key parameters for the evaluation of tropical forest management. Plant Res  
694 Dev 47/48:38–61

695 Brasil (2006) Instrução normativa n. 6 de 11 de dezembro de 2006. Diário Oficial da  
696 União, Brasília, DF, 13 dez. 2006. Seção 1, p. 155

697 Braz EM, Mattos PP (2015) Manejo De Produção Em Florestas Naturais Da Amazônia:  
698 Mitos E Verdades. Nativa 3:292–295. [https://doi.org/10.14583/2318-](https://doi.org/10.14583/2318-7670.v03n04a12)  
699 [7670.v03n04a12](https://doi.org/10.14583/2318-7670.v03n04a12)

700 Braz EM (2010) Subsídios para o planejamento do manejo de florestas tropicais da  
701 amazônia. Universidade Federal de Santa Maria

702 Braz EM, Canetti A, Mattos PP, et al (2018) Alternative criteria to achieve sustainable  
703 management of *Mezilaurus itauba* in the Brazilian Amazon. Pesqui Florest Bras  
704 38:1–8. <https://doi.org/10.4336/2018.pfb.38e201801648>

705 Braz EM, Mattos PP de, Oliveira MF, Basso RO (2014) Strategies for Achieving  
706 Sustainable Logging Rate in the Brazilian Amazon Forest. *Open J For* 04:100–105.  
707 <https://doi.org/10.4236/ojf.2014.42015>

708 Braz EM, Mattos PP, Thaines F, et al (2015) Criteria to be considered to achieve a  
709 sustainable second cycle in Amazon Forest. *Pesqui Florest Bras* 35:209.  
710 <https://doi.org/10.4336/2015.pfb.35.83.941>

711 Brienens RJW (2005) Tree rings in the tropics: a study on growth and ages of Bolivian  
712 rain forest trees

713 Brienens RJW, Zuidema PA (2006) The use of tree rings in tropical forest management :  
714 Projecting timber yields of four Bolivian tree species. *For Ecol Manage* 226:256–  
715 267. <https://doi.org/10.1016/j.foreco.2006.01.038>

716 Brienens RJW, Zuidema PA (2006) Lifetime growth patterns and ages of Bolivian rain  
717 forest trees obtained by tree ring analysis. *J Ecol* 94:481–493.  
718 <https://doi.org/10.1111/j.1365-2745.2005.01080.x>

719 Burkhardt HE, Tomé M (2012) *Modeling Forest Trees and Stands*. Springer Netherlands,  
720 Dordrecht

721 Canetti A, de Mattos PP, Braz EM, Netto SP (2017) Life pattern of urban trees: A  
722 growth-modelling approach. *Urban Ecosyst* 20:1057–1068

723 Carvalho PER (1981) Competição entre espécies florestais nativas em Irati - PR, cinco  
724 anos após o plantio. *Bol Pesqui Florest* 41–56



725 Clark D a., Clark DB (1992) Life history diversity of canopy and emergent trees in a  
726 neotropical rain forestClark, D. A. ., & .C. and D. B. (1992). Life history diversity  
727 of canopy and emergent trees in a neotropical rain forest. *Ecological monographs*,  
728 62(3), 315-344. Retrieved from. *Ecol Monogr* 62:315–344.  
729 <https://doi.org/10.2307/2937114>

730 Condit R, Sukumar R, Hubbell SSP, et al (1998) Predicting Population Trends from  
731 Size Distributions : A Direct Test in a Tropical Tree Community. *Am Nat*  
732 152:495–509. <https://doi.org/10.1086/286186>

733 Coomes DA, Duncan RP, Allen RB, Truscott J (2003) Disturbances prevent stem size-  
734 density distributions in natural forests from following scaling relationships. *Ecol*  
735 *Lett* 6:980–989. <https://doi.org/10.1046/j.1461-0248.2003.00520.x>

736 Dawkins HC, Philip MS (1998) *Tropical Moist Forest Silviculture and Management: A*  
737 *History of Success and Failure*. CAB International, Wallingford, UK

738 de Liocourt F (1898) De l’amenagement des sapinières. *Bull Trimest* 396–409

739 Dionisio LFS, Schwartz G, Lopes J do C, Oliveira F de A (2018) Growth, mortality,  
740 and recruitment of tree species in an Amazonian rainforest over 13 years of  
741 reduced impact logging. *For Ecol Manage* 430:150–156.  
742 <https://doi.org/10.1016/j.foreco.2018.08.024>

743 D’Oliveira MVN, Oliveira LC, Acuña MHA, Braz EM (2017) Twenty years monitoring  
744 growth dynamics of a logged tropical forest in Western Amazon. *Pesqui Florest*  
745 *Bras* 37:493–502. <https://doi.org/10.4336/2017.pfb.37.92.1398>

746 Felfili JM (1997) Diameter and height distributions in a gallery forest tree community  
747 and some of its main species in central Brazil over a six-year period (1985-1991).  
748 Rev Bras Botânica 20:155–162. [https://doi.org/10.1590/S0100-](https://doi.org/10.1590/S0100-84041997000200006)  
749 84041997000200006

750 Fortini LB, Zarin DJ (2011) Population dynamics and management of Amazon tidal  
751 floodplain forests: Links to the past, present and future. For Ecol Manage 261:551–  
752 561. <https://doi.org/10.1016/j.foreco.2010.11.007>

753 Fortini LB, Cropper WP, Zarin DJ (2015) Modeling the complex impacts of timber  
754 harvests to find optimal management regimes for Amazon tidal floodplain forests.  
755 PLoS One 10:1–18. <https://doi.org/10.1371/journal.pone.0136740>

756 Free CM, Grogan J, Schulze MD, et al (2017) Current Brazilian forest management  
757 guidelines are unsustainable for *Swietenia*, *Cedrela*, *Amburana*, and *Copaifera*: A  
758 response to da Cunha and colleagues. For Ecol Manage 386:81–83.  
759 <https://doi.org/10.1016/j.foreco.2016.09.031>

760 Free CM, Matthew Landis R, Grogan J, et al (2014) Management implications of long-  
761 term tree growth and mortality rates: A modeling study of big-leaf mahogany  
762 (*Swietenia macrophylla*) in the Brazilian Amazon. For Ecol Manage 330:46–54.  
763 <https://doi.org/10.1016/j.foreco.2014.05.057>

764 Glauner R, Ditzer T, Huth A (2003) Growth and yield of tropical moist forest for forest  
765 planning: an inquiry through modeling. Can J For Res 33:521–535.  
766 <https://doi.org/10.1136/bmj.2.2867.854>

- 767 Gotelli NJ (2008) *A Primer of Ecology*. Sinauer Associates Inc., Sunderland,  
768 Massachusetts
- 769 Gouveia D, Soares M, Silva W, et al (2011) Avaliação do crescimento de espécies  
770 florestais por grupo ecológico em áreas exploradas na FLONA do Tapajós. III  
771 Encontro Amaz Agrárias 1–5
- 772 Groenendijk P, Bongers F, Zuidema PA (2017) Using tree-ring data to improve timber-  
773 yield projections for African wet tropical forest tree species. *For Ecol Manage*  
774 400:396–407. <https://doi.org/10.1016/j.foreco.2017.05.054>
- 775 Higuchi N (1996) Utilização e manejo dos recursos madeireiros das florestas tropicais  
776 úmidas. *Acta Amaz* 24:275–288
- 777 Hossain MA, Hossain MK, Alam MS, Uddin MM (2015) Composition and Diversity of  
778 Tree Species in Kamalachari Natural Forest of Chittagong South Forest Division,  
779 Bangladesh. *J For Environ Sci* 31:192–201.  
780 <https://doi.org/10.7747/JFES.2015.31.3.192>
- 781 IBGE - Instituto Brasileiro de Geografia e Estatística (1992) Mapa de vegetação do  
782 Brasil -Escala 1:5.000.000. [http://www.dpi.inpe.br/Ambdata/mapa\\_sipam.php](http://www.dpi.inpe.br/Ambdata/mapa_sipam.php).  
783 Accessed 10 Mar 2017
- 784 IBGE - Instituto Brasileiro de Geografia e Estatística (2010) Censo Demográfico.  
785 <https://censo2010.ibge.gov.br>. Accessed 11 Nov 2017
- 786 Kohyama T (1986) Tree size structure of stands and each species in primary warm-  
787 temperate rain forests of Southern Japan. *Bot Mag Tokyo* 99:267–279.  
788 <https://doi.org/10.1007/BF02489543>

- 789 Lacerda AEB, Nimmo ER, Sebbenn AM (2013) Demography of *Hymenaea courbaril*.  
790 For Sci 59:15–26
- 791 Lamprecht H (1990) Silvicultura nos tropicos: ecossistemas florestais e respectivas  
792 especies arboreas - possibilidades e metodos de aproveitamento sustentado. GTZ,  
793 Eschborn
- 794 Lansanova LR (2012) Ajuste De Funções De Afilamento Para Espécies Florestais  
795 Comerciais Do Bioma Amazônico Matogrossense. Universidade Federal do Mato  
796 Grosso - UFMT
- 797 Lansanova LR, Silva FA da, Schons CT, Pereira ACDS (2018) Comparação Entre  
798 Diferentes Métodos Para Estimativa Volumétrica De Espécies Comerciais Da  
799 Amazônia. BIOFIX Sci J 3:109–115. <https://doi.org/10.5380/biofix.v3i1.57489>
- 800 Lappi J (1997) A Longitudinal Analysis of Height / Diameter Curves HEIGHT PLAYS  
801 TWO in modeling the growth. For Sci 43:555–570
- 802 Larson BC (1992) Pathways of development in mixed-species stands. In: Kelty MJ,  
803 Larson BC, Oliver CD (eds) The Ecology and Silviculture of Mixed-Species  
804 Forests. Springer, Dordrecht, Netherlands, pp 3–10
- 805 Lundqvist L (2017) Tamm Review: Selection system reduces long-term volume growth  
806 in Fennoscandic uneven-aged Norway spruce forests. For Ecol Manage 391:362–  
807 375. <https://doi.org/10.1016/j.foreco.2017.02.011>
- 808 Maciel MDNM, Watzlawick LF, Schoeninger ER, Yamaji FM (2017) Classificação  
809 ecológica das espécies arbóreas. Rev Acadêmica Ciência Anim 1:69.  
810 <https://doi.org/10.7213/cienciaanimal.v1i2.14922>

- 811 Madron LD De, Forni E (1997) Aménagement forestier dans l'Est du Cameroun. Bois  
812 Forêts des Trop 39–50. <https://doi.org/https://agritrop.cirad.fr/389267>
- 813 Mattioli W, Ferrari B, Giuliarelli D, et al (2015) Conversion of Mountain Beech  
814 Coppices into High Forest: An Example for Ecological Intensification. Environ  
815 Manage 56:1159–1169. <https://doi.org/10.1007/s00267-015-0549-2>
- 816 Mattos PP, Braz EM, Domene VD, et al (2015) Relação Clima-Crescimento De Árvores  
817 De Mimosa Tenuiflora Em Floresta Tropical Seca Sazonal, Brazil. Cerne 21:141–  
818 149. <https://doi.org/10.1590/01047760201521011460>
- 819 Meyer HA (1952) Balanced Growth , and Drain in Uneven-Aged Forests. J For
- 820 Miller DP (2004) Bootstrap 101: Obtain robust confidence intervals for any statistic. In:  
821 Twenty-Ninth Annual SAS Users Group International Conference. SAS Institute,  
822 Cary, NC, pp 193–29
- 823 Miranda DLC, Higuchi N, Trumbore SE, et al (2018) Using radiocarbon-calibrated  
824 dendrochronology to improve tree-cutting cycle estimates for timber management  
825 in southern Amazon forests. Trees - Struct Funct 32:587–602.  
826 <https://doi.org/10.1007/s00468-018-1658-3>
- 827 Miranda ZP, Guedes MC, Rosa SA, Schöngart J (2018) Volume increment modeling  
828 and subsidies for the management of the tree *Mora paraensis* (Ducke) Ducke based  
829 on the study of growth rings. Trees - Struct Funct 32:277–286.  
830 <https://doi.org/10.1007/s00468-017-1630-7>

- 831 Muller-Landau HC, Condit RS, Harms KE, et al (2006) Comparing tropical forest tree  
832 size distributions with the predictions of metabolic ecology and equilibrium  
833 models. *Ecol Lett* 9:589–602. <https://doi.org/10.1111/j.1461-0248.2006.00915.x>
- 834 Nyland RD (2002) *Silviculture: concepts and applications*, 2nd edn. McGraw-Hill,  
835 Boston, MA
- 836 Odum EP (1988) *Ecologia*, Guanabara. Rio de Janeiro, Brasil
- 837 O’Hara K (2014) Transformations to Multiaged Stand Structures. In: *Multiaged*  
838 *Silviculture: Managing for Complex Forest Stand Structures*. Oxford University  
839 Press, pp 1–15
- 840 Oliveira MF, Braz EM, Mattos PP de, et al (2015) Padrão de crescimento e diâmetro  
841 ótimo de corte de cupiúba na microrregião de Sinop, MT. Colombo, Brasil
- 842 Oliver CD (1992) Similarities of stand structures and stand development processes  
843 throughout the world—some evidence and applications to silviculture through  
844 adaptive management. In: Kelty MJ, Larson BC, Oliver CD (eds) *The Ecology and*  
845 *Silviculture of Mixed-Species Forests*. Springer, Dordrecht, Dordrecht,  
846 Netherlands, pp 11–26
- 847 Ong RC, Kleine M (1996) DIPSIM: Dipterocarp forest growth simulation model--a tool  
848 for forest-level management planning. *Dipterocarp For Ecosyst Sustain Manag*  
849 *World Sci Singapore* 228–246. <https://doi.org/10.1017/CBO9781107415324.004>
- 850 Orellana E, Figueiredo Filho A (2017) Uso do método da predição de parâmetros para  
851 projetar a distribuição diamétrica em florestas nativas com a função Weibull.  
852 *Ciência Florest* 27:981–991

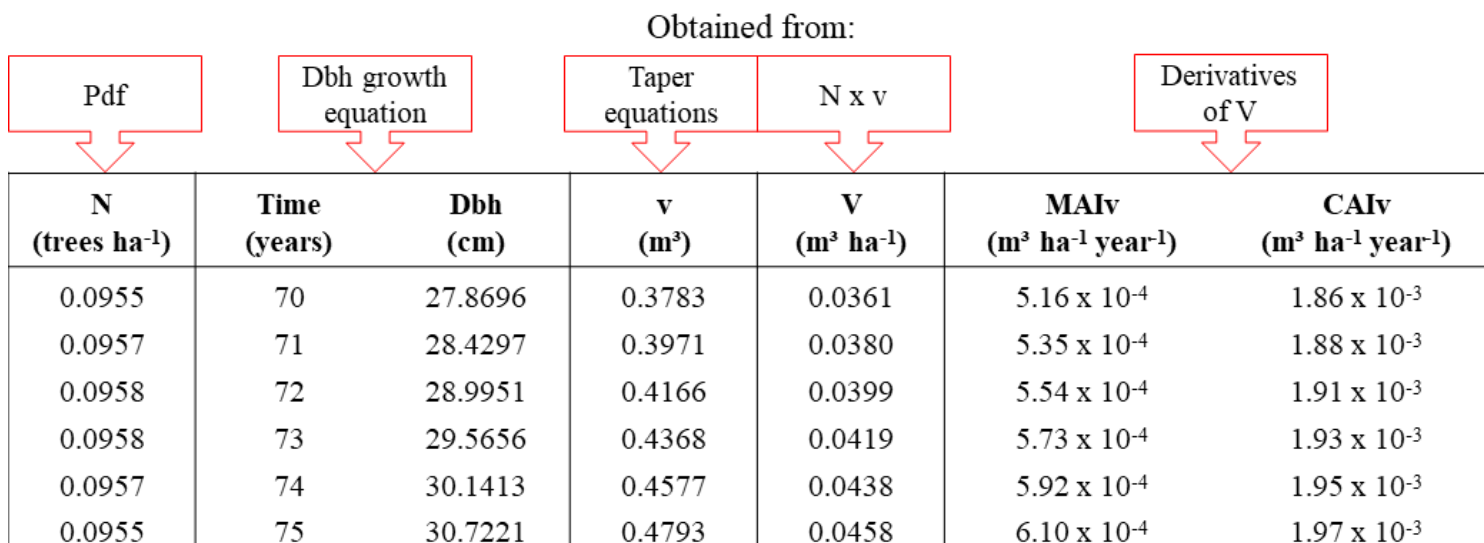
- 853 Pascal J (2003) Notions sur les structure et dynamique des forêts tropicales humides.  
854 Rev For Française 118. <https://doi.org/10.4267/2042/5765>
- 855 Reis LP, Ruschel AR, Coelho AA, et al (2010) Avaliação do potencial madeireiro na  
856 Floresta Nacional do Tapajós após 28 anos da exploração florestal. *Pesqui Florest*  
857 *Bras* 30:265–281. <https://doi.org/10.4336/2010.pfb.30.64.265>
- 858 Reynolds MR, Burk TE, Huang W-C (1988) Goodness-of-Fit Tests and Model  
859 Selection Procedures for Diameter Distribution Models. *For Sci* 34:373–399
- 860 Ribeiro ES, Paula MH De, Rossi R, et al (2016) Espécies florestais comercializadas  
861 pelo estado de mato grosso. *Biodiversidade* 15:2–20
- 862 Rinn F (1996) TSAP Win v. 3.6: Reference Manual: Computer Program for Tree-Ring  
863 Analysis and Presentation. RINNTECH, Heidelberg, Germany
- 864 Rosa SA, Barbosa ACMC, Junk WJ, et al (2017) Growth models based on tree-ring data  
865 for the Neotropical tree species *Calophyllum brasiliense* across different Brazilian  
866 wetlands: implications for conservation and management. *Trees - Struct Funct*  
867 31:729–742. <https://doi.org/10.1007/s00468-016-1503-5>
- 868 Rubin BD, Manion PD, Faber-Langendoen D (2006) Diameter distributions and  
869 structural sustainability in forests. *For Ecol Manage* 222:427–438.  
870 <https://doi.org/10.1016/j.foreco.2005.10.049>
- 871 Schaaf LB, Figueiredo Filho A, Galvão F, Sanquetta CR (2006) Alteração na estrutura  
872 diamétrica de uma floresta ombrófila mista no período entre 1979 e 2000. *Rev*  
873 *Árvore* 30:283–295. <https://doi.org/10.1590/S0100-67622006000200016>

- 874 Schöngart J (2011) Amazonian Floodplain Forests. Springer Netherlands, Dordrecht
- 875 Schöngart J (2008) Growth-Oriented Logging (GOL): A new concept towards  
876 sustainable forest management in Central Amazonian várzea floodplains. For Ecol  
877 Manage 256:46–58. <https://doi.org/10.1016/j.foreco.2008.03.037>
- 878 Scolforo J (1998) Manejo florestal. Lavras: UFLA/FAEPE 1158–1170
- 879 Sebbenn AM, Degen B, Azevedo VCR, et al (2008) Modelling the long-term impacts of  
880 selective logging on genetic diversity and demographic structure of four tropical  
881 tree species in the Amazon forest. For Ecol Manage 254:335–349.  
882 <https://doi.org/10.1016/j.foreco.2007.08.009>
- 883 Seydack AHW (2012) Regulation of Timber Yield Sustainability for Tropical and  
884 Subtropical Moist Forests: Ecosilvicultural Paradigms and Economic Constraints.  
885 In: Continuous Cover Forestry. Managing Forest Ecosystems. Springer, Dordrecht,  
886 pp 129–165
- 887 Seydack AHW (2000) Theory and Practice of Yield Regulation Systems for Sustainable  
888 Management of Tropical and Subtropical Moist Natural Forests. In: von Gadow K,  
889 Pukkala T, Tomé M (eds) Sustainable Forest Management. Springer Netherlands,  
890 Dordrecht, pp 257–318
- 891 Seydack AHW, Vermeulen WJ, Heyns HE, et al (1995) An unconventional approach to  
892 timber yield regulation for multi-aged, multispecies forests. II. Application to a  
893 South African forest. For Ecol Manage 77:155–168. [https://doi.org/10.1016/0378-](https://doi.org/10.1016/0378-1127(95)03578-X)  
894 [1127\(95\)03578-X](https://doi.org/10.1016/0378-1127(95)03578-X)



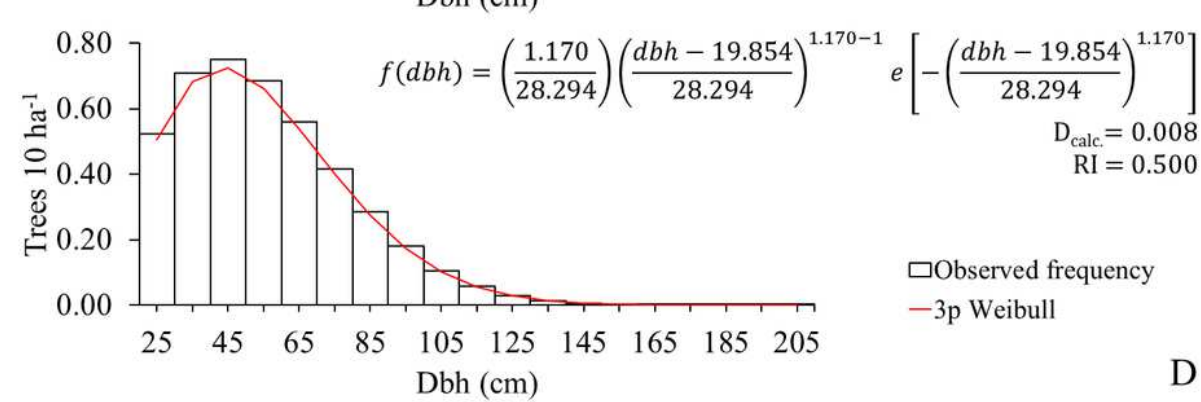
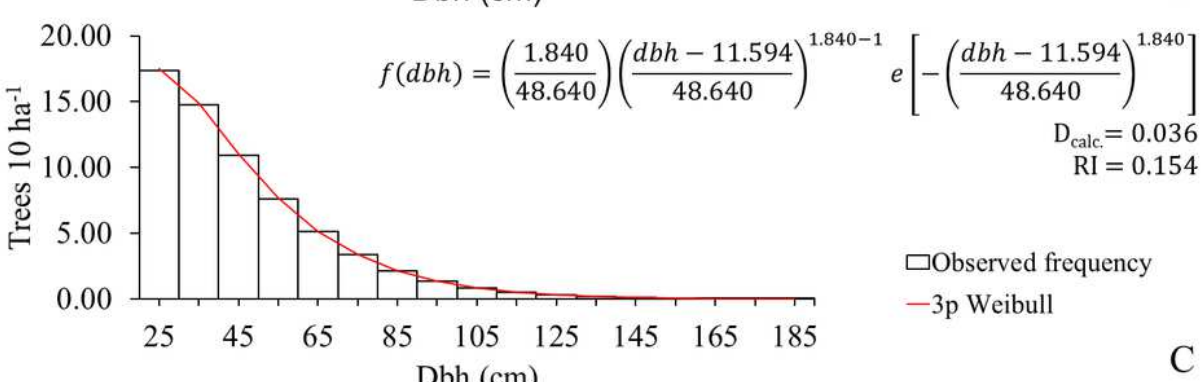
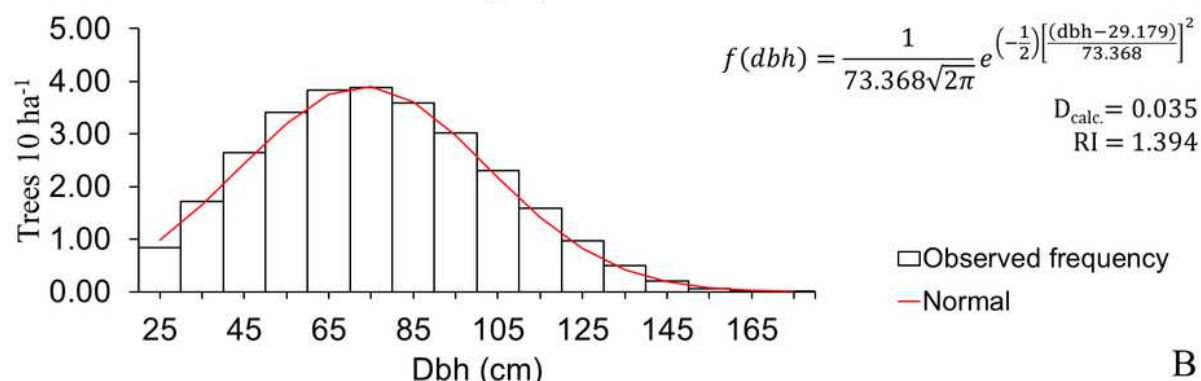
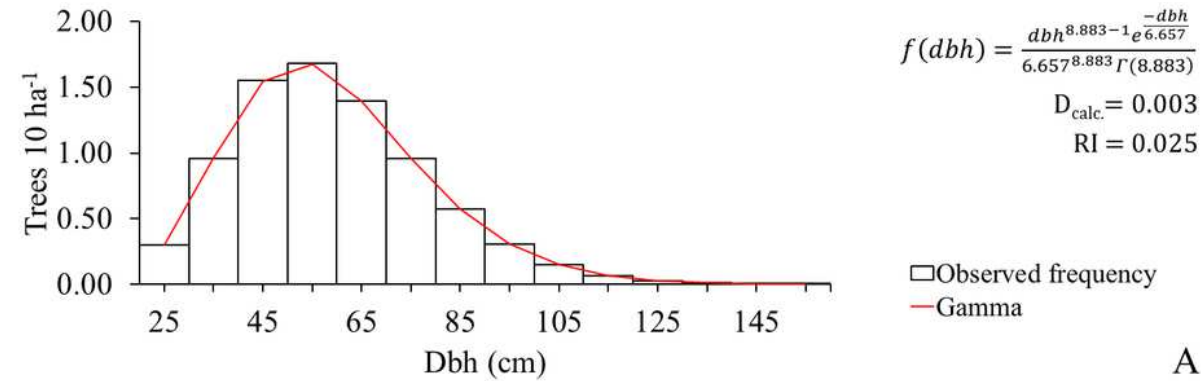
- 895 Silva JNM (1989) The behavior of the tropical rain forest of the Brazilian amazon after  
896 logging. Oxford University
- 897 Silva JNM (1997) Manejo de florestas de terra-firme da Amazônia Brasileira. In:  
898 Tópicos em manejo florestal sustentável. Embrapa Florestas, Colombo, pp 59–96
- 899 Silva JNM, Carvalho JOP de, Lopes J do CA (1985) Inventário florestal de uma área  
900 experimental na Floresta Nacional do Tapajós. Bol Pesqui Florest 38–110
- 901 SIVAM P (2002) Mapa de Vegetação da Amazônia Legal – RADAM/SIPAM. In:  
902 DNPM, Ministério Minas e Energ.  
903 [http://www.dpi.inpe.br/Ambdata/mapa\\_sipam.php](http://www.dpi.inpe.br/Ambdata/mapa_sipam.php). Accessed 10 Mar 2018
- 904 van Gardingen PR, Valle D, Thompson I (2006) Evaluation of yield regulation options  
905 for primary forest in Tapajós National Forest, Brazil. For Ecol Manage 231:184–  
906 195. <https://doi.org/10.1016/j.foreco.2006.05.047>
- 907 Vanclay JK (1989) A growth model for north Queensland rainforests. For Ecol Manage  
908 27:245–271. [https://doi.org/10.1016/0378-1127\(89\)90110-2](https://doi.org/10.1016/0378-1127(89)90110-2)
- 909 Wang X, Hao Z, Zhang J, et al (2009) Tree size distributions in an old-growth  
910 temperate forest. Oikos 118:25–36. [https://doi.org/10.1111/j.1600-  
911 0706.2008.16598.x](https://doi.org/10.1111/j.1600-0706.2008.16598.x)
- 912 Wright SJ, Muller-Landau HC, Condit R, Hubbell SP (2003) Shade tolerance, realized  
913 vital rates, and size distributions of tropical trees. Ecology 84:3174–3185
- 914 Yegang W, Jinxuan H (1988) A Spectral Analysis of the Population Dynamics of  
915 Korean Pine in the Mixed Broad-leaved Pinus koraiensis Forest. J Ecol 19–23

# Figures



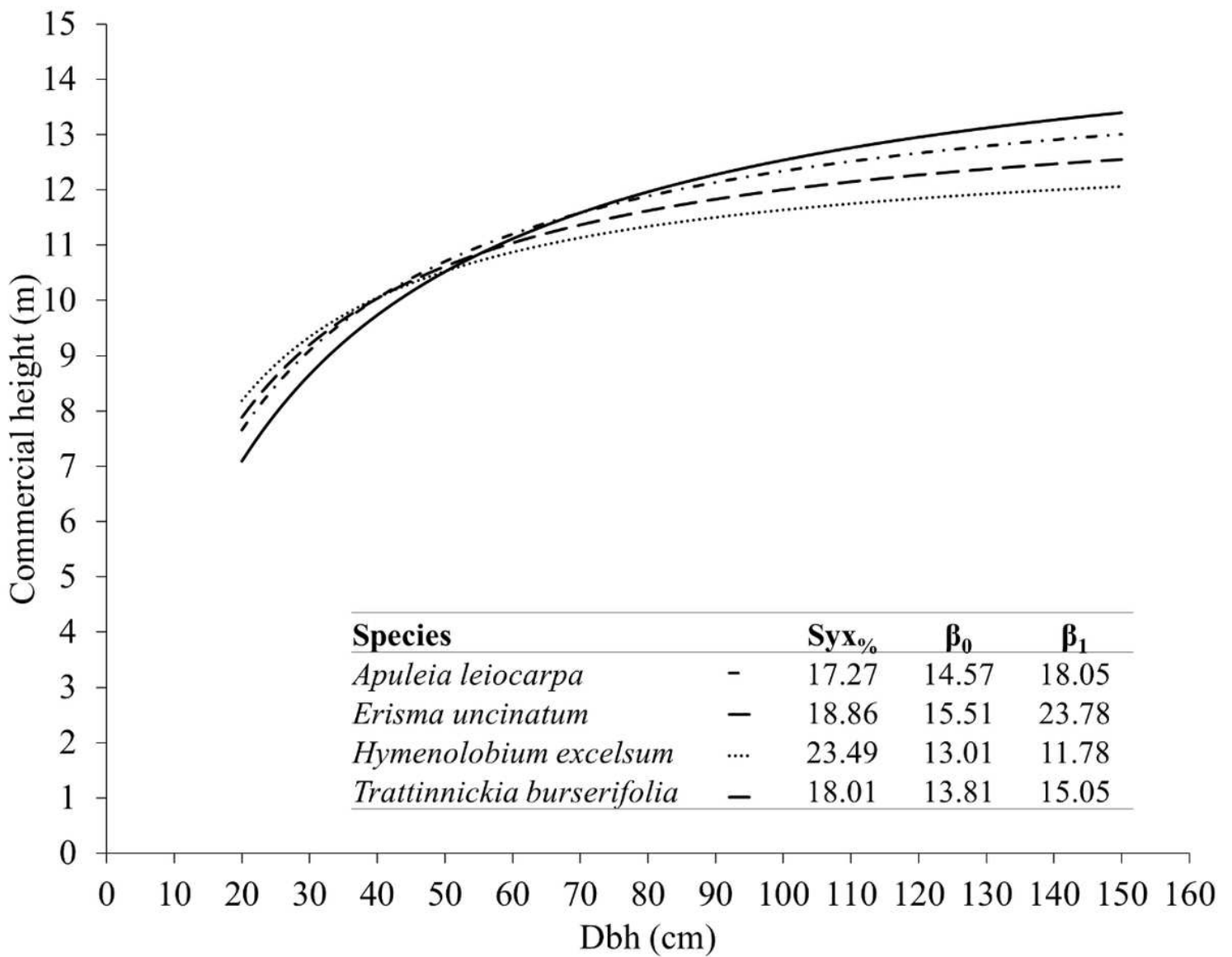
**Figure 1**

Scheme showing the methodological process for obtaining the curves of volume increment of the population. The boxes with red outline inform the data source of the column to which the arrows are directed. PDFs = probability density functions; N = number of trees per hectare; dbh = diameter at 1.30 m above ground level; v = volume of individual tree; V = species population volume (m<sup>3</sup> ha<sup>-1</sup>); MAIv = mean annual volume increment; and CAIv = current annual volume increment.



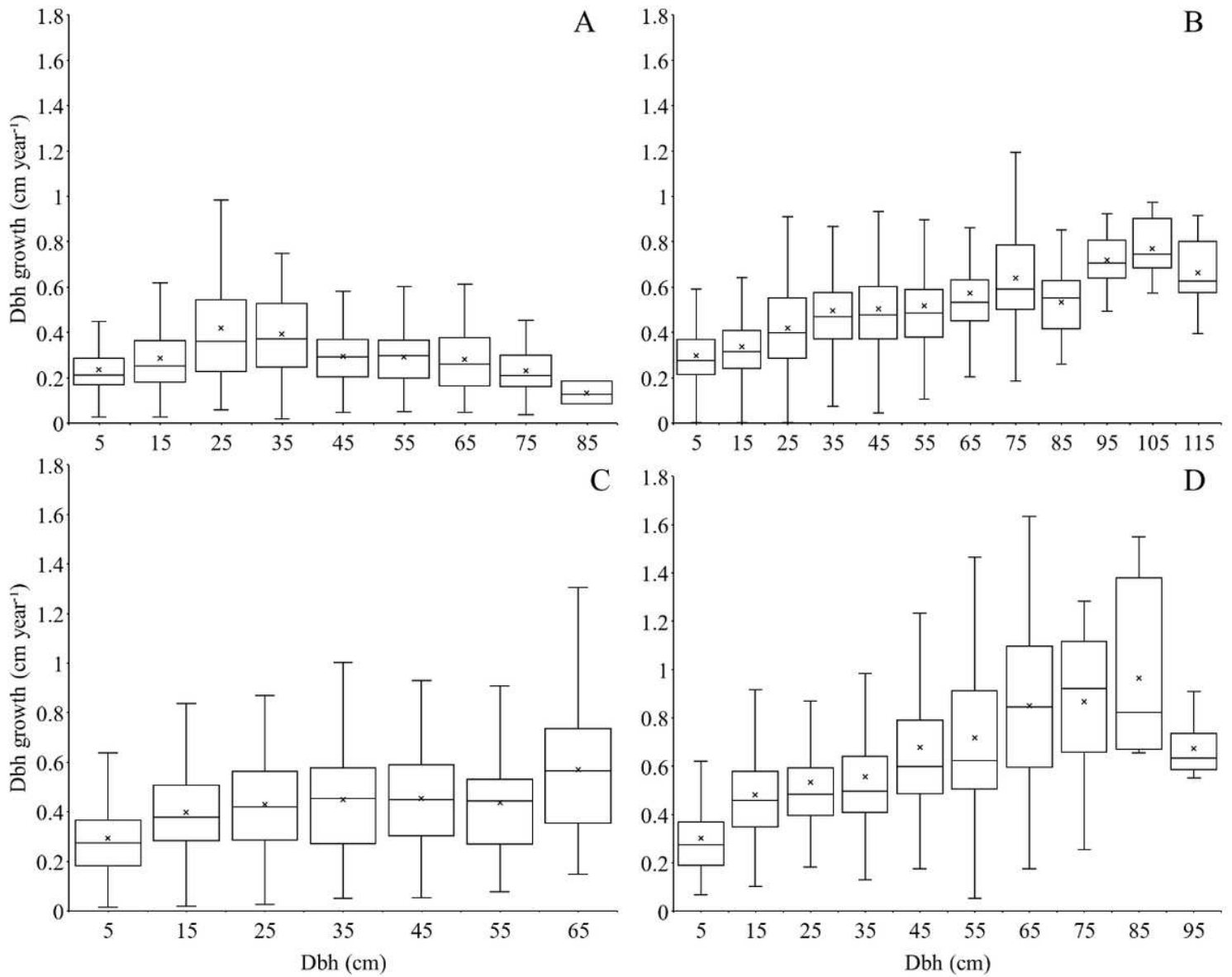
**Figure 2**

Probability density functions with better fitting for *Apuleia leiocarpa* (A), *Erisma uncinatum* (B), *Hymenolobium excelsum* (C) and *Trattinnickia burserifolia* (D). Dbh = diameter at 1.30 m above ground level (cm);  $\Gamma$  = gamma function;  $D_{calc.}$  = maximum absolute value between fitted pdf and observed values in each compartment; IR = error index (Reynolds et al., 1988).



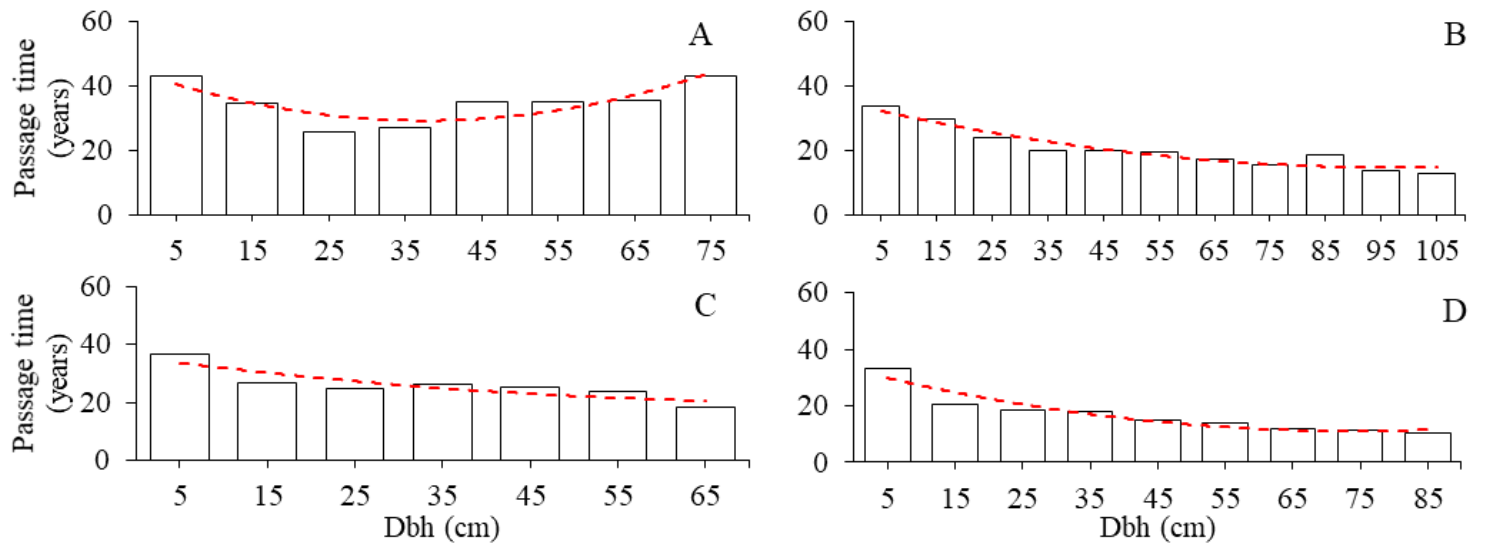
**Figure 3**

Heigh/dbh models of *Apuleia leiocarpa*, *Erisma uncinatum*, *Hymenolobium excelsum*, and *Trattinnickia burserifolia*. dbh = diameter at 1.30 m above ground level (cm); β<sub>0</sub> and β<sub>1</sub> = equation parameters fitted by non-linear regression; Sy<sub>x</sub>% = relative residual standard error (%)



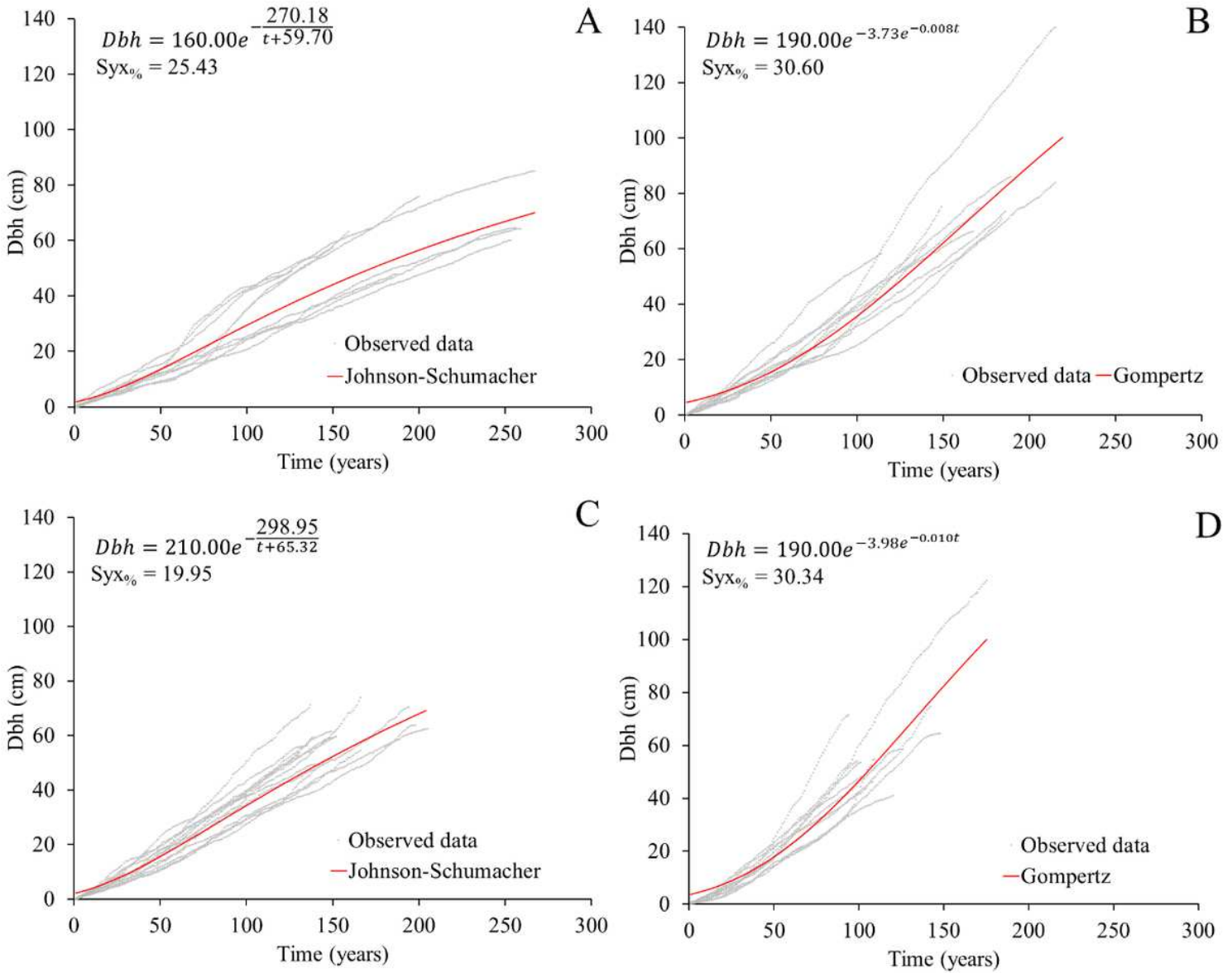
**Figure 4**

Boxplot of the mean annual periodic increment by diameter class of *Apuleia leiocarpa* (A), *Erisma uncinatum* (B), *Hymenolobium excelsum* (C), and *Trattinnickia burserifolia* (D). Markers (x) represent the mean increment per diameter class, which were calculated only for the diameter classes with three or more trees. Dbh = diameter at 1.30 m above ground level.



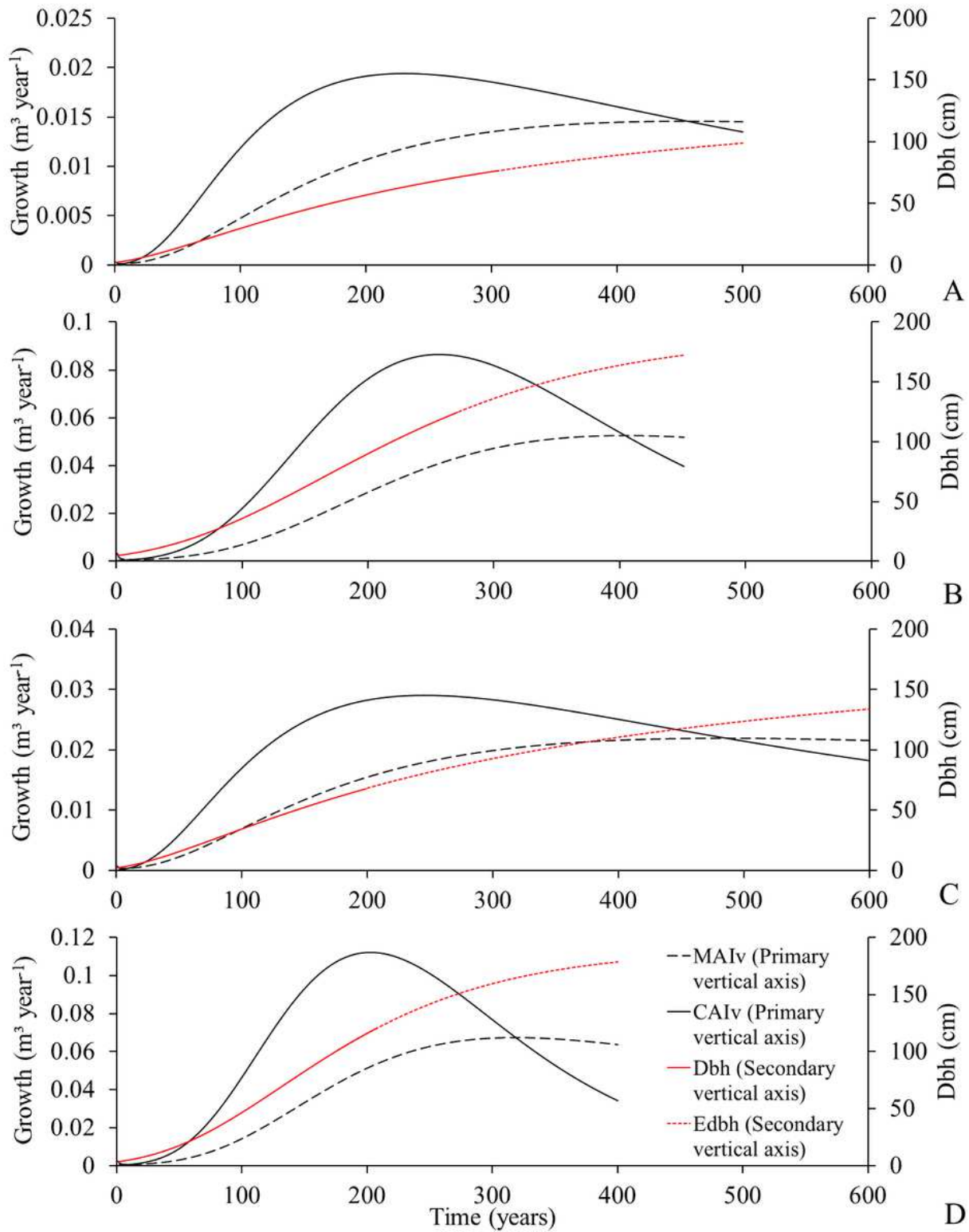
**Figure 5**

Mean passage time by diameter class for *Apuleia leiocarpa* (A), *Erisma uncinatum* (B), *Hymenolobium excelsum* (C), and *Trattinnickia burserifolia* (D) and their respective tendency lines. Only passage times for diameter classes with three or more trees were considered. Dbh = diameter at 1.30 m above ground level.



**Figure 6**

Accumulated diametric growth equations fitted for *Apuleia leiocarpa* (A), *Erisma uncinatum* (B), *Hymenolobium excelsum* (C), and *Trattinnickia burserifolia* (D). The growth equations were fitted for the ages represented by three or more samples. Dbh = diameter at 1.30 m above ground level; t = time (years); Syx% = Relative Residual Standard Error (%).

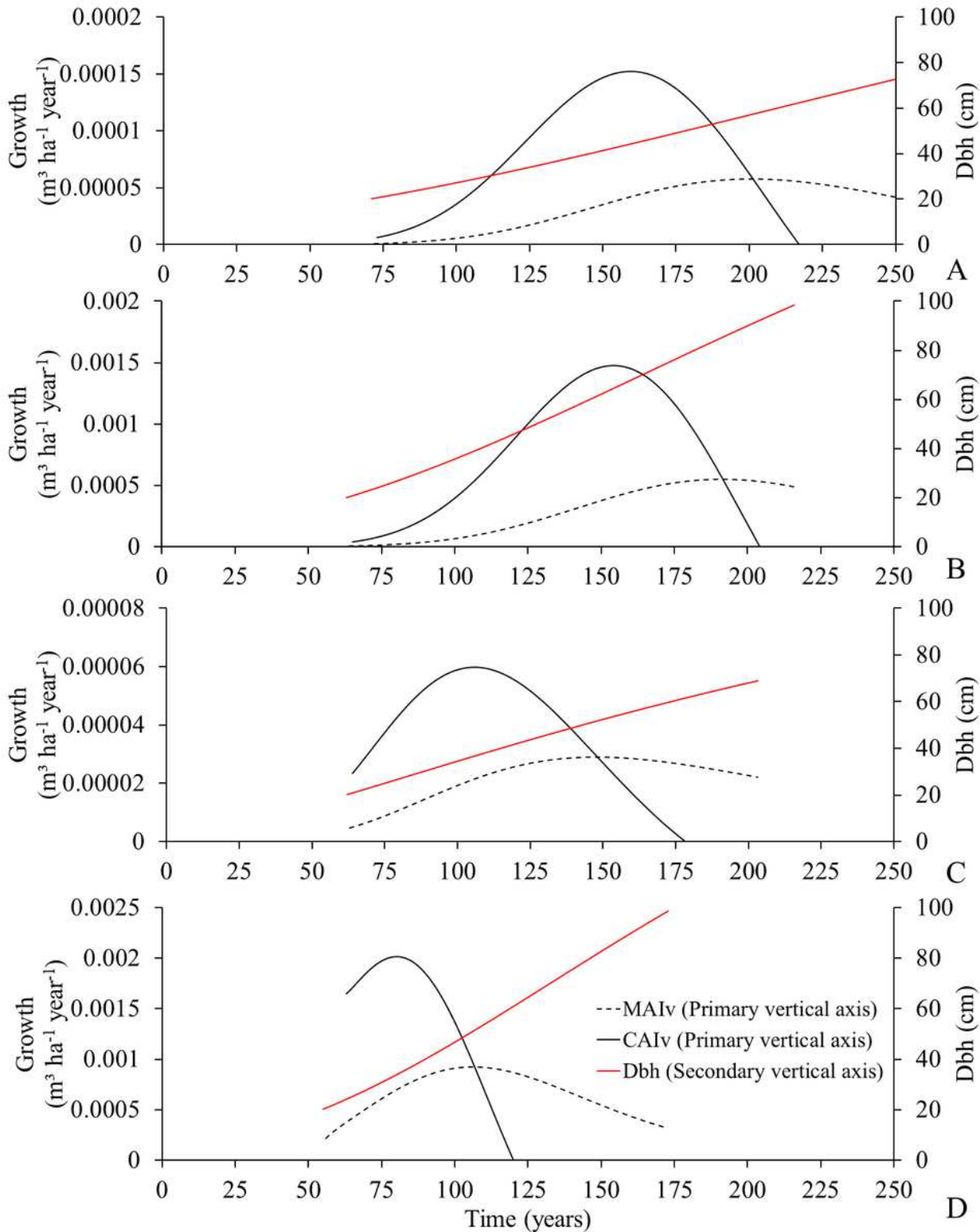


**Figure 7**

Volumetric increment curves (black) and dbh growth equation (red) for *Apuleia leiocarpa* (A), *Erisma uncinatum* (B), *Hymenolobium excelsum* (C), and *Trattinnickia burserifolia* (D). In the primary y-axis: MAIv = mean annual volumetric increment; CAIv = current annual volumetric increment. On the secondary y-axis: dbh = diameter at 1.30 m above ground level (cm) (growth equation fitted within the measured data



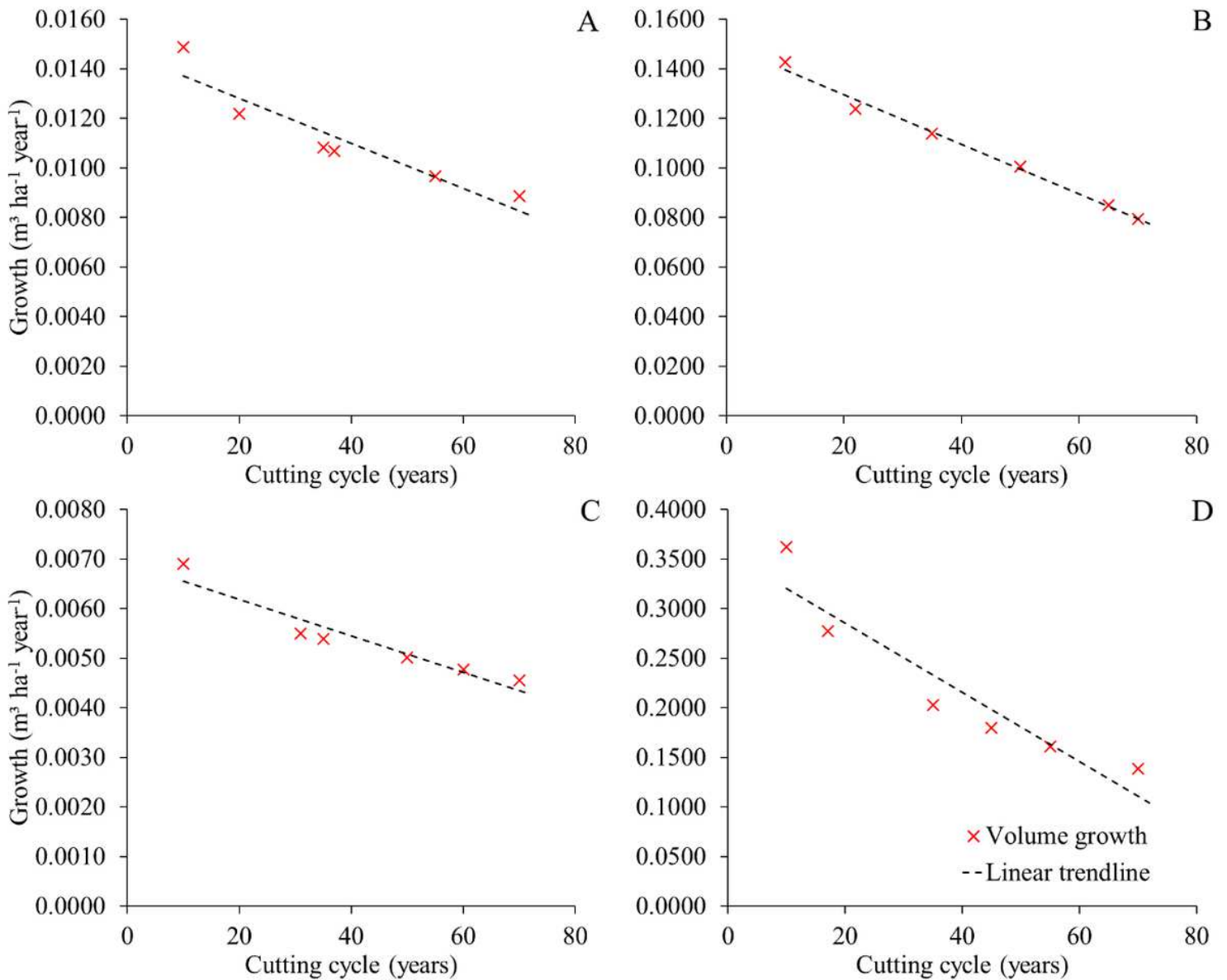
range); Edbh = accumulated diameter at 1.30 m above ground level (growth equation outside the measured data range).



**Figure 8**

Volumetric and diametric increment curves for the population of *Apuleia leiocarpa* (A), *Erisma uncinatum* (B), *Hymenolobium excelsum* (C) and *Trattinnickia burserifolia* (D) (dbh  $\geq$  20 cm). In the primary y-axis:

MAIv = mean annual increment in volume; CAIv = current annual increment in volume. In the secondary y-axis: dbh = diameter at 1.30 m above ground level (cm), obtained from the growth equation.



**Figure 9**

Annual volume increment by the MCD defined by the population's biological rotation age and different cutting cycles for *Apuleia leiocarpa* (A), *Erismia uncinatum* (B), *Hymenolobium excelsum* (C), and *Trattinnickia burserifolia* (D).